

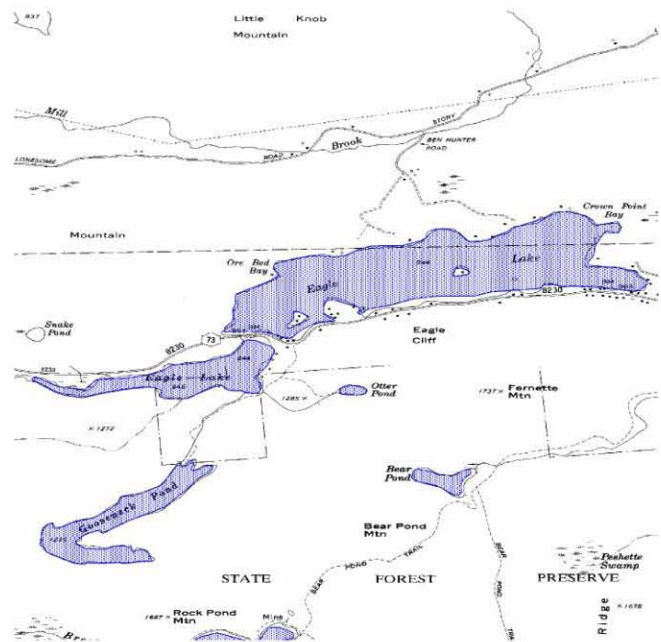
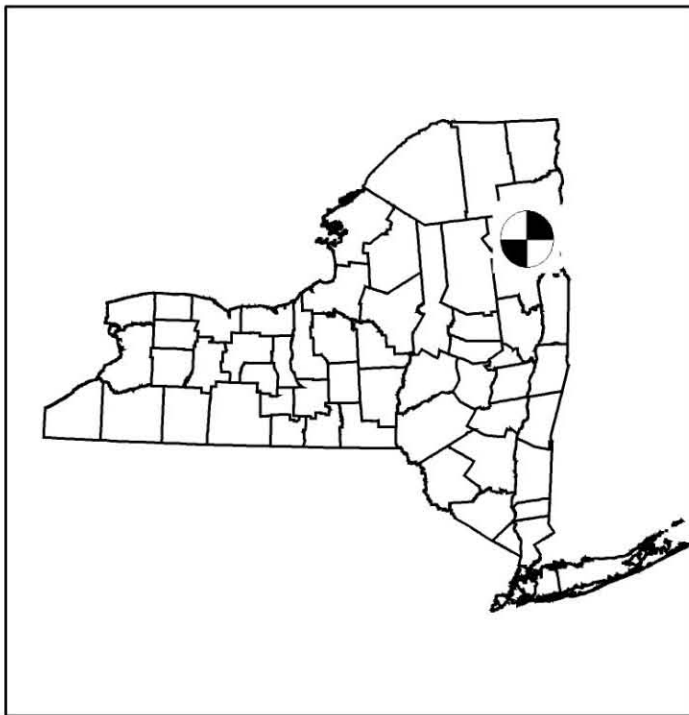


Division of Water

New York
Citizens Statewide Lake Assessment Program
(CSLAP)

2007 Annual Report- Eagle Lake

January, 2008



2007 INTERPRETIVE SUMMARY

NEW YORK CITIZENS STATEWIDE LAKE ASSESSMENT PROGRAM (CSLAP)

EAGLE LAKE

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NY Federation of Lake Associations

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BACKGROUND AND ACKNOWLEDGMENT

The Citizens Statewide Lake Assessment Program (CSLAP) is a volunteer lake monitoring program conducted by the NYS Department of Environmental Conservation (NYSDEC) and the NYS Federation of Lake Associations (FOLA). Founded in 1986 with 25 pilot lakes, the program has involved more than 200 lakes, ponds, and reservoirs and 1000 volunteers from eastern Long Island to the northern Adirondacks to the western-most lake in New York, and from 10-acre ponds to several Finger Lakes, Lake Ontario, Lake George, and lakes within state parks. In this program, lay volunteers trained by the NYSDEC and FOLA collect water samples, observations, and perception data every other week in a 15 week interval between May and October. Water samples are analyzed by certified laboratories. Analytical results are interpreted by the NYSDEC and FOLA and utilized for a variety of purposes by the State of New York, local governments, researchers, and, most importantly, participating lake associations. This report summarizes the 2007 sampling results for **Eagle Lake**.

Eagle Lake is a 422 acre, class B lake found in the Town of Ticonderoga in Essex County, within the Eastern Adirondack region of New York State. Eagle Lake was first sampled as part of CSLAP in 2000. The following volunteers have participated in CSLAP, and deserve most of the credit for the success of this program at **Eagle Lake**: **Paul and Mary Lloyd Burroughs**.

In addition, the authors wish to acknowledge the following individuals, without whom this project and report would never have been completed:

From the Department of Environmental Conservation, Dick Draper, and Margaret Novak for on-going support of the program; Jay Bloomfield and James Sutherland, for their work in developing and implementing the program, and the technical staff from the Lake Services Section and the Statewide Water Monitoring Section, for continued technical review of program design.

From the Federation of Lake Associations, Anne Saltman, Dr. John Colgan, Don Keppel, Nancy Mueller and the Board of Directors, for their continued strong support of CSLAP.

The New York State Department of Health (prior to 2002) and Upstate Freshwater Institute (since 2002), particularly Steve Effler, MaryGail Perkins, and Elizabeth Miller, provided laboratory materials and all analytical services, reviewed the raw data, and implemented the quality assurance/quality control program.

Finally, but most importantly, the authors would like to thank the more than 1,500 volunteers who have made CSLAP a model for lay monitoring programs throughout the country and the recipient of a national environmental achievement award. Their time and effort have served to greatly expand the efforts of the state and the public to protect and enhance the magnificent water resources of New York State.

WHAT'S NEW IN THE 2007 CSLAP REPORT?

In a never ending quest to make the CSLAP reports more useful and comprehensive, or at least more interesting and worthy of a cover-to-cover read, the NYSDEC makes small changes in the CSLAP report each year. Some of these changes are small and include fixing previous errors, based on corrections provided by readers or re-editing. Others are more substantial and reflect improvements in technology (better graphics or layout capabilities) or information about the lake or its watershed. For example, the 2005 CSLAP report included information about regulated activities in the area around the lake and a compendium of other state water quality data for the lake. The 2006 report included fish stocking, fisheries regulations, and fish consumption advisory information for the first time, as well as site location maps, information about rare, threatened, or endangered plant species in lake, and detailed discussions about lake use impacts and their implications for the state Priority Waterbody List.

The 2007 CSLAP report has been improved by the following new information:

- RIBS water quality data for streams entering the lake. The state Rotating Intensive Basins monitoring program began collecting stream water quality data in 1987 and information about benthic (bottom) organisms in the rivers and streams throughout New York state since 1972. Some of these streams and rivers are tributary to CSLAP lakes. Where available, these data are summarized in this report.
- NOAA weather data summaries have been expanded in two ways. Annual data summaries have shifted from a single assessment (“wet”, “normal”, etc.) to two assessments—precipitation in the winter and spring, and summer rainfall. Statewide summaries are still used for evaluating statewide water quality trends, but “divisional” data, corresponding to 10 regions in the state, are used for evaluating lake specific trends. General information about regional and statewide precipitation patterns is also provided.
- Historical aquatic plant identifications, such as those provided by the 1980s Adirondack Lake Survey or the 1930s Biological Surveys, were provided by scientific (Latin) name. The common names for these plants have now been included.
- Ammonia, total nitrogen, and calcium data have been included in the statewide summaries, based on five years of CSLAP water quality data for these indicators.
- The exotic plant distribution map section has been expanded to include Brazilian elodea (*Egeria densa*), variable watermilfoil (*Myriophyllum heterophyllum*), parrotfeather (*Myriophyllum aquaticum*), and European frogbit (*Hydrocharis morsus-ranae*), four submergent exotic plants that are threatening to significantly increase in distribution or are growing very invasively in an increasing number of New York state lakes.
- Finally, the report now includes a “So What Have We Learned Through CSLAP” section that summarizes the long-term trends evaluation and other important findings from the CSLAP dataset.

We hope this report satisfies the needs of lake associations and CSLAP participants, and we continue to welcome suggestions for improving the program, reporting, and other avenues for gaining greater knowledge about the lakes of New York State.

FINDINGS AND EXECUTIVE SUMMARY

Eagle Lake was sampled as part of the New York Citizens Statewide Lake Assessment Program in 2007. For all program waters, water-quality conditions and public perception of the lake each year and historically have been evaluated within annual reports issued after each sampling season. This report attempts to summarize both the 2007 CSLAP data and an historical comparison of the data collected within the 2007 sampling season and data collected at Eagle Lake prior to 2007.

The majority of the short- and long-term analyses of the water quality conditions in Eagle Lake are summarized in Table 2, divided into assessments of eutrophication indicators, other water quality indicators, and lake perception indicators. The CSLAP data indicate that the lake can be classified as oligotrophic, or highly unproductive. Eagle Lake was about as productive as in the typical CSLAP sampling season, based on phosphorus and chlorophyll *a* readings close to the long-term average for the lake. The drop in water clarity in the last several years has not been mirrored by an increase in algae levels, and it is likely that the small changes were probably within the normal variability for this lake, or at least is due to higher color and wetter weather. CSLAP data suggest that water clarity is probably closely influenced by both algae and nutrients (based on the similar trophic classification generated from each of the trophic indicators), and the nitrogen to phosphorus ratios indicate that algae levels in Eagle Lake are controlled by phosphorus. Lake productivity decreases slightly but not significantly over the course of a typical sampling season, due in large part to deepwater nutrient levels that are similar to those at the lake surface (and thus do not “enrich” the surface waters in late summer and fall, despite slight deepwater oxygen deficits dating back to the 1930s). Phosphorus levels in the lake fall well below the state phosphorus guidance value, resulting in water transparency readings that significantly exceed the minimum recommended water clarity for swimming beaches.

The lake is weakly colored (low levels of dissolved organic matter); while the recent (slight) rise in these color readings is coincident with the slight drop in water transparency, these indicators have not been closely correlated. As with most CSLAP lakes, it is likely that the recent rise in color is in response to wetter weather. Eagle Lake has moderately softwater, alkaline (above neutral) pH readings, and mostly undetectable nitrate and ammonia readings. Conductivity readings have varied annually and sample to sample. pH readings consistently fall between the NYS water quality standards (=6.5 to 8.5), and should be adequate to support most organisms found in the lake, although pH has been slightly higher than usual in the last few years. Neither nitrate nor ammonia levels in surface or bottom waters are high enough to represent a threat to the lake. Calcium levels are probably high enough to support zebra mussels, although these exotic animals have not been found in the lake.

The recreational suitability of Eagle Lake has been stable since first evaluated in 2000. Recreational conditions are most often described as “slightly impaired” for most uses, and “not quite crystal clear”. The latter is slightly less favorable than expected given the water quality conditions in the lake. These recreational assessments are atypical of (less favorable than) other lakes with similar water quality characteristics, but indicative of lakes with similar plant densities (aquatic plants regularly grow to, but not densely at, the lake surface, according to the sampling volunteers, and “excessive weed growth” is regularly reported as impacting lake use). The recreational assessments are seasonally stable, consistent with seasonally stable aquatic plant and water quality conditions.

The 1996 NYSDEC Priority Waterbody Listings (PWL) for the Upper Hudson River drainage basin do not include Eagle Lake. The CSLAP datasets suggest that recreation (bathing and non-contact recreation such as boating) may be *stressed* by excessive weed growth. The next PWL review for the Upper Hudson River drainage basin will likely occur in 2008.

General Comments and Questions:

- ***What is the condition of Eagle Lake?***

Water quality conditions in Eagle Lake are more than adequate to support most recreational uses of the lake during the summer, and can be best described as oligotrophic, or highly unproductive, with nutrient levels low enough to indicate that water quality changes are unlikely in at least the near future. These recreational assessments, however, indicate “slightly impaired” conditions, but continue to most strongly influenced by excessive weed growth (or poor weather).

- ***What about the dark and murky bottom waters of the lake?***

Although Eagle Lake is thermally stratified, deepwater nutrient (phosphorus and nitrogen) levels are nearly identical to those measured at the lake surface. This suggests that nutrient release from bottom sediments are probably negligible, and that deepwater oxygen levels probably remain high during the summer (although there probably remains some oxygen deficits at the lake bottom, based on data collected in the 1932 Biological Survey and 1999 LCI survey of the lake).

- ***How does this condition change from spring showers thru changing of the leaves?***

The productivity of Eagle Lake (clarity, nutrient and algae levels) decreases slightly during the summer, though in a manner that is not statistically significant, and recreational assessments are seasonally stable, coincident with seasonal stability in weed coverage and densities.

- ***How has the condition changed since CSLAP sampling began on the lake and/or relative to historical values?***

Color readings have increased slightly since 2000, a pattern common to many CSLAP lakes, perhaps in response to wetter weather or the 2002 shift in laboratories, although color readings remain fairly low. Water transparency readings have also been lower in recent years, perhaps due to the rise in color and the wetter weather, although these readings remain high. None of the other water quality indicators have exhibited significant or long-term changes.

- ***How does Eagle Lake compare to other similar lakes (nearby lakes,....)?***

Eagle Lake is less productive than other nearby (Upper Hudson River basin) lakes, other lakes classified for swimming and other contact recreation (Class B), and other NYS lakes. However, recreational assessments have consistently been less favorable than in these other lakes, due to the influence of excessive weed growth (and/or poor weather conditions).

- ***Based on these data, what should be done to improve or maintain Eagle Lake?***

Given the low lake productivity, and little evidence of water quality threats, management of water quality conditions in Eagle Lake should focus on reducing nutrient loading to the lake, through maintaining septic systems, shoreline buffer zones, limited use of lawn fertilizers, minimizing land disturbances in the near-lake watershed, and localized stormwater management. The lake association is also advised to minimize introductions of exotic plants and animals from public and private launch areas into the lake, particularly given the strong connection between weeds and recreational assessments of the lake, as well as the increasing presence of zebra mussels in nearby lakes.

Context and Qualifiers

The NY Citizens Statewide Lake Assessment Program (CSLAP) is intended to be a long-term, standardized, trophic-based, water-quality monitoring program to facilitate comparison of water-quality data from season to season, year to year, and from lake to lake. The data and information collected through CSLAP can be utilized to identify water-quality problems, detect seasonal and long-term patterns, and educate sampling volunteers and lake residents about water-quality conditions and stressors at their lakes. It is particularly useful in evaluating the over-enrichment of aquatic plant (algae and rooted plant) communities in a lake, and the response of the lake to these trophic stressors.

Shorefront residents, lake managers, and government agencies are increasingly tasked to better assess and evaluate water-quality conditions and lake uses in NYS lakes, including those sampled through CSLAP, whether to address localized problems, meet water-quality standards, satisfy state and federal environmental reporting requirements, or enhance and balance a suite of lake uses. CSLAP data should be a part of this process, but only a part. For some lakes, particularly small lakes and ponds with limited public access by those who don't reside on the lake shore, CSLAP may be the sole source of data used to assess lake conditions. In addition, studies conducted through CSLAP find strong similarities between sampling sites in many, but not all, large lakes, and generally find a strong convergence of perceptions about lake and recreational use conditions within most lakes, based on a local familiarity with "normal" conditions and factors that might affect lake use. For the purpose of broad water-quality evaluations and understanding the connection between measured water-quality indicators and the support of broadly based recreational uses of the lake, CSLAP can be a singularly effective tool for standardizing the lake-assessment process. CSLAP volunteers, lake associations, and others engaged in lake assessment and management should continue to utilize CSLAP in this context.

However, for large, multi-use lakes, or those lakes that are threatened by pollutants not captured in eutrophication-based monitoring programs, CSLAP becomes a less effective primary tool for assessing lake condition and use impairments. For example, CSLAP data have only limited utility in evaluating the following:

- (a) contamination from bacteria or other biological toxins, particularly related to the safety of water use for potable intake or swimming
- (b) contamination from inorganic (e.g., metals) and organic (e.g., PCBs, DDT) compounds
- (c) portions of a lake not well mixed with the "open water" or otherwise distant from the primary sampling site(s), including the shoreline, bottom sediment and isolated coves
- (d) rooted aquatic plant impacts in areas of the lake not evaluated by the sampling volunteers
- (e) diverging perceptions of recreational-use impacts, particularly in lakes with shorelines or isolated coves exhibiting conditions very different from those sampled or evaluated by the sampling volunteers
- (f) impacts to fish or other fauna due to factors unrelated to eutrophication
- (g) PWL or 303(d) listings for other pollutants or portions of the lake not sampled through CSLAP

For these waterbodies, CSLAP can and should continue to be part of an extensive database used to comprehensively evaluate the entirety of the lake and its uses, but absent a more complete dataset, CSLAP data should be used with caution as a sole means for evaluating the lake. Water-quality evaluations, recommended PWL listings, and other extrapolations of the data and analyses should be utilized in this context and by no means should be considered "the last word" on the lake.

I. INTRODUCTION: CSLAP DATA AND YOUR LAKE

Lakes are dynamic and complex ecosystems. They contain a variety of aquatic plants and animals that interact and live with each other in their aquatic setting. As water-quality changes, so too will the plants and animals that live there, and these changes in the food web also may affect water-quality. Water-quality monitoring provides a window into the numerous and complex interactions of lakes. Even the most extensive and expensive monitoring program **cannot completely assess** the water-quality of a lake. However, by looking at some basic chemical, physical, and biological properties, it is possible to gain a greater understanding of the general condition of lakes. CSLAP monitoring is a basic step in overall water-quality monitoring.

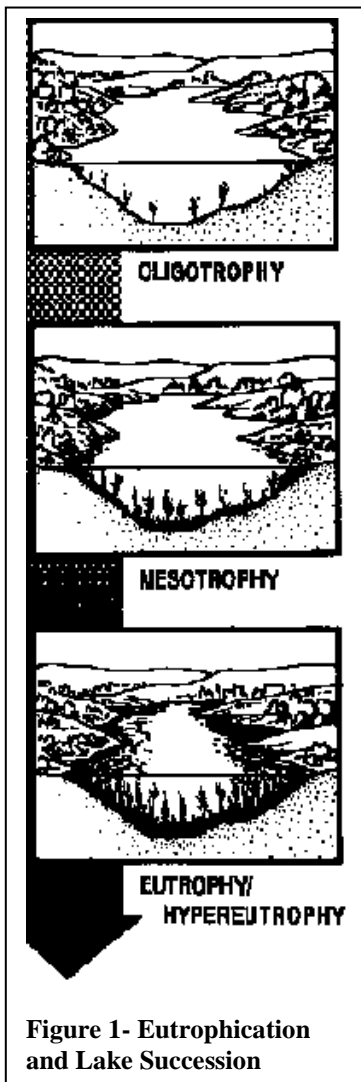
Understanding Trophic States

All lakes and ponds undergo **eutrophication**, an aging process, that involves stages of succession in biological productivity and water-quality (Figure 1). **Limnologists** (scientists who study freshwater systems) divide these stages into **trophic** states. Each trophic state can represent a wide range of biological, physical, and chemical characteristics and any lake may “naturally” be categorized within any of these trophic states. In general, the increase in productivity and decrease in clarity corresponds to an enrichment of nutrients, plant and animal life. Lakes with low biological productivity and high clarity are considered **oligotrophic**. Highly productive lakes with low clarity are considered **eutrophic**. Lakes that are **mesotrophic** have intermediate or moderate productivity and clarity. It is important to remember that eutrophication is a natural process and is not necessarily indicative of man-made pollution.

In fact, some lakes are thought to be “naturally” productive. Trophic classifications are not interchangeable with assessments of water-quality. Water-quality degradation from the perspective of one user may contrast with the perception of favorable conditions by a different lake user. For example, a eutrophic lake may support an excellent warm-water fishery because it is nutrient rich, but a swimmer may describe that same lake as polluted. A lake’s trophic state is still important because it provides lake managers with a reference point to view changes in a lake’s water-quality and they begin to understand how these changes may cause **use impairments** (threaten the use of a lake or swimming, drinking water or fishing).

When human activities accelerate lake eutrophication, it is referred to as **cultural eutrophication**. Cultural eutrophication may result from shoreline erosion, agricultural and urban runoff, wastewater discharges or septic seepage, and other non-point source pollution sources. These can greatly accelerate the natural aging process of lakes, cause successional changes in the plant and animal life within the lake, shoreline and surrounding watershed, and impair the water-quality and value of a lake. They may ultimately extend aquatic plants and emergent vegetation throughout the lake, resulting in the transformation of the lake into a marsh, prairie, and forest. The extent of cultural eutrophication and the corresponding pollution problems can be signaled by significant changes in the trophic state over a short period.

Why is this important? New York State lakes can be affected by a variety of stressors, from acid rain to zebra mussels and almost everything in between. In any given part of the state, some of these stressors are more important than others. For example, there are probably more lakes affected by acid rain than any other pollutant, but these impacts are typically associated with a particular region (the



Adirondacks and Catskills) and particular type of lake (small, high-elevation lakes in basins with thin soils and little buffering capacity). But for most lakes in New York, cultural eutrophication represents the most significant source of pollutants and threat to water-quality. As a result, water-quality indicators related to eutrophication comprise the foundation of most water-quality monitoring programs.

II. CSLAP SAMPLING PARAMETERS

CSLAP monitors several parameters related to the trophic state of a lake, including the clarity of the water, the amount of nutrients in the water, and the amount of algae resulting from those nutrients. Three parameters are the most important measures of eutrophication in most New York lakes: **total phosphorus, chlorophyll *a*** (estimating the amount of algae), and **Secchi disk transparency**. Because these parameters are closely linked to the growth of weeds and algae, they provide insight into “how the lake looks” and its suitability for recreation and aesthetics. Other CSLAP parameters help characterize water-quality at the lake. Each of these sampling parameters is outlined in Figure 3. In addition, CSLAP also uses the responses on the Field Observation Forms to gauge volunteer perceptions of lake water-quality. Most water-quality “problems” arise from impairment of accepted or desired lake uses, or the perception that such uses are somehow degraded. As such, any water-quality monitoring program should attempt to understand the link between perception and measurable quality.

The parameters analyzed in CSLAP provide valuable information for characterizing lakes. By adhering to a consistent sampling protocol provided in the CSLAP Sampling Protocol sampling volunteers collect and use data to assess both seasonal and yearly fluctuations in these parameters and to evaluate the water-quality conditions in their lake. By comparing a specific year's data to historical water-quality information, lake managers can pinpoint trends and determine whether water-quality is improving, degrading or remaining stable. Such a determination answers a first critical question posed in the lake-management process.

Ranges for Parameters Assessing Trophic Status and Eagle Lake

The relationship between phosphorus, chlorophyll *a*, and Secchi disk transparency has been explored by many researchers, to assess the trophic status (the degree of eutrophication) of lakes. Figure 2 shows the ranges for phosphorus, chlorophyll *a*, and Secchi disk transparency (summer median) that are representative for the major trophic classifications:

These classifications are valid for clear-water lakes only (with less than 30 platinum color units). Some humic or “tea color” lakes, for example, naturally have high levels

Figure 2. Trophic Status Indicators

Parameter	Eutrophic	Mesotrophic	Oligotrophic	Eagle Lake
Phosphorus (mg/l)	> 0.020	0.010 - 0.020	< 0.010	0.006
Chlorophyll <i>a</i> (µg/l)	> 8	2- 8	< 2	1.1
Secchi Disk Clarity (m)	< 2	2- 5	> 5	6.3

of dissolved organic material, resulting in color readings that exceed 30 color units. This will cause the water transparency to be lower than expected, given low phosphorus and chlorophyll *a* levels in the lake. Water transparency can also be unexpectedly lower in shallow lakes due to influences from the bottom (or the inability to measure the maximum water clarity due to the visibility of the Secchi disk on the lake bottom). Even shallow lakes with high water clarity, low nutrient concentrations, and little algal growth may also have significant weed growth due to shallow water conditions. While such a lake may be considered unproductive by most water-quality standards, that same lake may experience severe aesthetic problems and recreational impairment related to weeds, not trophic state. Generally, however,

the trophic relationships described above can be used as an accurate "first" gauge of productivity and overall water-quality.

Figure 3. CSLAP Parameters	
PARAMETER	SIGNIFICANCE
Water Temperature (°C)	Water temperature affects many lake activities, including the rate of biological growth and the amount of dissolved oxygen. It also affects the length of the recreational season.
Secchi Disk Transparency (m)	Determined by measuring the depth at which a black and white disk disappears from sight, the Secchi disk transparency estimates the clarity of the water. In lakes with low color and rooted macrophyte ("weed") levels, it is related to algal productivity.
Conductivity (µmho/cm)	Specific conductance measures the electrical current that passes through water, and is used to estimate the number of ions (charged particles). It is somewhat related to both the hardness and alkalinity (acid-buffering capacity) of the water and may influence the degree to which nutrients remain in the water. Generally, lakes with conductivity of <100 µmho/cm are considered softwater, while conductivity readings >300 µmho/cm are found in hardwater lakes.
pH	pH is a measure of the (free) hydrogen ion concentration in solution. Most clearwater lakes must maintain a pH between 6 and 9 to support most types of plant and animal life. Low pH waters (<7) are acidic, while high pH waters (>7) are basic.
Color (true) (platinum color units)	The color of dissolved materials in water usually consists of organic matter, such as decaying macrophytes or other vegetation. It is not necessarily indicative of water-quality but may significantly influence water transparency or algae growth. Color in excess of 30 ptu indicates sufficient quantities of dissolved organic matter to affect clarity by imparting a tannic color to the water.
Phosphorus (total, mg/l)	Phosphorus is one of the major nutrients needed for plant growth. It is often considered the "limiting" nutrient in NYS lakes, for biological productivity is often limited if phosphorus inputs are limited. Nitrogen-to-phosphorus ratios of >25 generally indicate phosphorus limitation. Many lake management plans are centered on phosphorus controls. Phosphorus is reported as total phosphorus (TP)
Nitrogen (nitrate, ammonia, and total (dissolved), mg/l)	Nitrogen is another nutrient necessary for plant growth and can act as a limiting nutrient in some lakes, particularly in the spring and early summer. Nitrogen to phosphorus ratios <10 generally indicate nitrogen limitation (for algae growth). For much of the sampling season, many CSLAP lakes have very low or undetectable levels of one or more forms of nitrogen. It is measured in CSLAP in three forms_ nitrate/nitrite (NO _x) ammonia (NH _{3/4}), and total nitrogen (TN or TDN).
Chlorophyll <i>a</i> (µg/l)	The measurement of chlorophyll <i>a</i> , the primary photosynthetic pigment found in green plants, provides an estimate of phytoplankton (algal) productivity, which may be strongly influenced by phosphorus.
Calcium (mg/l)	Calcium is a required nutrient for most aquatic fauna and is required for the shell growth for zebra mussels (at least 8-10 mg/l) and other aquatic organisms. It is naturally contributed to lakes from limestone deposits and is often strongly correlated with lake buffering capacity and conductivity.

By each of the trophic standards described above, Eagle Lake would be considered to be an **oligotrophic, or highly unproductive, lake**. This has been a consistent assessment for each of the eight CSLAP sampling seasons at the lake. The trophic condition of Eagle Lake will be discussed in greater detail later in this report.

III. CSLAP LAKES

CSLAP sampling began in 1986 on 25 lakes generally distributed throughout the state, and in the following 20 years has expanded to more than 210 lakes. The program was developed primarily to identify water-quality problems, develop long-term databases, and educate lakefront property owners on small lakes with little historical information and few other contemporary studies. However, the program has been utilized by lake residents, lake associations and managers, municipalities, state and federal government and environmental organizations to gain insights about small ponds, large high-profile lakes and multi-use reservoirs from eastern Long Island to the northern Adirondacks, to the western border of New York State. A map showing each of the lakes sampled through CSLAP since 1986 is shown in

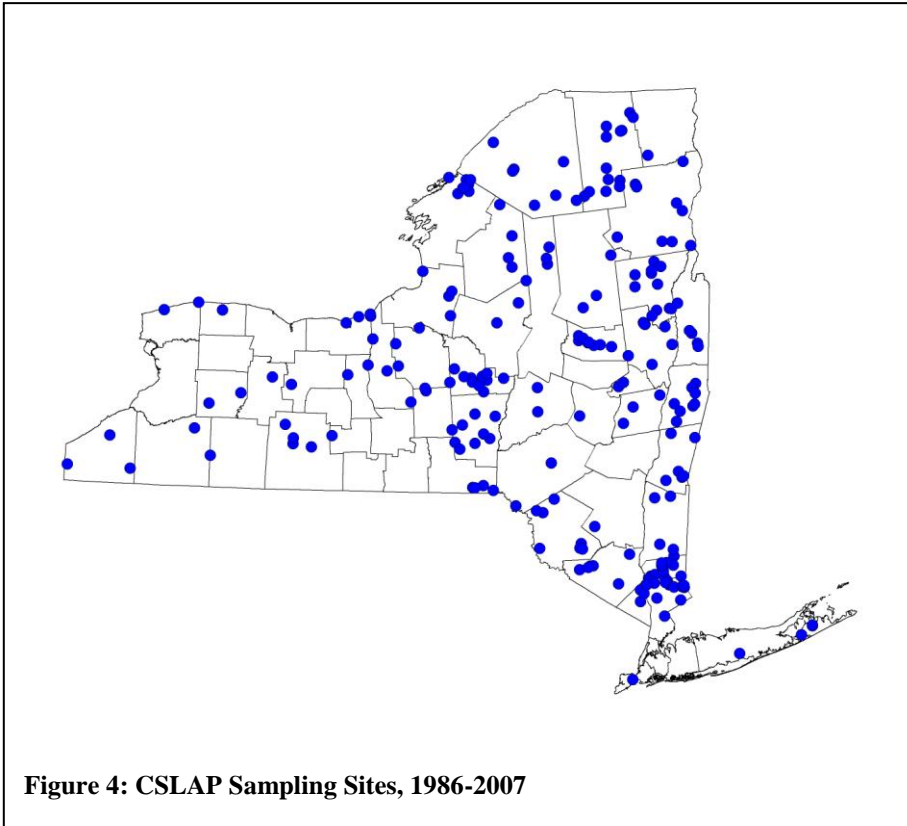


Figure 4. The distribution of lakes roughly matches the distribution of lake associations in the state (or at least those affiliated with the NY Federation of Lake Associations, the largest lake association organization in the state). The relative paucity of CSLAP lakes in the Finger Lakes region reflects the small number of lakes in a region dominated by very large lakes, while the small number of lakes sampled in the Catskills, Long Island, and western NY reflects the shortage of organized lake associations in those areas.

CSLAP lakes have ranged from the very small (five acre Cranberry Lake in the downstate region) to the

great (two state park beaches on Lake Ontario). It has included perhaps the clearest lake in New York State (Skaneateles Lake, one of the Finger Lakes, with as high as 50 feet of water transparency) and several lakes with clarity as low as one foot. There are a large number of lakes used for potable water, as well as those classified only for fishing and non-contact recreation. Some lakes (those on Long Island) sit just above sea level, while others are perched high in the clouds, including Summit Lake in central NY and Twitchell Lake in the Adirondacks, more than 2,000 feet above sea level.

Figures 5a and 5b summarize the variety of lakes sampled through CSLAP. In short, these lakes constitute a comprehensive cross-section of the lake conditions, uses, and settings encountered in New York State.

The typical CSLAP lake is slightly larger than the typical New York State lake and is more likely to be found in the Adirondacks, downstate, and central New York (generally the region bound by the Adirondacks, Finger Lakes, and the downstate region). Specifically, the “average” CSLAP lake is about 125 acres in size, at an elevation of about 1000 feet (300 meters), and can be found in Otsego

County in the Leatherstocking region of New York State, the approximate geographic center of the CSLAP lake population. The typical New York state lake, on the other hand, would be in Fulton County in the southern Adirondacks, and would be about 20 acres in size and perched at an elevation of about 1700 feet (530 meters). The vast majority of lakes in New York state are small, and an inordinate number of lakes are found in the Adirondacks, although there are many other lake-rich regions in the state.

However, this CSLAP profile, as well as the preponderance toward “mid-elevation” regions, is probably more typical of the “lake community” regions of the state. This corresponds to those regions in which large numbers of lakes are heavily populated, which in turn represents lower elevation waterbodies that support siting septic systems and have close proximity to roads and other non-lake communities (comprised of visitors and seasonal lake residents). The relatively higher percentage of Class B lakes in CSLAP and Class C lakes in the rest of the state reflects the large number of uninhabited Class C

lakes in the Adirondacks. These lakes have been classified as Class C lakes, often by default, due in part to the lack of information about historical or contemporary lake uses and water-quality conditions. On the other hand, most of the more densely populated lakes closer to the major population centers of the state have been designated as Class B lakes, owing to their long-standing use for contact recreation. As noted in the individual summary reports for many of the Class C lakes, it is likely that these lakes actively support swimming and other contact recreation, and the state classification system will eventually “catch up” to these recreational uses.

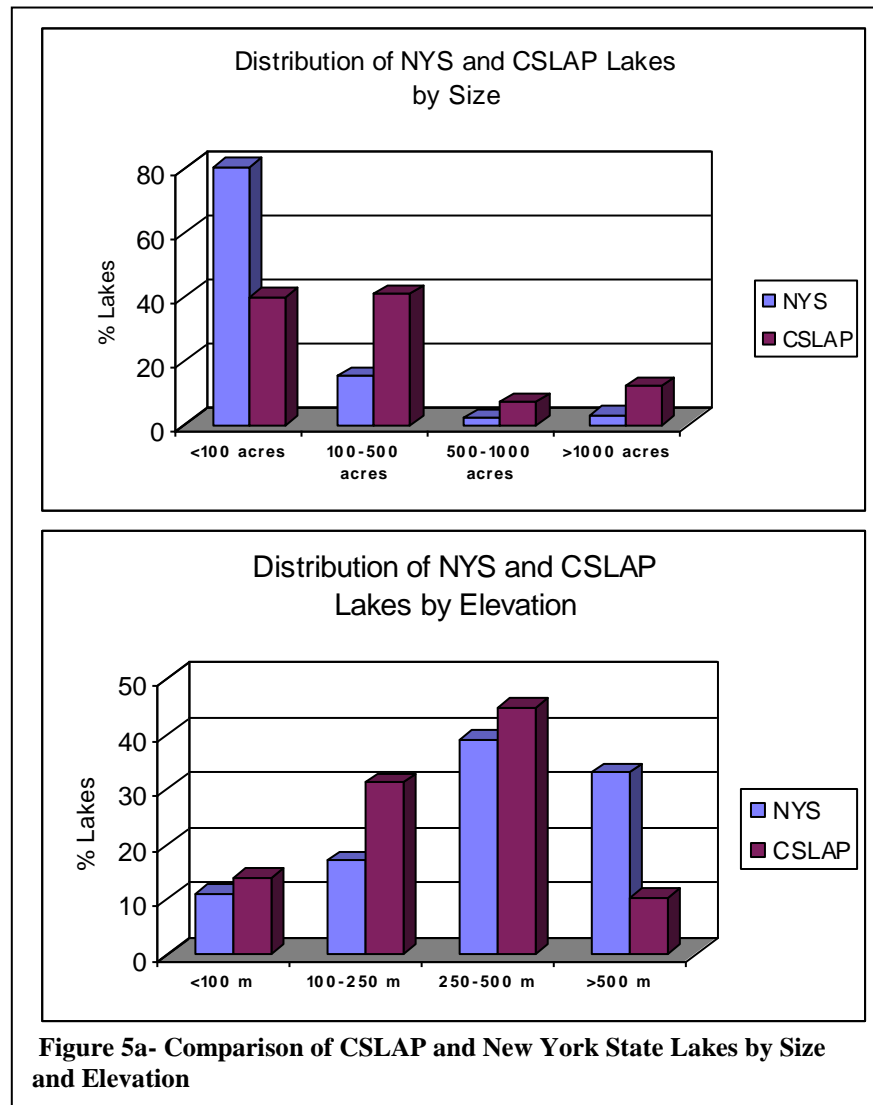
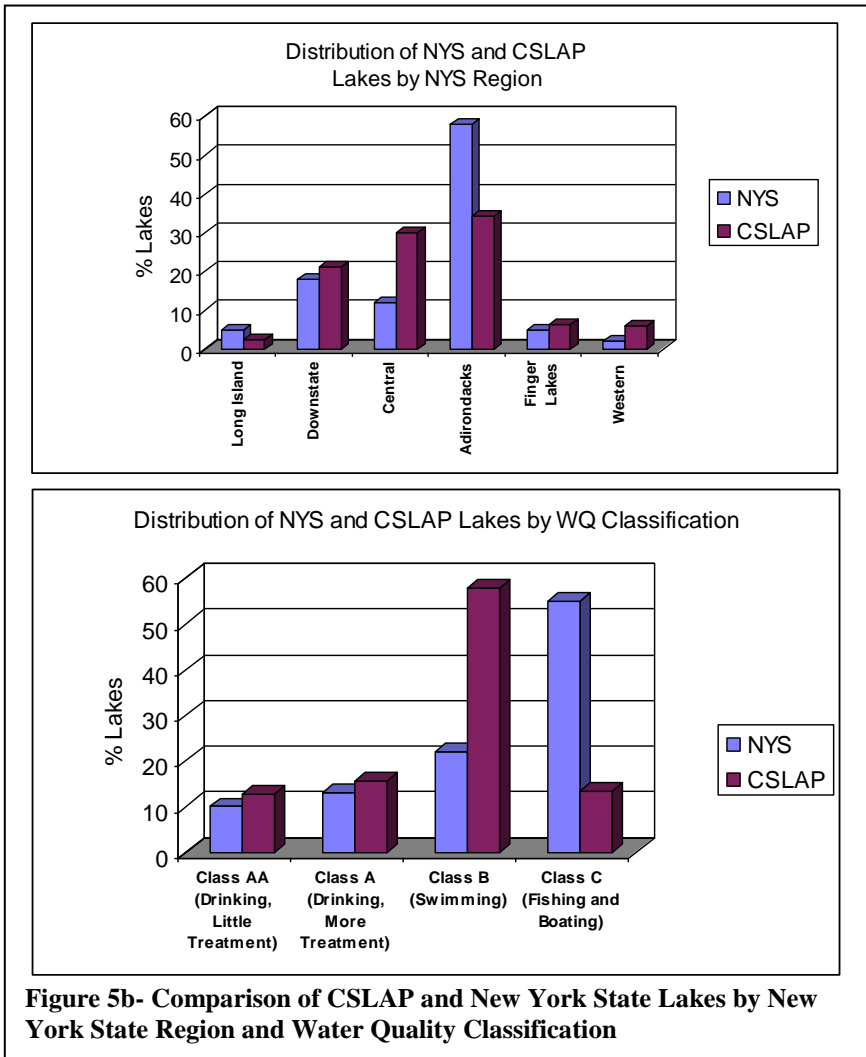


Figure 5a- Comparison of CSLAP and New York State Lakes by Size and Elevation



However, many of the lake distribution categories displayed in Figures 5a and 5b indicate similar cross-sections of lakes. There are relatively few lakes in Long Island, Western New York and the Finger Lakes region, whether looking at the entirety of New York state or just those lakes in CSLAP. There are also few Class AA and A lakes—those used for potable water intake—in New York state or within the CSLAP database.

The distribution of lakes in these categories does suggest that CSLAP lakes are mostly comparable to other New York State lakes, and that an evaluation of CSLAP data may serve as a reasonable surrogate for statewide water-quality evaluations, particularly since CSLAP serves as the primary long-term database maintained and supported by New York State.

IV: EAGLE LAKE- BACKGROUND INFORMATION

Eagle Lake is a 422 acre, class B lake found in the Town of Ticonderoga in Essex County, within the Eastern Adirondack region of New York State. Eagle Lake was first sampled as part of CSLAP in 2000. Figure 6 shows the location of Eagle Lake. It is one of 9 CSLAP lakes among the >270 lakes found in Essex County, and one of 24 CSLAP lakes among the >470 lakes and ponds in the Upper Hudson River drainage basin. Eagle Lake is a Class B lake; this means that the best intended use for the lake is for contact recreation—swimming and bathing—and for non-contact recreation—boating and aesthetics. These “categories” will be used to evaluate water-quality conditions later in the report.

CSLAP samples have been collected from the deepest part of the lake, corresponding to a depth of about 38 feet (11.5 meters). Most lakes with a maximum depth of >20 feet are thermally stratified, so bottom samples have been collected through CSLAP at the lake.

Historical Water-Quality Information for Eagle Lake

Eagle Lake was sampled by the NYSDEC in 1999 as part of the Lake Classification and Inventory (LCI) survey, the Division of Water ambient lake monitoring program. It was also sampled by the New York State Conservation Department (the predecessor to the NYSDEC) as part of the Biological Survey of the Upper Hudson River basin in 1932. The results from both of these surveys, at least as they pertain to water quality indicators measured through CSLAP, are reported in Table 1. The LCI results from 1999 indicate that water transparency was slightly higher than in the contemporary CSLAP study of the lake, although nutrient and algae levels were comparable. These indicators suggest that water quality conditions from 1999 were similar to those measured in recent years through CSLAP.

The Biological Survey was intended to evaluate water quality conditions as they relate to fisheries management, so much of the information collected cannot be easily compared to the CSLAP dataset. The lake was described as follows:

“Eagle Lake comprises an area of 409 acres with little shallow water and few weed beds. The rocky bottom slopes rapidly away from the shores. Eighty to ninety percent of the lake is over twenty feet deep and temperature and oxygen relationships are satisfactory for fish everywhere..... ..few weed beds...This is a rather narrow lake with rocky shores. The weed areas are limited mostly to a long narrow bay which includes the part of the lake west of the highway crossing. This bay gradually merges with Paragon Creek. This whole area, for a distance of over one mile, supports a very luxuriant growth of plants”

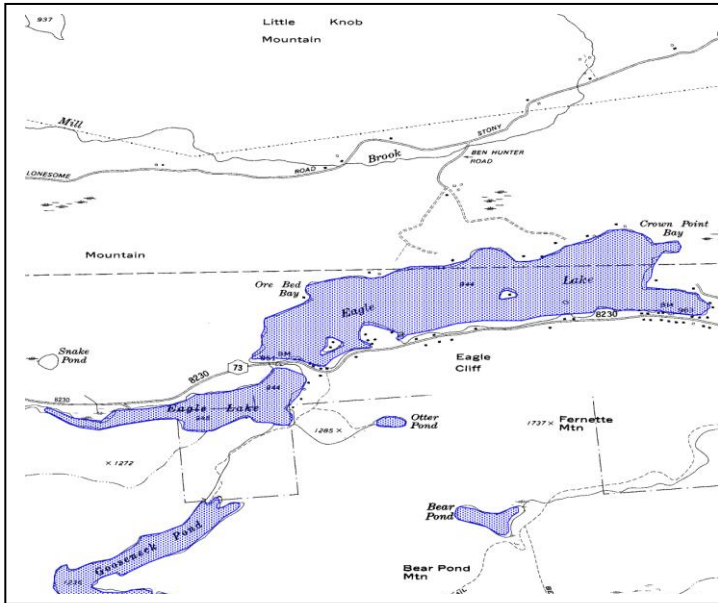
The water quality data suggests that water clarity and pH readings were probably similar to those in the present studies, suggesting water quality changes may not have been significant. Oxygen levels were somewhat depleted very near the lake bottom, a phenomenon found on occasion in recent studies.

Water quality data have not been collected through any of the NYSDEC Division of Water stream monitoring programs.

Historical Fisheries Information for Eagle Lake

Eagle Lake is stocked annually with about 3,800 9” brown trout, and 1,500 7” brown trout. Fish species in the lake include black crappie, brown bullhead, brown trout, largemouth bass, northern pike, pickerel, rock bass, smallmouth bass, pumpkinseed sunfish, lake whitefish, and yellow perch.

General statewide fishing regulations are applicable in Eagle Lake. In addition, the open season on sunfish and yellow perch lasts all year, with no size or take limits. There is also a year-long open season on trout, with no size limits, but a daily take limit of 5.



Permitted Facilities Associated with Eagle Lake

There appear to be a few facilities or activities on Eagle Lake that requires permits or is otherwise regulated by the NYSDEC; the map below shows facilities on or near the lake (represented by “derricks”). These corresponds to a shoreline improvement projects.

V. NEW YORK STATE, CSLAP AND EAGLE LAKE WATER-QUALITY DATA: 1986-2006

Overall Summary:

Although water-quality conditions at each CSLAP lake have varied each year since 1986, and although detailed statistical analyses of the entire CSLAP dataset has not yet been conducted, general water-quality trends can be evaluated after 5-21 years' worth of CSLAP data from these lakes. Overall (regional and statewide) water-quality conditions and trends can be evaluated by a variety of different means. Each of the tested parameters ("analytes") can be evaluated by looking at how the analyte varies from year to year from the long-term average ("normal") condition for each lake, and by comparing these parameters across a variety of categories, such as across regions of the state, across seasons (or months within a few seasons), and across designated best uses for these lakes. Such evaluations are provided in the second part of this summary, via figures 7 through 17. The annual variability is expressed as the difference in the annual average (mean) from both the long-term average and the normal variability expected from this long-term average. The latter can be presented as the "standard error" (SE, calculated here within the 95% confidence interval)—one standard error away from the long-term average can be considered a "moderate" change from "normal," with a deviation of two or more standard errors considered to be a "significant" change. For each of these parameters, the percentage of lakes with annual data falling within one standard error from the long-term average are considered to exhibit "no change," with the percentage of lakes demonstrating moderate to significant changes also displayed on these graphs (figures 7a through 17a). Annual changes in these lakes can also be evaluated by standard linear regressions- annual means over time, with moderate correlation defined as $R^2 > 0.33$, and significant correlation defined as $R^2 > 0.5$. These methods are described in greater detail in Appendix D. Assessments of weather patterns—whether a given year was wetter or drier than usual—accounts for broad statewide patterns, not weather conditions at any particular CSLAP lake. As such, weather may have very different impacts at some (but not most) CSLAP lakes in some of these years.

Long-term trends can also be evaluated by looking at the summary findings of individual lakes and attempting to extrapolate consistent findings to the rest of the lakes. Given the (non-Gaussian) distribution of many of the water-quality parameters evaluated in this report, non-parametric tools may be the most effective means for assessing the presence of a water-quality trend. However, these tools do not indicate the magnitude of the trend. As such, a combination of parametric and non-parametric tools is employed here to evaluate trends. The Kendall tau ranking coefficient has been utilized by several researchers and state water-quality agencies to evaluate water-quality trends via non-parametric analyses and is utilized here. For parametric analyses, best-fit analysis of summer (June 15 through September 15) averages for each of the eutrophication indicators can be evaluated, with trends attributable to instances in which deviations in annual means exceed the deviations found in the calculation of any single annual mean. "Moderate" change is defined as $\tau > 0.33$, and "significant" change is defined as $\tau > 0.5$. It has been demonstrated in many of these programs that long-term trend analyses cannot be utilized to evaluate lake datasets until at least five years' worth of data have been collected.

As of 2007, there were 157 CSLAP lakes that have been sampled for at least five years; of these, 113 were sampled within the last five years. The change in these lakes is demonstrated in figures 7 and 8; figures 7a through 7l indicate "moderate" long-term change, while figures 8a through 8l indicate "significant" long-term change. When these lakes are analyzed by this combination of parametric and non-parametric analyses, these data suggest that while most NYS lakes have not demonstrated a significant change (either τ or $R^2 > 0.5$) or even a moderate changes (τ or $R^2 > 0.33$).

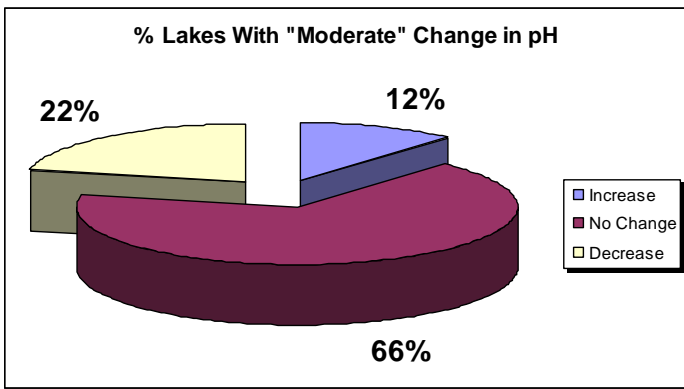


Figure 7a. %CSLAP Lakes Exhibiting Moderate Long-Term Change in pH

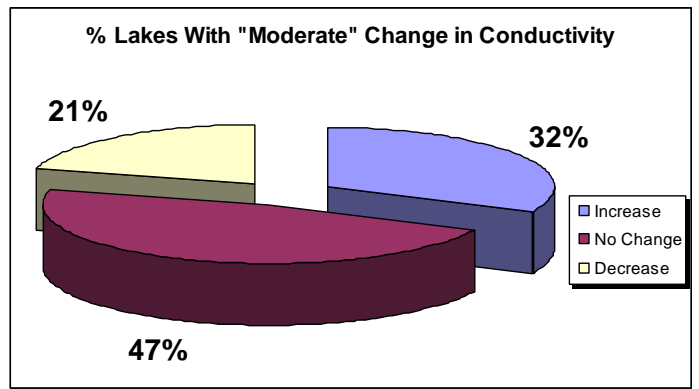


Figure 7b. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Conductivity

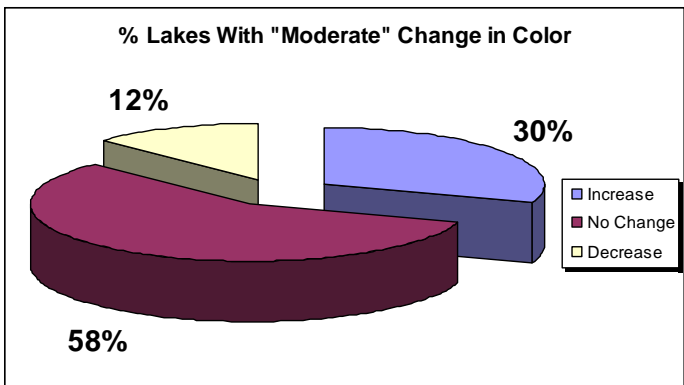


Figure 7c. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Color

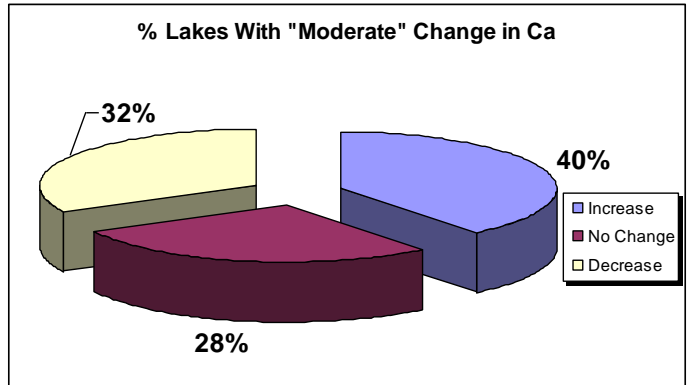


Figure 7d. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Calcium

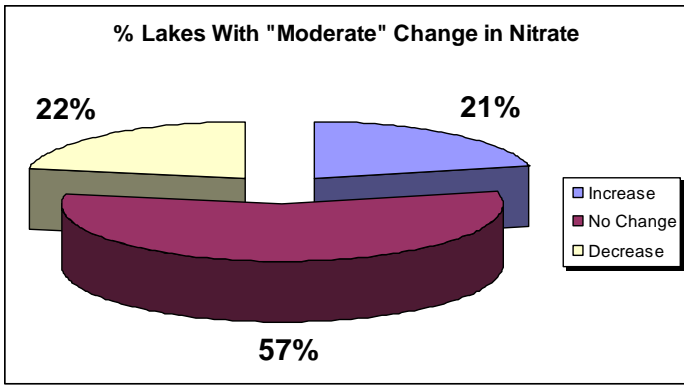


Figure 7e. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Nitrate

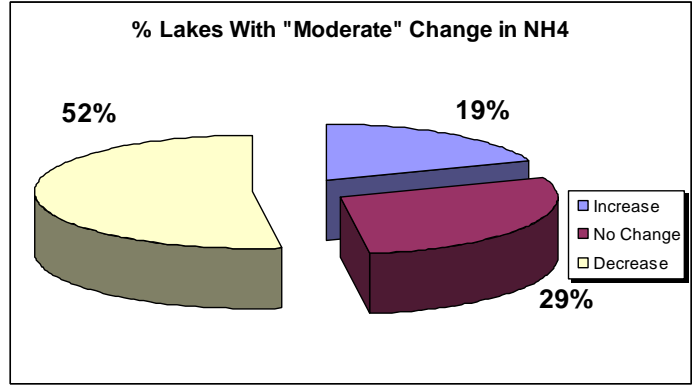


Figure 7f. %CSLAP Lakes Exhibiting Moderate Long-Term Changes in Ammonia

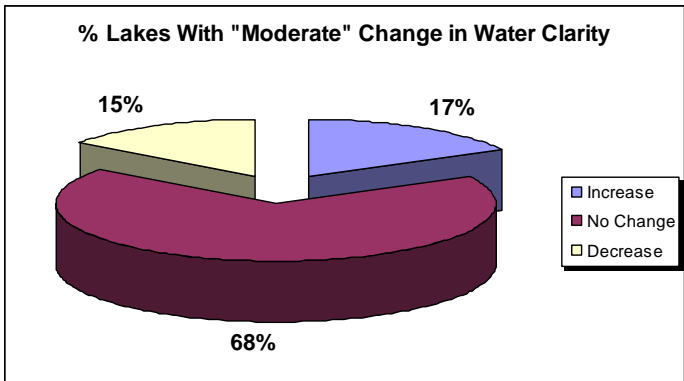


Figure 7g. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Water Clarity

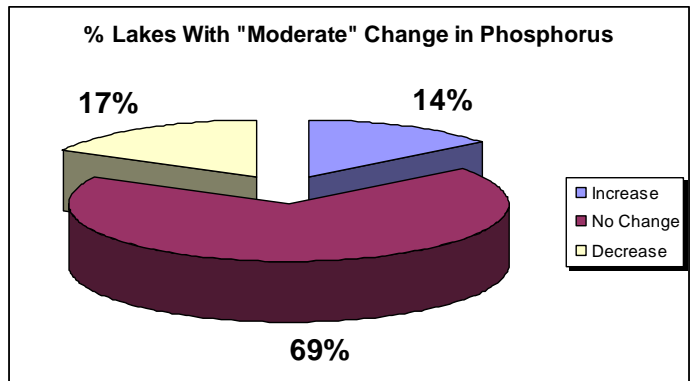


Figure 7h. %CSLAP Lakes Exhibiting Moderate Long-Term Changes in Phosphorus

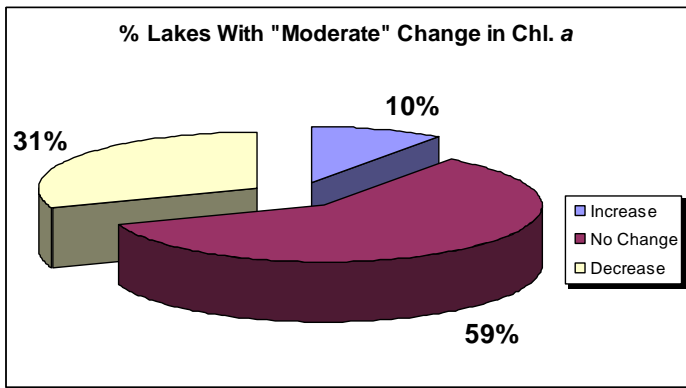


Figure 7i. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Chlorophyll a

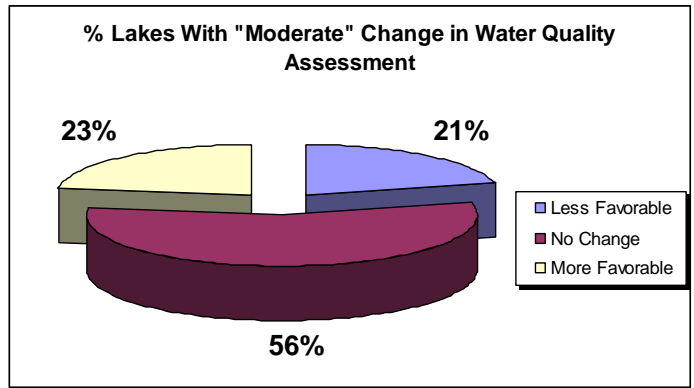


Figure 7j. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Water-quality Assessment

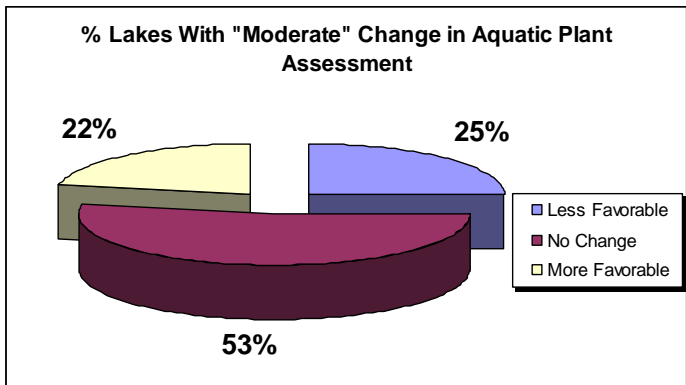


Figure 7k. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Aquatic Plant Assessment

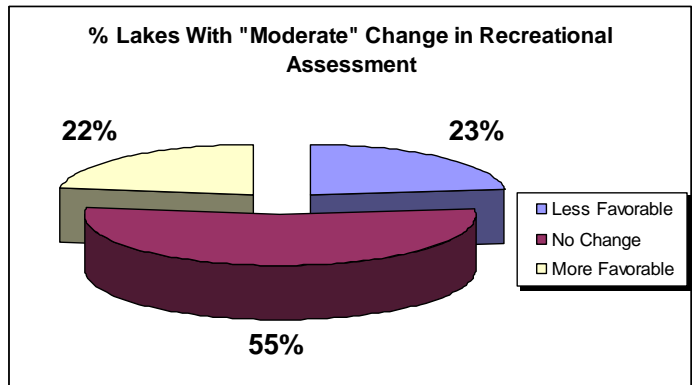


Figure 7l. %CSLAP Lakes Exhibiting Moderate Long-Term Change in Recreational Assessment

Some of the lakes sampling through CSLAP have demonstrated a moderate change since CSLAP sampling began in 1986, at least for some of the sampling parameters measured through CSLAP. In general, between 50% and 65% of the CSLAP lakes have not exhibited even moderate changes. Some of the parameters that have exhibited moderate changes may not reflect actual water-quality change. For example, it appears that the increase in color (Figure 7c) and decrease in nitrate (Figure 7e) and chlorophyll *a* (Figure 7i) is probably due to the shift in laboratories, even though the analytical methods are comparable. The increase in conductivity (Figure 7b) and decrease in pH (Figure 7a) are probably real phenomena—both changes were evident to some degree prior to the shift in laboratories, and both are largely predictable. The difference between the increase and decrease in the other sampling parameter (or between more favorable and less favorable conditions) does not appear to be important and probably indicates random variability.

Figures 8a through 8l indicate that, not surprisingly, “substantial” change is less common. Substantial change follows the same patterns as discussed above with the evaluation of “moderate” change in CSLAP lakes, except that the percentage of CSLAP lakes not exhibiting significant change is much higher, rising to about 65-80% of these lakes. For those CSLAP lakes exhibiting substantial change, it is most apparent in the same parameters described above. About 25% of the CSLAP lakes have exhibited a substantial increase in conductivity, consistent with a broad (and expected) successional pattern, in which lakes generally concentrate materials washed in from the surrounding watershed (and as the runoff itself concentrates materials as these watersheds move from forested to more urbanized, whether via residential development or other uses. The comparison between figures 8b and 8e through 8h indicate that this has not (yet) translated into higher nutrient loading into lakes.

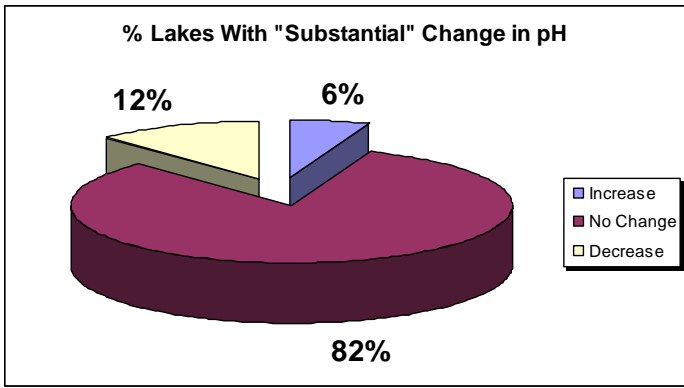


Figure 8a. %CSLAP Lakes Exhibiting Substantial Long-Term Change in pH

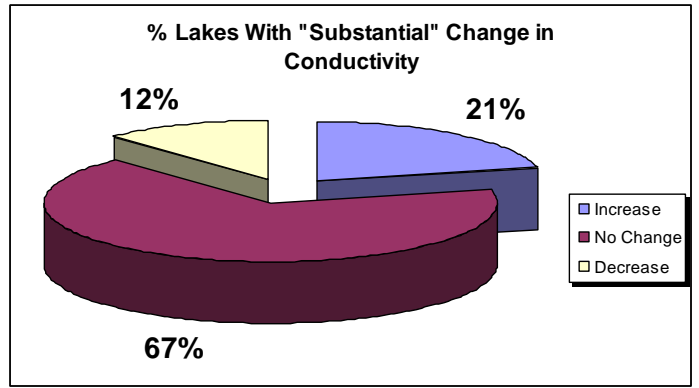


Figure 8b. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Conductivity

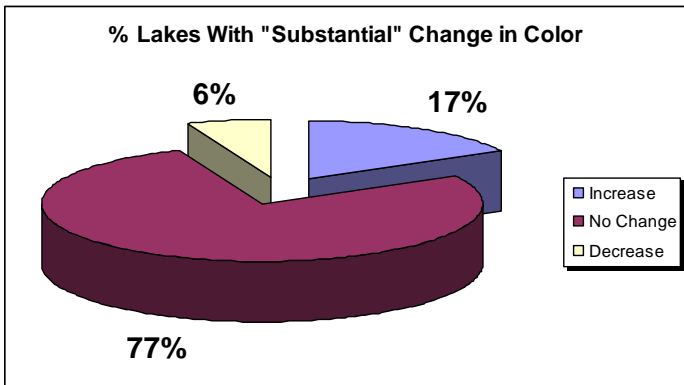


Figure 8c. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Color

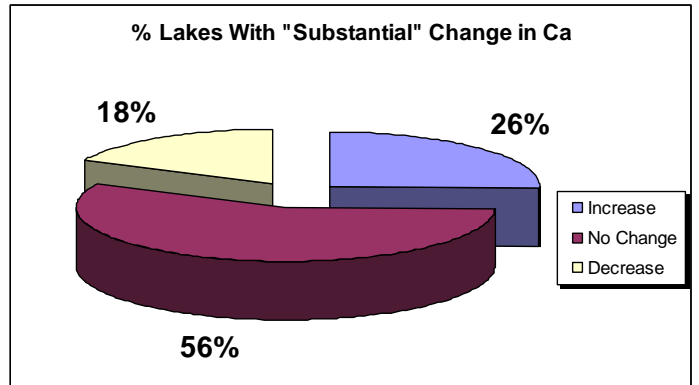


Figure 8d. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Calcium

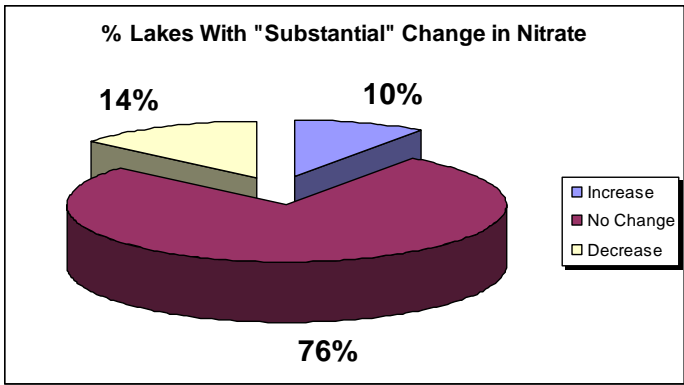


Figure 8e. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Nitrate

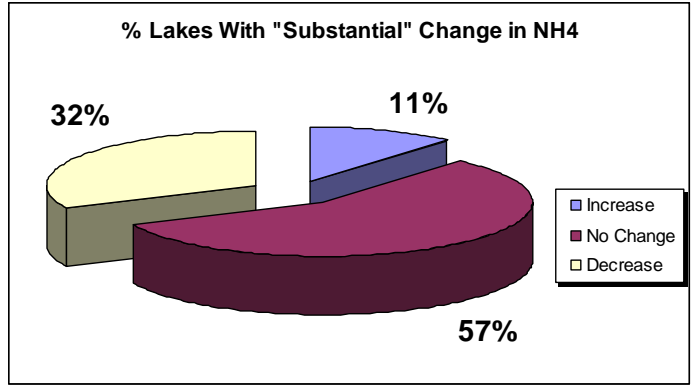


Figure 8f. %CSLAP Lakes Exhibiting Substantial Long-Term Changes in Ammonia

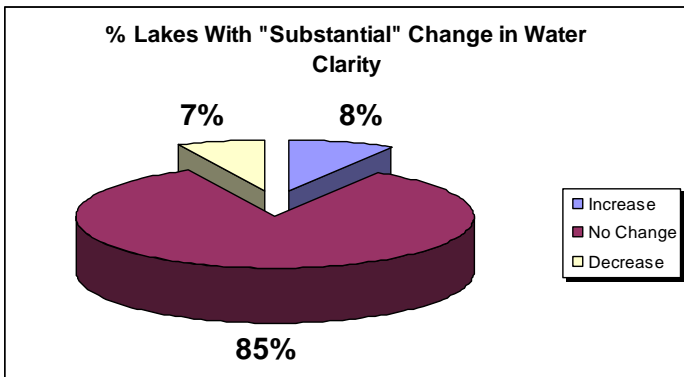


Figure 8g. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Water Clarity

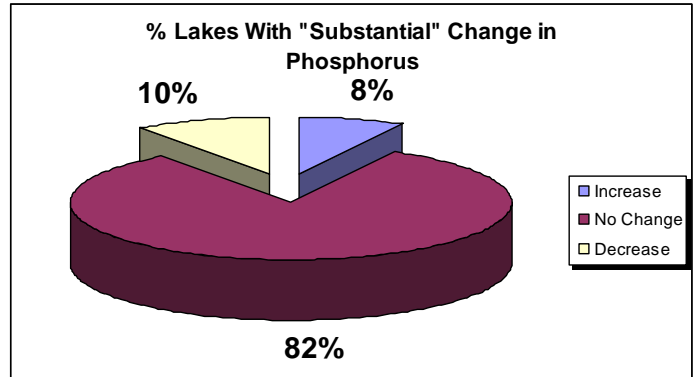


Figure 8h. %CSLAP Lakes Exhibiting Substantial Long-Term Changes in Phosphorus

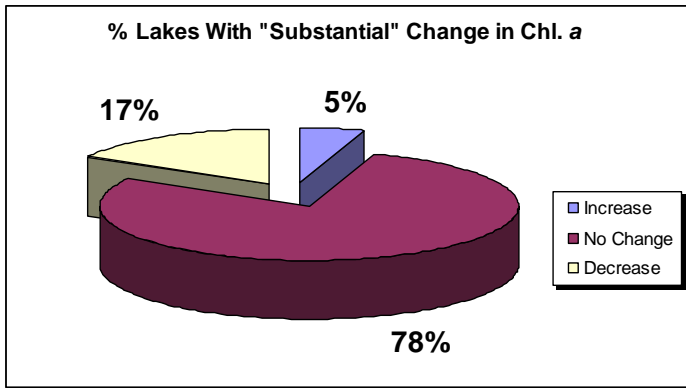


Figure 8i. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Chlorophyll a

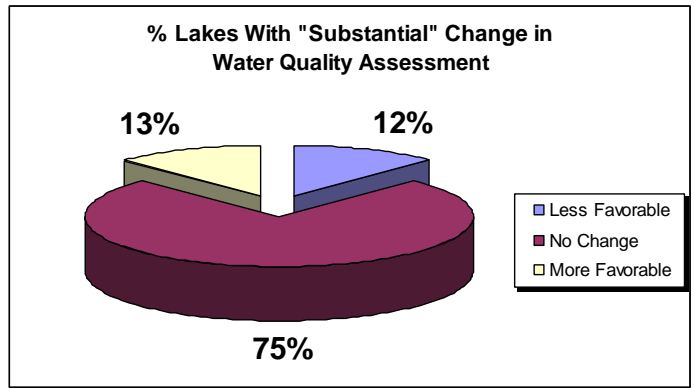


Figure 8j. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Water-quality Assessment

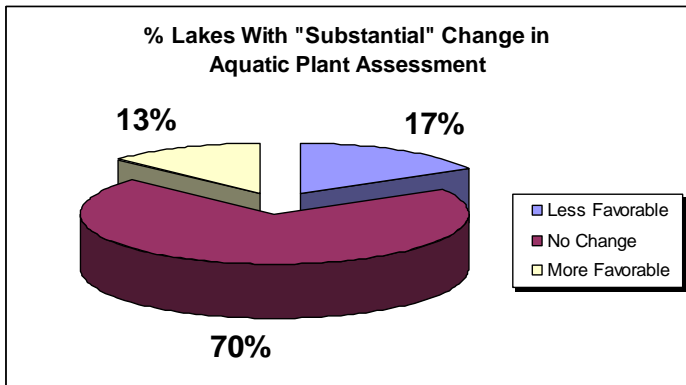


Figure 8k. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Aquatic Plant Assessment

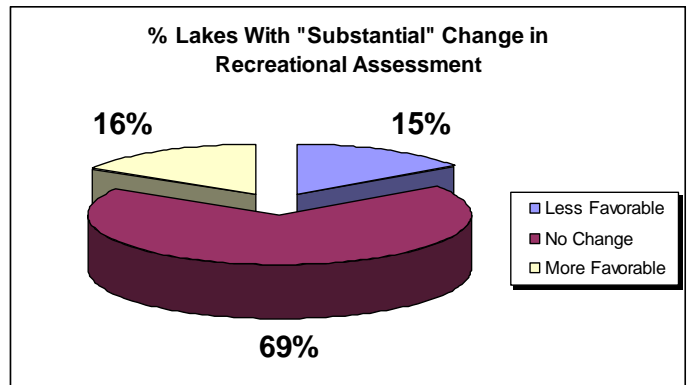


Figure 8l. %CSLAP Lakes Exhibiting Substantial Long-Term Change in Recreational Assessment

As noted above, there does not appear to be any clear pattern between weather and water-quality changes, although some connection between changes in precipitation and changes in some water-quality indicators is at least alluded to in some cases. However, all of these lakes may be the long-term beneficiaries of the ban on phosphorus in detergents in the early 1970s, which, with other local circumstances (perhaps locally more “favorable” weather, local stormwater or septic management, etc.), has resulted in less productive conditions. Without these circumstances, water-quality conditions in many of these lakes might otherwise be more productive in the creeping march toward aging, eutrophication, and succession (as suggested from the steady rise in conductivity). In other words, the higher materials loading into these lakes may be largely balanced by a reduction in nutrients within the corresponding runoff.

The drop in pH in NYS lakes has been studied at length within the Adirondacks and may continue to be attributable on a statewide basis to acid rain, which continues to fall throughout the state. The CSLAP dataset is not adequate to evaluate any ecological changes associated with higher lake acidity, and it is certainly worth noting that the slight drop in pH in most CSLAP lakes does not bring these lakes into an acidic status (these lakes have, at worse, become slightly less basic). In addition, for lakes most susceptible to acidification, laboratory pH is only an approximation of actual pH. Fully accurate pH readings require field measurements using very specialized equipment, although for most lakes with even modest buffering capacity, laboratory pH is a good estimate of *in situ* pH readings. So while the decrease in pH in some CSLAP lakes should continue to be watched, it does not appear to be a cause for concern, at least relative to the low pH in small, undeveloped, high-elevation lakes within the Adirondack Park.

Lake perception has changed more significantly than water-quality (except conductivity). None of the lake perception indicators—water-quality, weeds, or recreation—have varied in a consistent manner, although variability is more common in each of these indicators. The largest change is in recreational assessments, with about one third of all lakes exhibiting substantial change and nearly half demonstrating moderate change. A more detailed analysis of these assessments (not presented here) indicates that the Adirondacks have demonstrated more “positive” change than other regions of the state, due to the perception that aquatic weed densities have not increased as significantly (and water-quality conditions have improved in some cases). However, the rapid spread of *Myriophyllum spicatum* into the interior Adirondacks will likely reverse this “trend” in coming years, and it is not clear if these “findings” can be extrapolated to other lakes within the Adirondack Park.

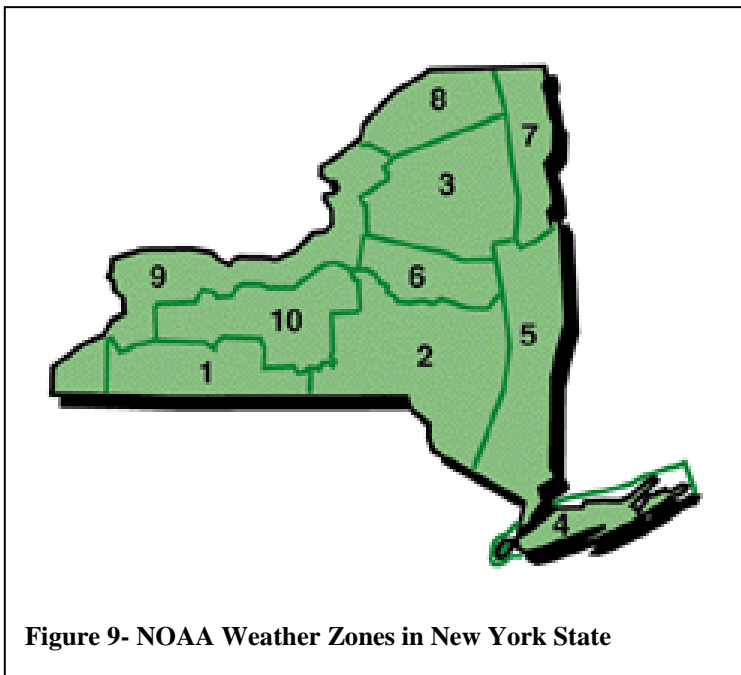


Figure 9- NOAA Weather Zones in New York State

Larger trends and observations about each of the CSLAP sampling parameters are presented below in figures 10 through 21. Information about general precipitation and runoff patterns—whether a particular year was wet or dry—is reported to provide a basis for understanding the connection between weather and water quality for lakes in New York state. It is clear that weather patterns are highly variable within the state. While this is also apparent down at the individual lake scale—storms can fall at a lake but not a neighboring lake—the National Oceanographic and Atmospheric Administration (NOAA) has established ten weather zones in New York state corresponding to regions exhibiting similar weather patterns. Weather data for the state can be summarized by each of these zones, in

an attempt to fine-tune individual lake analyses to local weather data.

The individual parameter summaries provided in figures 10-20 correspond to the predominant weather patterns found from 1986 to 2006 in the state. A code can be located above the columns for each year; a “↑” corresponds to wetter (>50%) than normal weather, while “↓” corresponds to drier (<50%) than normal weather, and “0” corresponds to normal weather. In this code, the first symbol corresponds to the winter and spring precipitation, and the second symbol corresponds to summer precipitation. So, for example, a code of “↑↓” corresponds to a wet spring and dry summer, while “00” corresponds to normal spring and summer precipitation. While ideally the individual parameter summaries and weather summaries could be delineated by weather zone, the CSLAP lake dataset is not sufficient large for most of these weather zones to generate statistically meaningful data summaries. However, these weather zone data are used in the individual lake data summaries in **Section IV: Detailed Eagle Lake Water Quality Summary.**

Eagle Lake is in NOAA weather zone 3, the Northern Plateau. The precipitation patterns for this zone are summarized in Section IV and Figures 24 through 31.

Statewide and Eagle Lake Regional Weather Patterns

Weather patterns in New York state have varied significantly from year to year since at least 1986. This may be a response to global climatic change, since greater weather variance has been observed by both climatologists and casual observers.

Using the criteria above (wetter = >50% more precipitation than the long-term average, drier = >50% less precipitation than normal) and equally weighing each of the 10 NOAA weather zones in New York state, Table 1 shows the winter (January through March) and spring (April through June) precipitation and “summer” (June through September) precipitation patterns for New York state and the NOAA zone corresponding to Eagle Lake. Summer was defined here to overlap with spring to include the entirety of the sampling season for most CSLAP lakes.

Year	Statewide Avg: Winter-Spring / Summer	NOAA Zone 3 Avg: Winter-Spring / Summer
1986	Normal / Wet	Normal / Wet
1987	Dry / Normal	Dry / Normal
1988	Very Dry / Normal	Very Dry / Normal
1989	Wet / Normal	Normal / Very Wet
1990	Very Wet / Normal	Very Wet / Dry
1991	Normal / Normal	Normal / Normal
1992	Normal / Wet	Normal / Normal
1993	Wet / Normal	Wet / Dry
1994	Wet / Normal	Wet / Normal
1995	Very Dry / Normal	Very Dry / Normal
1996	Very Wet / Normal	Wet / Normal
1997	Normal / Normal	Wet / Normal
1998	Very Wet / Normal	Very Wet / Normal
1999	Normal / Normal	Normal / Normal
2000	Very Wet / Normal	Very Wet / Normal
2001	Normal / Normal	Normal / Normal
2002	Very Wet / Dry	Wet / Dry
2003	Normal / Wet	Normal / Normal
2004	Dry / Very Wet	Normal / Normal
2005	Normal / Normal	Dry / Wet
2006	Wet / Wet	Wet / Normal

Table 1: Statewide and NOAA Zone 3 Weather Patterns

occurred in this region in the last ten to fifteen years.

The weather data in Table 1 shows that wetter than normal summers have occurred in three of the last four years, although more variable weather patterns have occurred in the winter and spring. The wettest years have been 1990, 1996, 1998, 2004 and 2006, while the driest years were 1988 and 1995. The only dry seasons since 1995 were the winter of 2004 and the summer of 2002.

Data from the Northern Plateau—the northern Adirondack region that includes Eagle Lake—have indicated more normal weather patterns, or at least that seasonal averages have been closer to normal over most of the last four years. The wettest years generally occurred from 1996 through 2000, especially in the winter and spring, while the driest years were 1988 and 1995. It should be noted that only one dry summer (2002) and one dry winter (2005) has

pH

Annual Variability:

The pH of most CSLAP lakes has consistently been well within acceptable ranges for most aquatic organisms during each sampling season. The average pH has not varied significantly from one sampling season to the next, although pH was highest in 1988 (one of the driest years) and 1992, and lowest in 1987 and 2004. pH readings were slightly lower than normal in 1996 but higher than normal in 2006, the two wettest years, and were not significantly different than normal in 1995, perhaps the driest year. There do not appear to be any significant annual pH trends in the CSLAP dataset, at least as evaluated in Figure 10a. 90% of all samples had pH between 6.5 and 8.5 (the state water-quality standards); 6% of samples have pH > 8.5, and 4% have pH < 6.5.

What Was Expected in 2007?

2007 was a relatively wet year, at least in most of the state during much of the summer sampling season. There is not a strong correlation between weather and pH during most of the CSLAP sampling seasons. However, pH readings have slightly higher in the last few years, perhaps due to phenomena unrelated to weather. This suggests that pH readings may be slightly higher than normal in 2007.

What Happened at Eagle Lake in 2007?

pH readings in Eagle Lake have been higher than usual in the last three years, but there does not appear to be any clear long-term trend or connection between weather and pH in Eagle Lake.

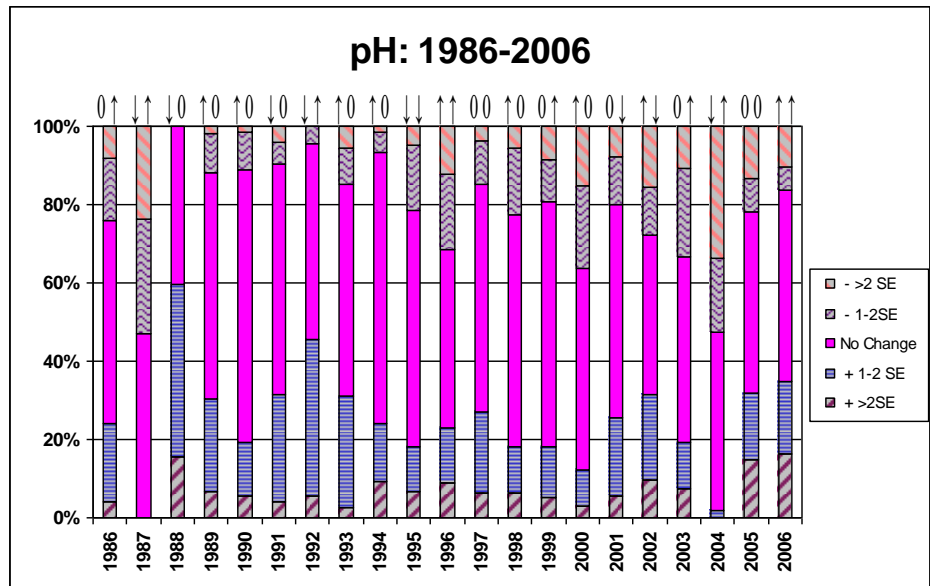


Figure 10a. Annual Change from "Normal" pH in CSLAP Lakes (SE = Standard Error)

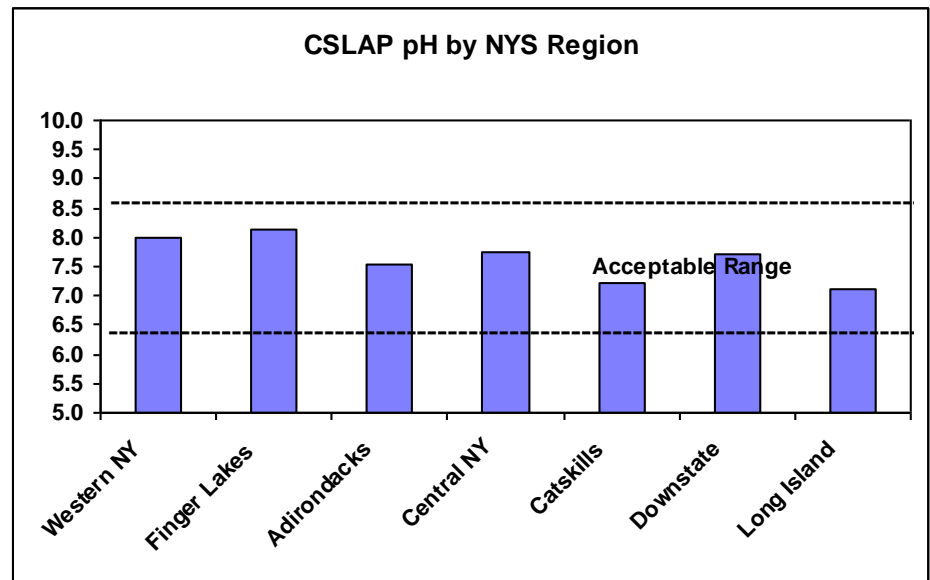


Figure 10b. pH in CSLAP Lakes by NYS Region

Statewide Variability:

As expected, pH readings are lowest in the high-elevation regions (Adirondacks and Catskills) or Long Island, which has primarily shallow and slightly colored lakes, and the highest in regions with relatively high conductivity (western NY and the Finger Lakes region). All of these readings are consistently within the acceptable range for most aquatic organisms. However, the CSLAP dataset does not reflect the low pH found in many high elevation NYS lakes overlying granite and poorly buffered soils, because the typical CSLAP lake resides in geological settings (primarily limestone) that allow for residential development. In other words, pH is one of the few CSLAP sampling parameters that does not yield comparable results when comparing CSLAP results to overall NYS results, because CSLAP lakes are not really representative of the typical NYS lake as related to pH.

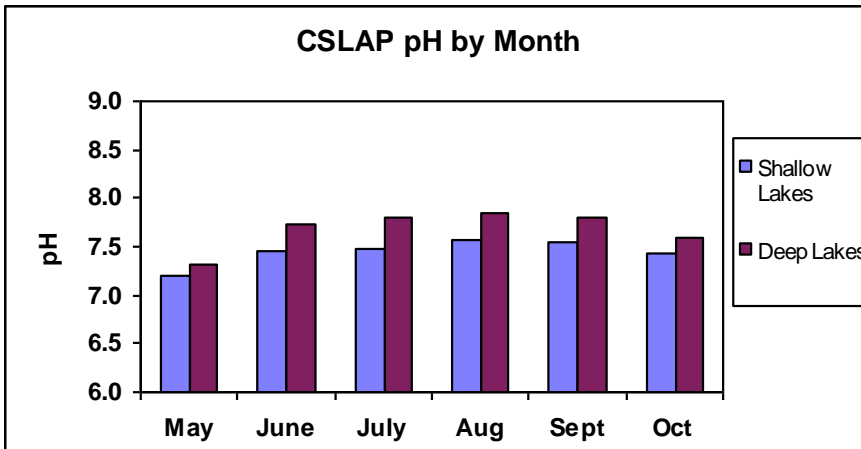


Figure 10c. pH in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

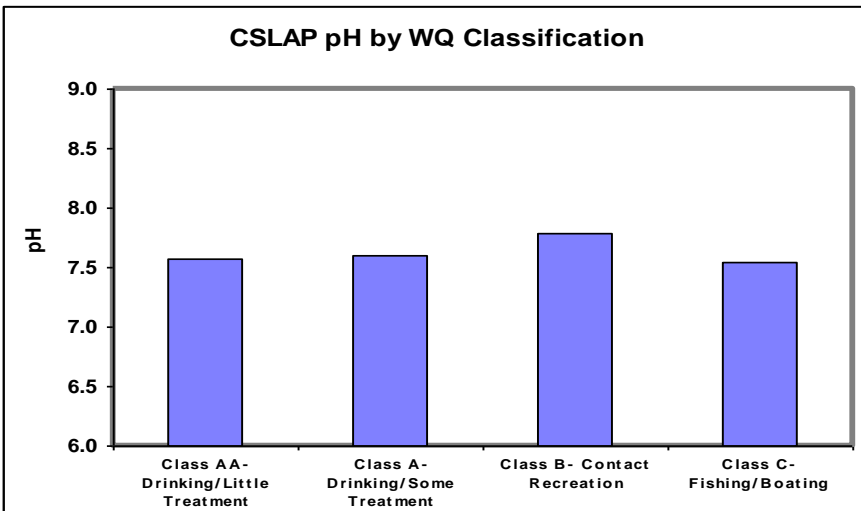


Figure 10d. pH in CSLAP Lakes by Lake Use usage.

Seasonal Variability:

pH readings tend to increase slightly during the course of the summer, due largely to increasing algal photosynthesis (which consumes CO₂ and drives pH upward), although these seasonal changes are probably not significant. Low pH depressions are most common early in the sampling season (due to lingering effects from snowpack runoff), and high pH spikes occur mostly in mid- to late summer.

Lake-Use Variability:

pH does not vary significantly from one lake use to another, although in general, pH readings are slightly higher for lakes used primarily for contact recreation (Class B). However, this is probably more reflective of geographical differences (there are relatively more Class B CSLAP lakes in higher pH regions, and more Class A lakes in lower pH regions) than any inherent link between pH and lake

Conductivity

Annual Variability:

There appeared to be a clear trend toward increasing lake conductivity from 1986 through 2004. While conductivity often increased after storm events, the highest conductivity occurred in drier years, since as 1995, with lower readings occurring in wetter years, such as 1996 and 1998. This suggests that other factors may have influenced the rise in conductivity over this period. However, conductivity was much lower than usual in 2006, a wet year, with about half of the CSLAP lakes exhibiting conductivity readings at least one standard error lower than usual.

What Was Expected in 2007?

2007 was a relatively wet year, at least in most of the state during much of the summer sampling season. The relationship between conductivity and precipitation is not consistent, since higher conductivity often occurs immediately after storm events and in drier years. Therefore, it is anticipated that conductivity readings may again be lower than normal, although higher conductivity readings may occur after heavy storms.

What Happened at Eagle Lake in 2007?

Conductivity readings in Eagle Lake were lower in 2007 than in any previous sampling seasons, although additional data will be needed to determine if a long-term trend toward decreasing conductivity is occurring.

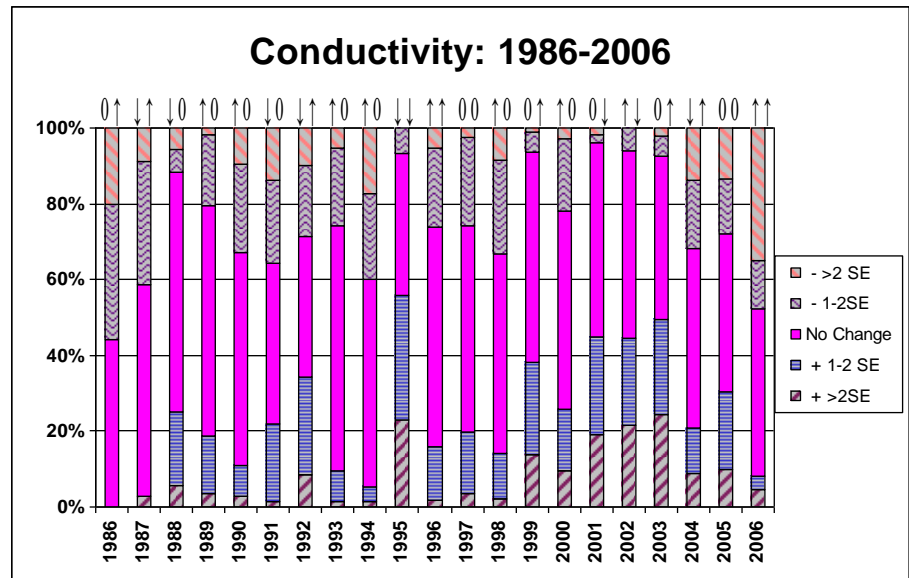


Figure 11a. Annual Change from "Normal" Conductivity in CSLAP Lakes (SE = Standard Error)

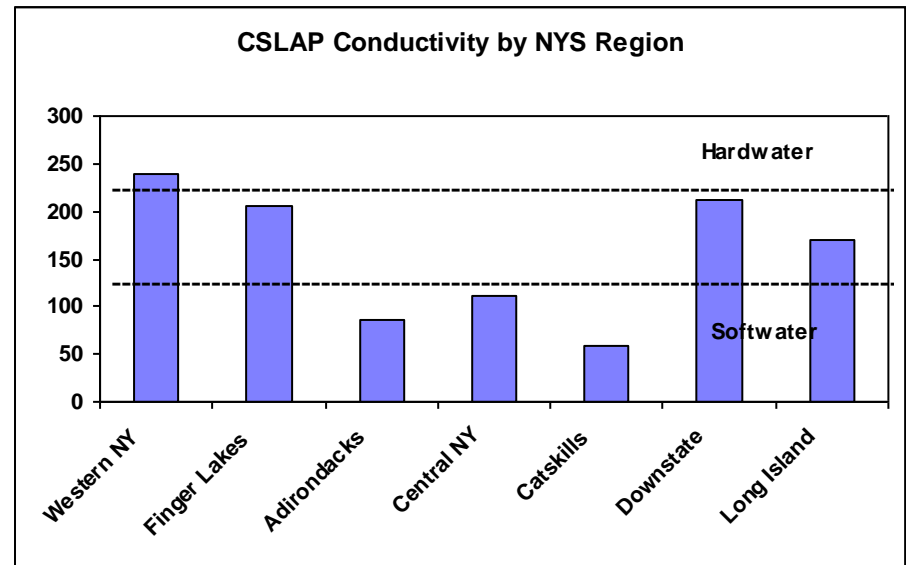


Figure 11b. Conductivity in CSLAP Lakes by NYS Region

Statewide Variability:

Although “hardwater” and “softwater” are not consistently defined by conductivity, in general lakes in the Adirondacks and Catskills have lower conductivity (softer water), and lakes downstate, in western NY, and in the Finger Lakes region have higher conductivity (harder water). These regional differences are due primarily to surficial geology and “natural” conditions in these areas. However, within each of these broad geographical areas, there are usually some lakes with higher conductivity and some lakes with lower conductivity readings.

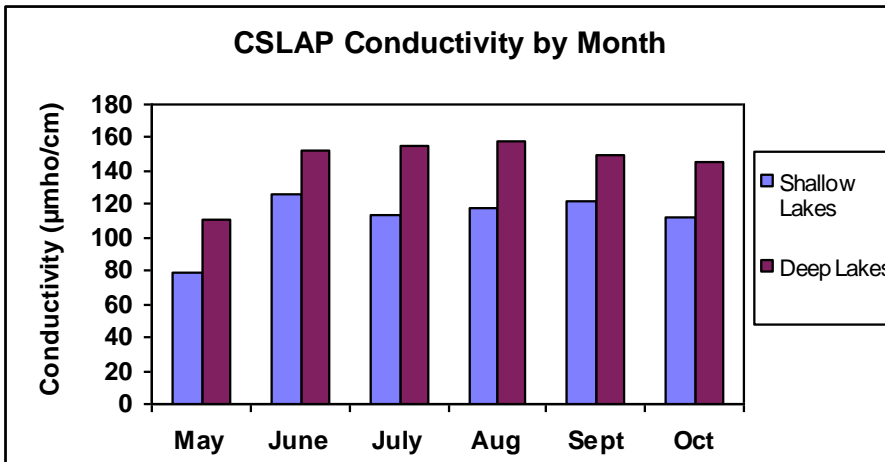


Figure 11c. Conductivity in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

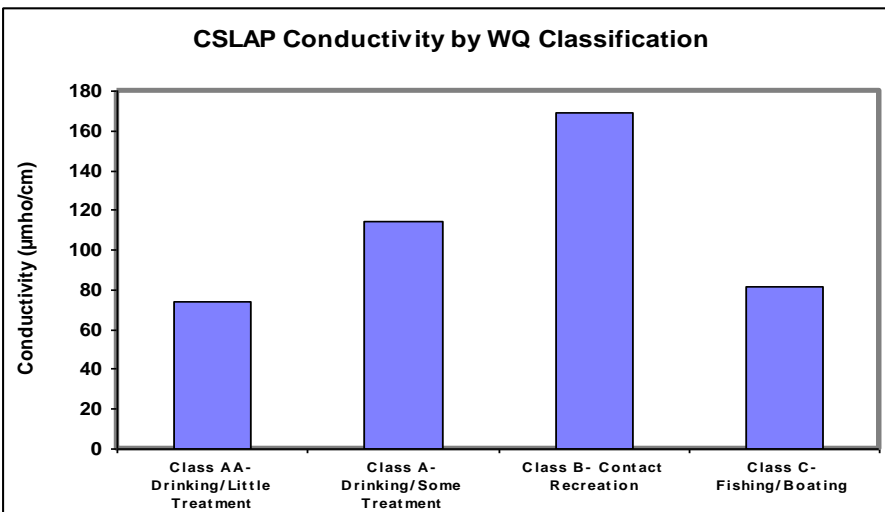


Figure 11d. Conductivity in CSLAP Lakes by Lake Use

Seasonal Variability:

Conductivity readings are higher in the summer than in the late spring in many CSLAP lakes. These readings decreased in deep lakes in the late summer and fall but remained fairly steady in shallow lakes during this period (actual readings within specific lakes, however, may often vary significantly from week to week). Although lake destratification (turnover) brings bottom waters with higher conductivity to the lake surface in deeper lakes, conductivity readings dropped in the fall. It is possible that fully mixed conditions may be missed in some NYS lakes by discontinuing sampling after the end of October. Conductivity readings overall were higher in deep lakes, although this may be an artifact of the sampling set (there are more CSLAP deep lakes in areas that “naturally” have harder water).

Lake-Use Variability:

Conductivity readings are substantially higher for lakes used primarily for contact recreation (Class B) and are somewhat higher for lakes used for drinking water with some treatment (Class A). However, this is probably more reflective of geographical differences (there are relatively more softwater CSLAP lakes in the Adirondacks, which tend to have more Class A or Class AA lakes, at least in CSLAP, and more Class B lakes are found in hardwater regions) than any *de facto* connection between conductivity and lake usage.

Color

Annual Variability:

Color readings in many CSLAP lakes have increased in recent years. One of the years with the lowest color readings, 1995, was the driest of the CSLAP sampling seasons, while the highest color occurred in two of the wettest years (2004 and 2006). Most lake samples (88%) correspond to water-color readings too low (< 30 ptu) to significantly influence water clarity, although nearly 30% of the samples in 2007 corresponded to color readings exceeding this threshold. Color readings were much higher in 2006 than in any other CSLAP sampling season. Given that color readings were also highest in four of the last five years, the increase in color may be attributable in part to the shift in laboratories, which occurred prior to the 2003 sampling season. The higher color has also been coincident with wet summers and/or wet winters during most of these years (the lower color in 2005 may have been due to more normal weather patterns).

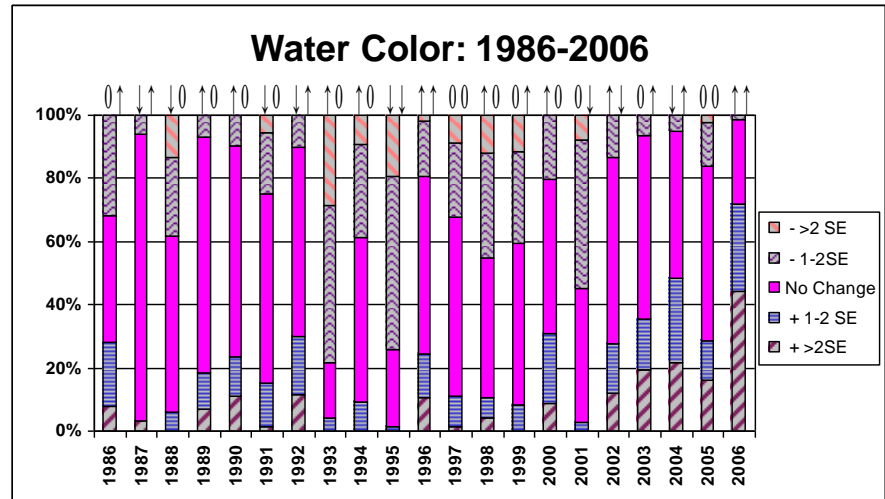


Figure 12a. Annual Change from “Normal” Color in CSLAP Lakes (SE = Standard Error)

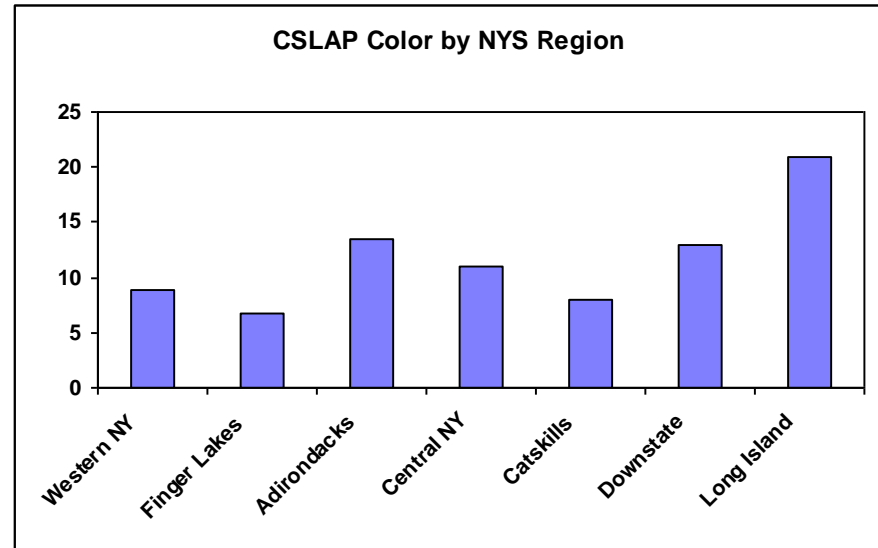


Figure 12b. Color in CSLAP Lakes by NYS Region

What Was Expected in 2007?

As noted above, color readings have generally been higher during wet years, and readings have been higher in four of the last five years, perhaps due to slightly different analytical methodology. Since 2007 generally corresponded to a wet year, it is likely that color readings in 2006 will at least be higher than the long-term average.

What Happened at Eagle Lake in 2007?

Water color readings in Eagle Lake were lower in the last two years than in 2005 and 2004, although there has been a trend toward increasing water color since 2000. The higher color in the last four years has generally corresponded to wetter weather.

Statewide Variability:

Water color is highest in Long Island and the Adirondacks, and lowest in the Finger Lakes, Catskill and western NY regions. This is mostly coincident with the statewide conductivity distribution (with softwater lakes more likely to be colored). The CSLAP dataset may be a representative cross-section of NYS lakes as related to color.

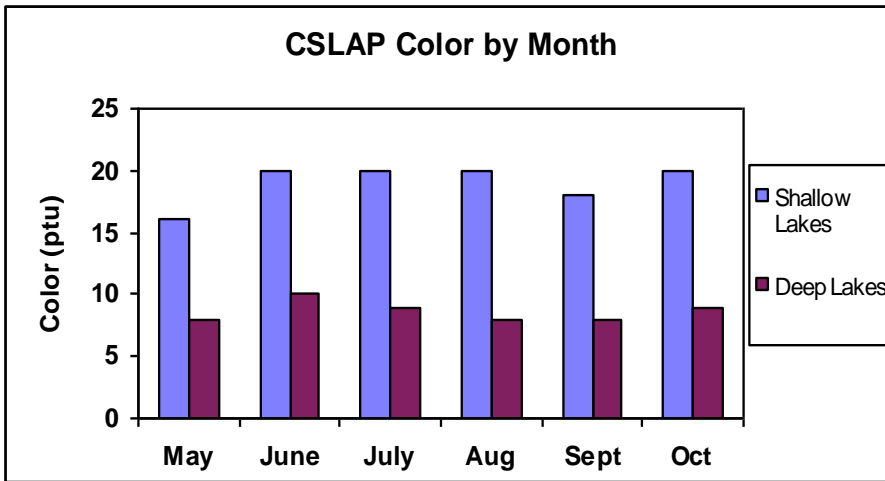


Figure 12c. Color in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

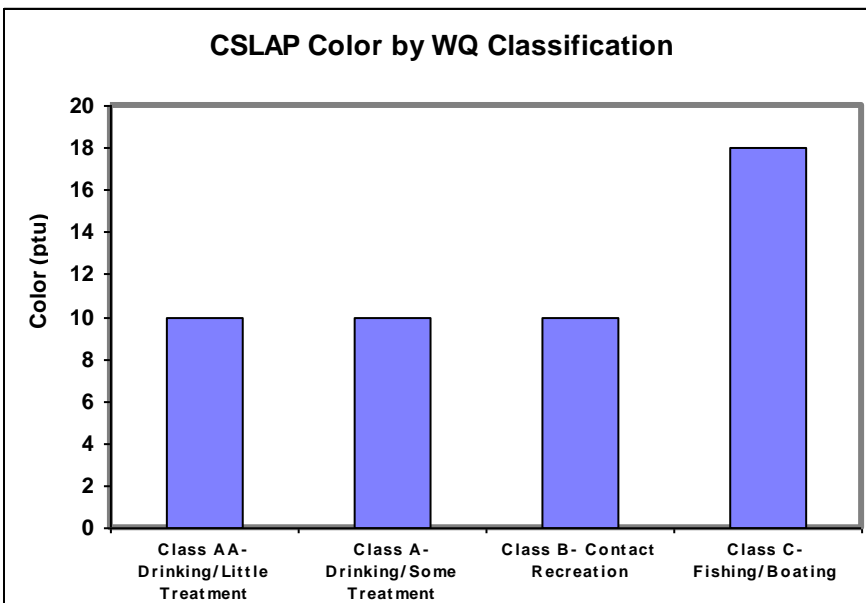


Figure 12d. Color in CSLAP Lakes by Lake Use

Seasonal Variability:

Color readings are significantly higher in shallow lakes than in deepwater lakes; these readings increase from spring to summer in these shallower lakes (perhaps due to dissolution of organic material, including algae, and wind-induced mixing during the summer) and then drop off again in late summer into the fall. Color generally follows the opposite trend in deeper lakes, with slightly decreasing color readings perhaps due to more particle setting in the summer and remixing in the fall, although the seasonal trend in the deeper lakes is not as pronounced as in shallow lakes.

Lake-Use Variability:

Color readings are substantially higher for lakes used primarily for non-contact recreation (Class C), but this is probably more reflective of morphometric differences, for Class C lakes tend to be shallow lakes (mean depth = 4 meters), while the other classes tend to

be deeper lakes (mean depth = 9 meters). However, the elevated color readings correspond to elevated levels of dissolved organic matter and may also reflect impediments (via economically viable water treatment, aesthetics, and potential formation of hazardous compounds during chlorination) to the use of these waters for drinking.

Calcium

Annual Variability:

Calcium was analyzed for the first time in 2002, so long-term analyses are limited by the relative lack of data. Readings were highest in 2004 and lowest in 2002; the latter corresponded to a year in which calcium was analyzed by a different laboratory. While 2004 was the only year since 2001 with a relatively dry winter, it is not known if there is a connection between winter and spring weather and summer calcium readings. Likewise, it is also not known if the drier summer in 2002 triggered the lower calcium readings. Additional data will help to determine if calcium levels are changing, but these data suggest that a significant long-term trend is not apparent.

What Was Expected in 2007?

There did not appear to be a strong predictive connection between weather and calcium levels in the lake, notwithstanding the observations about spring and summer precipitation levels in 2002 and 2004. So the calcium readings in 2007 are likely to be “unexpected”.

What Happened at Eagle Lake in 2007?

Calcium readings in 2007 were similar to those measured in most previous CSLAP sampling seasons.

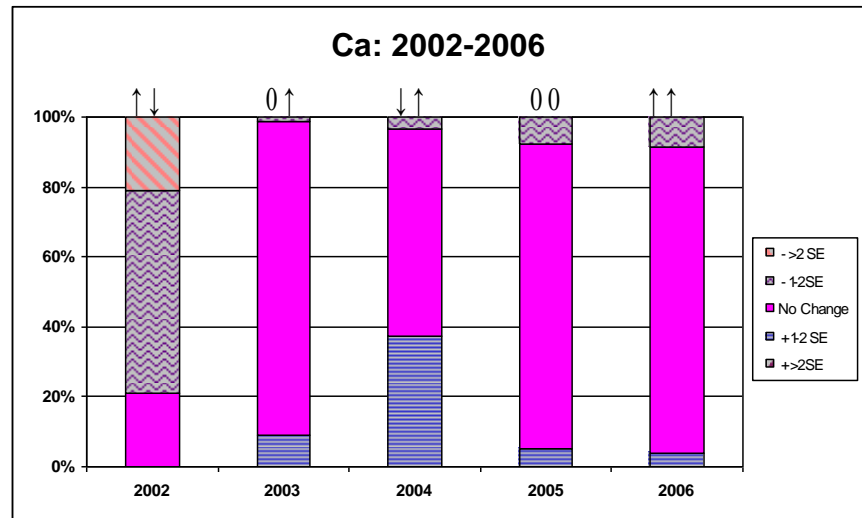


Figure 13a. Annual Change from “Normal” Calcium in CSLAP Lakes (SE = Standard Error)

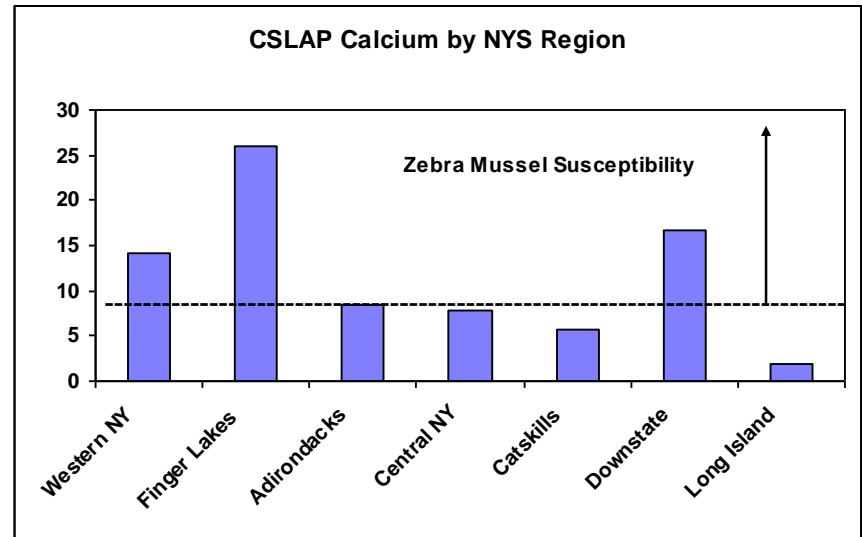


Figure 13b. Calcium in CSLAP Lakes by NYS Region

Statewide Variability:

Calcium readings are highest in the Finger Lakes, western, and downstate New York regions. This is mostly coincident with the statewide conductivity distribution (since the ions that contribute to conductivity are often found in the same proportions as calcium). While the former two regions are already populated by zebra mussel-infested lakes, the downstate region at present does not possess many lakes with these exotic organisms. The data in Figure 13b suggest many of the downstate lakes may be susceptible to zebra mussels, while some lakes in many of the other regions may have already crossed the susceptibility threshold. The CSLAP dataset is most likely a reasonably representative cross-section of NYS lakes as related to calcium.

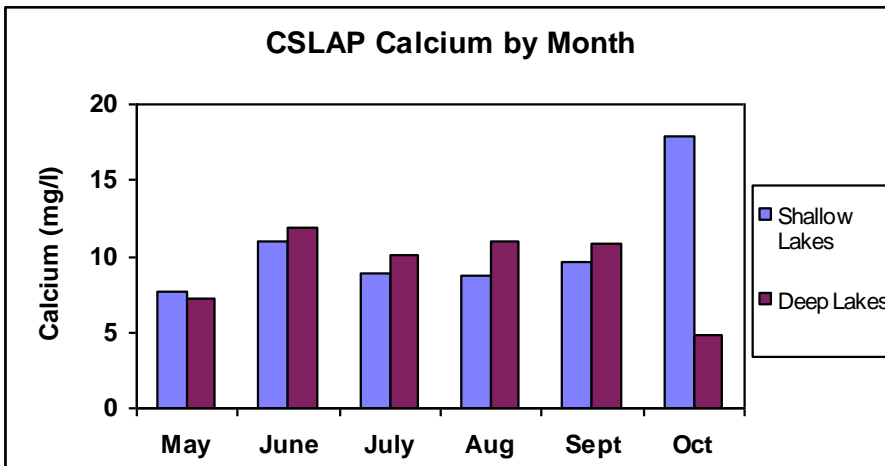


Figure 13c. Color in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

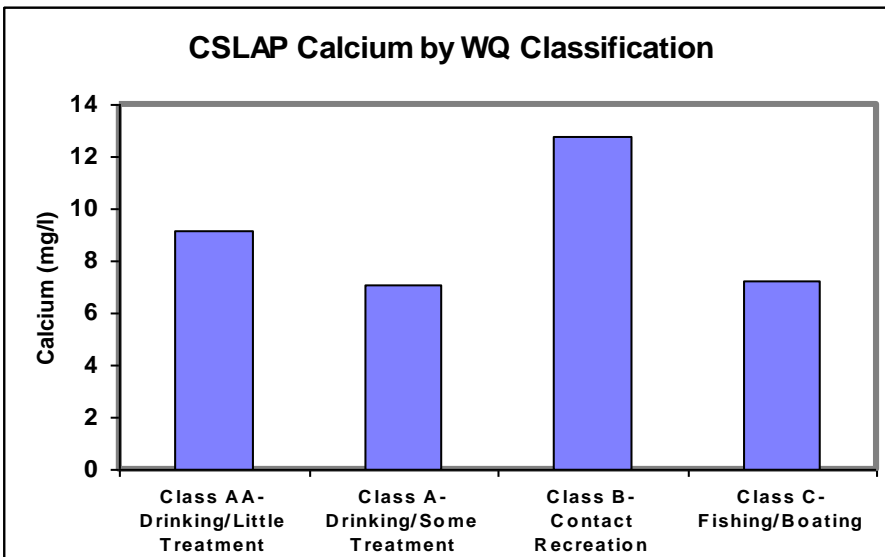


Figure 13d. Color in CSLAP Lakes by Lake Use

Seasonal Variability:

Calcium readings appear to increase during the sampling season at many shallow CSLAP lakes, with the highest readings occurring in the fall. The opposite appears to occur with deeper lakes, but it is more likely that the seasonal distribution noted in Figure 13c reflects a relatively larger number of low calcium lakes sampled in the fall rather than an actual fall decrease in calcium levels in these lakes.

Lake-Use Variability:

Calcium readings are substantially higher for lakes used primarily for contact recreation (Class B), but this is probably more reflective of regional differences, for Class B lakes are more likely to be found in the regions with higher conductivity and calcium readings, such as the Finger Lakes region, downstate, and western New York. As noted earlier, many of the Class C

lakes in CSLAP are found in the Adirondacks, where calcium readings are lower.

Nitrate

Annual Variability:

Evaluating nitrate in CSLAP lakes is confounded by the relative lack of nitrate data for many sampling seasons (it was analyzed in water samples at a lower frequency, or not at all, for many years), the high number of undetectable nitrate readings, and some changes in detection levels. The limited data indicated that nitrate was highest in 1986 and 1989, two early CSLAP years in which nitrate was analyzed more frequently (including a relatively large number of early season samples), and in 2004 and 2005, which corresponded to the use of a new analytical tool. Readings were lowest in 1995, 2002 and 2003. Although nitrate levels are probably closely related to winter and spring precipitation levels (due to the higher nitrate readings in snowpacks), this is not apparent from Figure 14a. There was not a predictable relationship between either winter runoff or summer rains and nitrate levels. No readings have approached the state water-quality standard (= 10 mg/l) in any CSLAP sample.

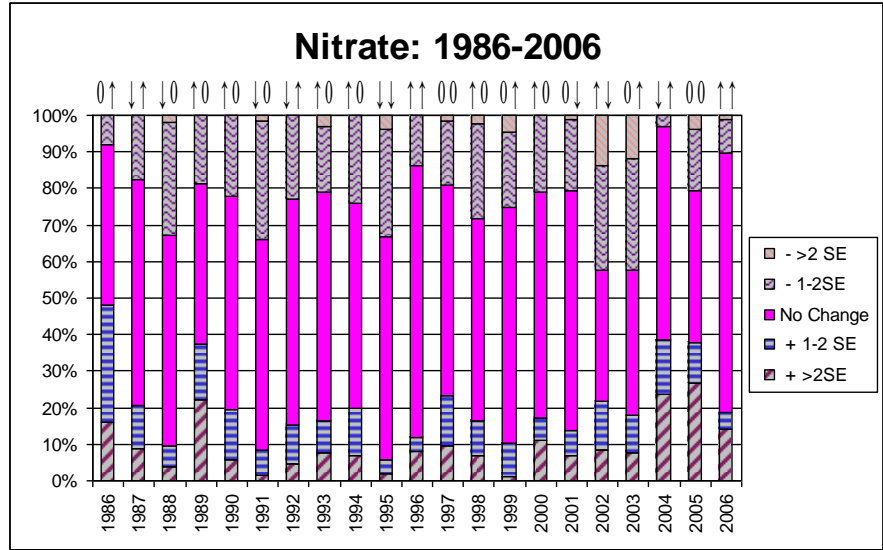


Figure 14a. Annual Change from “Normal” Nitrate in CSLAP Lakes (SE = Standard Error)

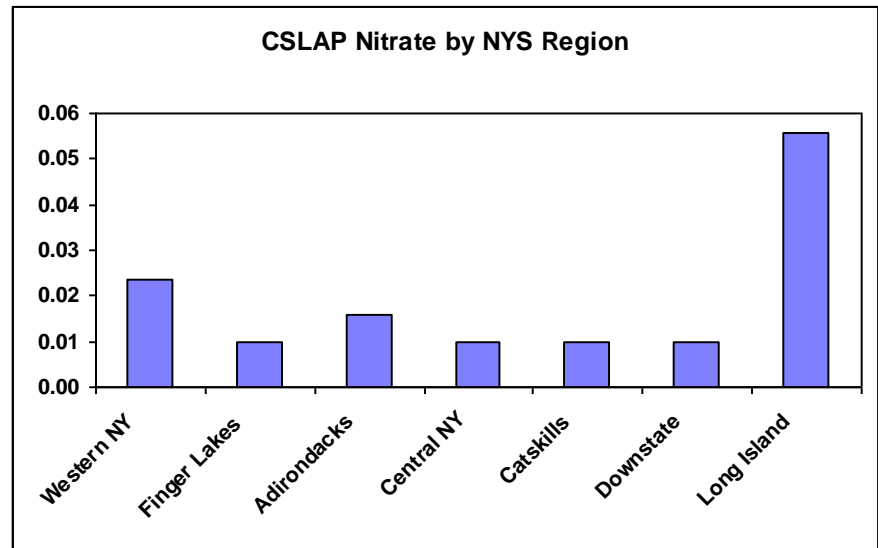


Figure 14b. Nitrate in CSLAP Lakes by NYS Region

What Was Expected in 2007?

Nitrate readings have been very unpredictable, although at nearly all times, nitrate readings have been low. Given the higher readings found in 2004 and lower readings found in 2005, nitrate levels cannot be easily predicted in 2007.

What Happened at Eagle Lake in 2007?

Nitrate readings in Eagle Lake in 2007 were higher than in any previous year, continuing a trend toward increasing nitrate concentrations in the lake since 2002. However, even the highest readings in 2007 were only slightly higher than the analytical detection limit.

Statewide Variability:

Nitrate levels are highest in Long Island, western NY, and the Adirondacks, and lowest in the other NYS regions. However, none of these regions demonstrate readings that are particularly high. Readings from individual lakes in Long Island, Madison County, and the Adirondacks (spring only) are often elevated, although still well below water-quality standards.

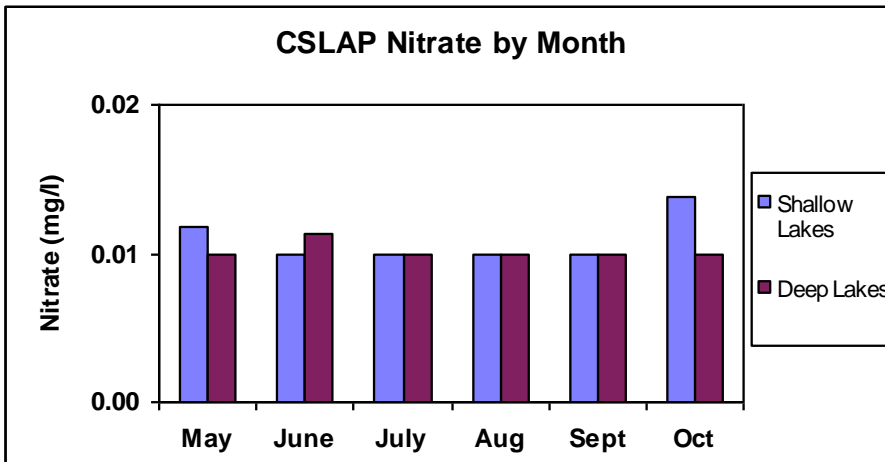


Figure 14c. Nitrate in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

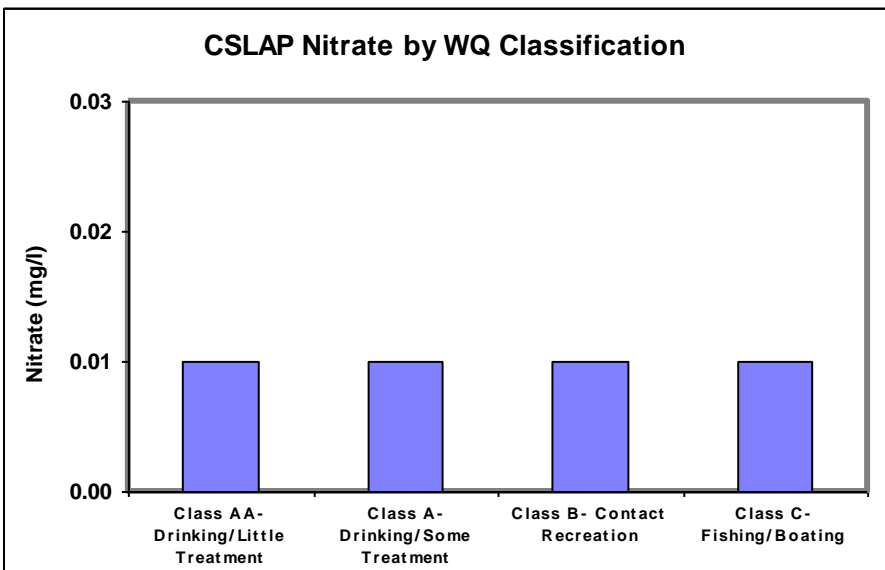


Figure 14d. Nitrate in CSLAP Lakes by Lake Use

by the melting of large winter snowpacks, such as some Class AA and A lakes in the Adirondacks, but these statistics cannot be easily teased from datasets strongly influenced by the large number of lakes with undetectable nitrate readings.

Seasonal Variability:

Nitrate readings are not seasonally variable on a program-wide basis, as indicated in Figure 14c. However, in some individual lakes, in the regions listed above, nitrate is often detectable until early summer and then undetectable through the rest of the sampling season (the large number of lakes with undetectable nitrate levels throughout the year overwhelm the statistics in Figure 14c). Nitrate levels in shallow lakes were slightly higher in October, but the difference between September and October nitrate readings is probably within the rounding error for these analyses.

Lake-Use Variability:

Nitrate readings appeared to be identical for all classes of lake uses, as indicated in Figure 14d. Higher early-season nitrate readings are found in some lakes influenced

Ammonia

Annual Variability:

Ammonia was analyzed for the first time in 2002, so long-term analyses are limited by the relative lack of data. The limited data indicated that ammonia was highest in 2002 and 2006, and lowest in 2005. 2006 was a wet year, and 2005 was dry, and while 2002 was not a wet year, these data suggest that ammonia increases with precipitation and decreases in dry conditions. It is more likely that the higher ammonia readings were associated with wet winter and spring conditions, as were apparent in both 2002 and 2006. No surface readings have approached the state water-quality standard (= 2 mg/l) in any CSLAP sample, although this threshold has been reached in some anoxic (oxygen-depleted) deepwater samples.

What Was Expected in 2007?

As noted above, ammonia readings were higher when winter and spring runoff was heaviest, and lowest when spring and summer precipitation was lower. Given the higher than normal winter and spring precipitation levels in 2007, ammonia readings were expected to increase in 2007.

What Happened at Eagle Lake in 2007?

Ammonia readings in Eagle Lake in 2007 were higher than in any previous CSLAP sampling seasons, although these readings continued to be low and not a part of any significant long-term trend.

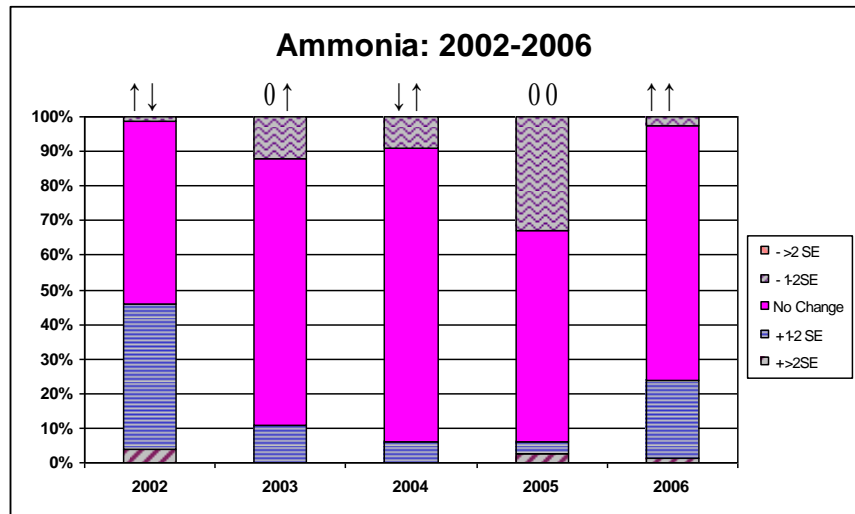


Figure 15a. Annual Change from "Normal" Ammonia in CSLAP Lakes (SE = Standard Error)

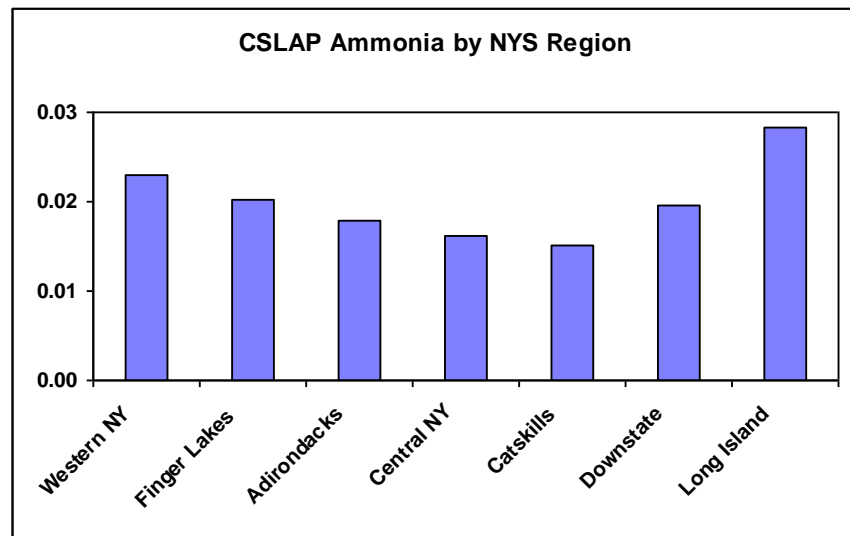


Figure 15b. Ammonia in CSLAP Lakes by NYS Region

Statewide Variability:

Ammonia levels are highest in Long Island, western NY, and the Finger Lakes, and lowest in Central New York and the Catskills. However, none of these regions demonstrate readings that are particularly high.

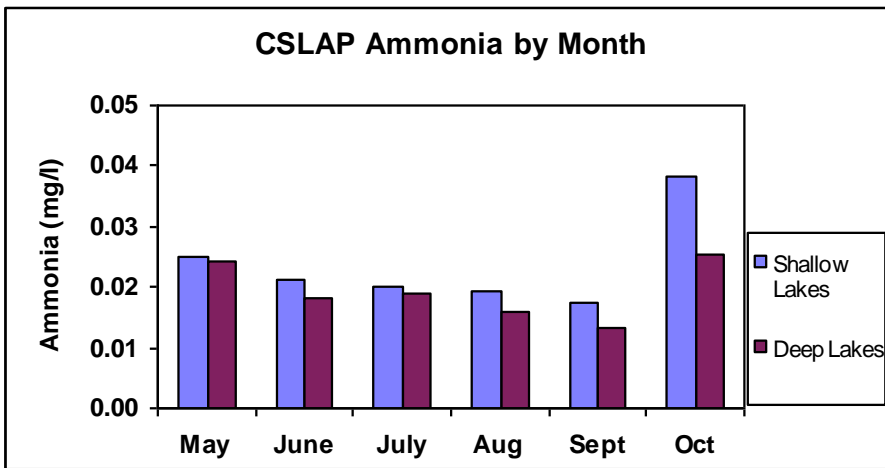


Figure 15c. Ammonia in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

Seasonal Variability:

Ammonia readings appear to decrease during the summer, and then increase in the fall, as indicated in Figure 15c. For the deeper lakes, this may be due to the migration of deepwater ammonia levels (which may have risen in response to deepwater anoxia) to the surface after the lake has been destratified. However, the rise in ammonia levels was greater for shallow lakes, suggesting other factors may also be in play.

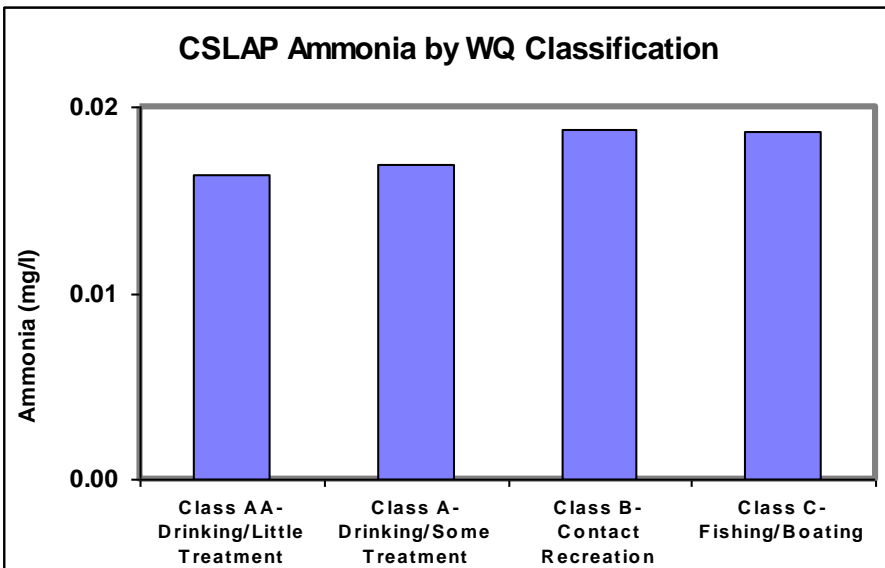


Figure 15d. Ammonia in CSLAP Lakes by Lake Use

Lake-Use Variability:

Ammonia readings appeared to be identical for all classes of lake uses, as indicated in Figure 15d. In nearly all classes of lakes, ammonia levels are close to the analytical detection limit, and far below the state water quality standard (= 2.0 mg/l).

Total (Dissolved) Nitrogen

Annual Variability:

Total dissolved nitrogen (TDN) was analyzed for the first time in 2002, so long-term analyses are limited by the relative lack of data. The limited data indicated that TDN was highest in 2006, when the winter/spring and summer precipitation levels were higher than normal. TDN data were lowest in 2005, which was perhaps drier than any other CSLAP sampling season since 2002, at least on a statewide basis. These patterns generally follow the trends observed with the ammonia data, but were inconsistent with the nitrate data.

What Was Expected in 2007?

Given the apparent connection between ammonia and precipitation noted in Figure 16a (readings highest in wet weather and lowest in dry weather), ammonia readings could be expected to be higher than normal in 2007, since precipitation levels were higher in most parts of the state.

What Happened at Eagle Lake in 2007?

TDN readings in Eagle Lake were higher in 2007 and in 2006 than in previous CSLAP sampling seasons, but no long-term trends are apparent.

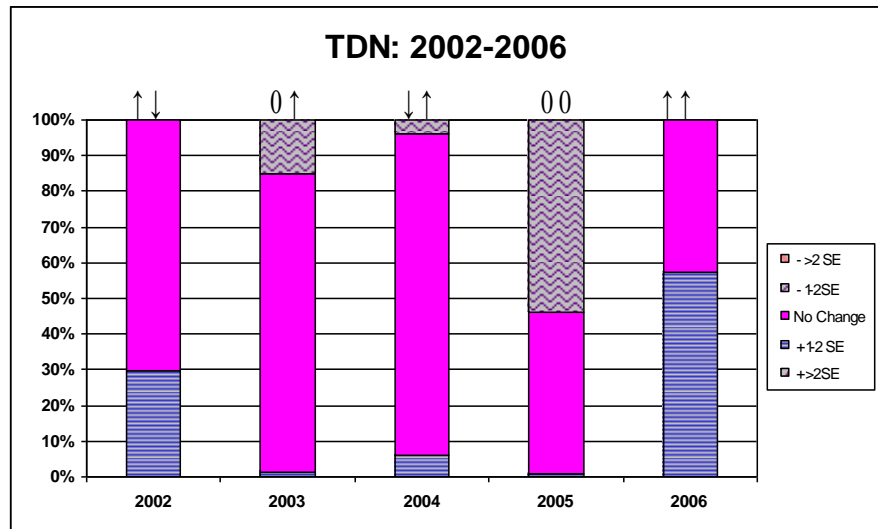


Figure 16a. Annual Change from "Normal" TDN in CSLAP Lakes (SE = Standard Error)

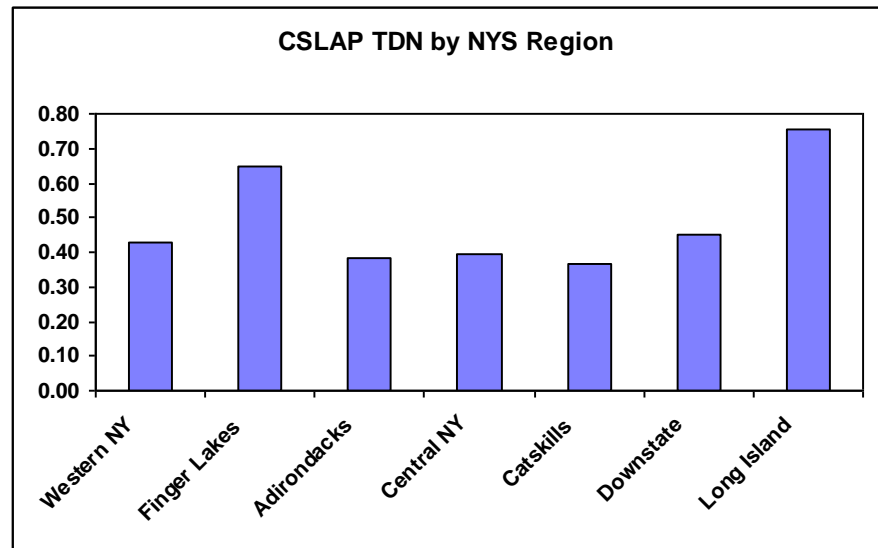


Figure 16b. TDN in CSLAP Lakes by NYS Region

Statewide Variability:

Total dissolved nitrogen levels are highest in Long Island and the Finger Lakes, and consistently lower everywhere else. The higher readings from both regions are probably associated with dissolved organic nitrogen, since nitrate and ammonia readings are much lower than total nitrogen. This does not appear to have translated into higher algae levels in these regions (see the discussion below re: chlorophyll *a*).

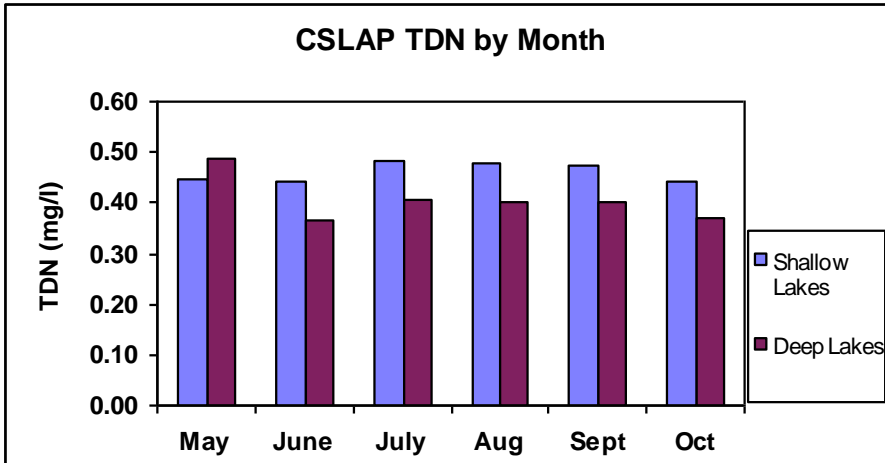


Figure 16c. TDN in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

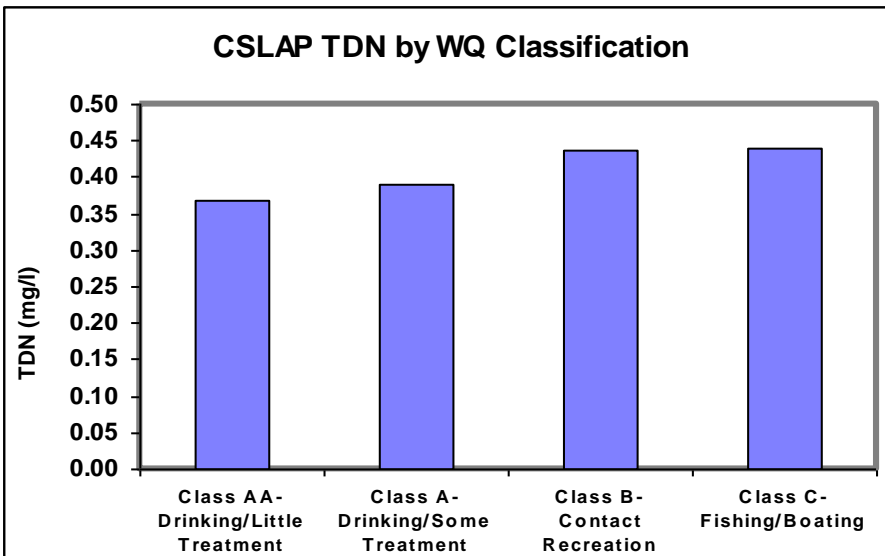


Figure 16d. TDN in CSLAP Lakes by Lake Use

Seasonal Variability:

Total dissolved nitrogen readings are not seasonally variable, particularly in shallow lakes, as indicated in Figure 16c. TDN readings in deeper lakes were higher in May than in any subsequent sampling month, although this is probably due to more May sampling of deep lakes with “normally” high dissolved nitrogen readings rather than higher early season readings in all deep lakes. Shallow lake TDN readings were fairly stable throughout the summer, and at nearly all times were higher than deep lake TDN levels.

Lake-Use Variability:

Total dissolved nitrogen readings were higher in Class B and Class C lakes than in Class AA or Class A lakes, as can be seen in Figure 16d. This “finding” cannot be easily explained, but additional data in the coming years may help to determine if the pattern shown

in Figure 16d represents a real phenomenon or one influenced by relatively small datasets.

**Trophic Indicators:
Water Clarity**

Annual Variability:

Water clarity (transparency) has varied annually in most CSLAP lakes. There does not appear to be much of a correlation between clarity and precipitation—the highest clarity occurred in 1995, 1997, and 1999, which corresponded to normal precipitation (statewide), although the lowest clarity occurred during three wet years (1996, 2000, and 2006). There are no significant broad statewide water clarity trends, although (as described in other portions of this report), clear trends do exist on some lakes. The majority of water clarity readings in CSLAP lakes (59%) correspond to *mesotrophic* conditions (clarity between 2 and 5 meters), with 26% corresponding to *eutrophic* conditions ($Z_{sd} < 2$) and 15% corresponding to *oligotrophic* conditions ($Z_{sd} > 5$).

What Was Expected in 2007?

While water transparency readings do not appear to be strongly affected by dry weather, water clarity seems to be lowest during wet years. Since 2007 was a wet year in much of the state, it is likely that more lakes would exhibit slightly lower water transparency readings in 2007.

What Happened at Eagle Lake in 2007?

Water transparency readings in 2007 were among the lowest recorded through CSLAP at Eagle Lake, following a trend toward decreasing water clarity since 2000. This has occurred despite algae (chlorophyll *a*) readings that have been lower and phosphorus readings that have been fairly stable over the same period, although this was also coincident with an increase in water color readings.

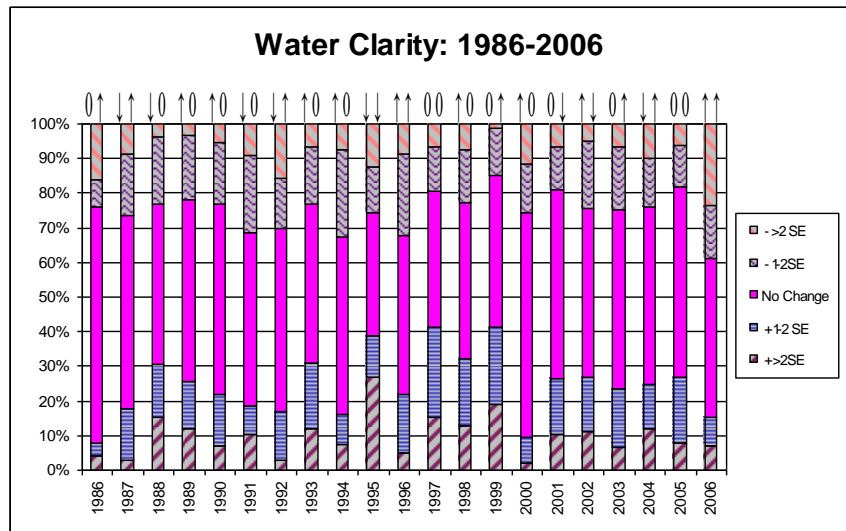


Figure 17a. Change from “Normal” Water Clarity in CSLAP Lakes (SE = Standard Error)

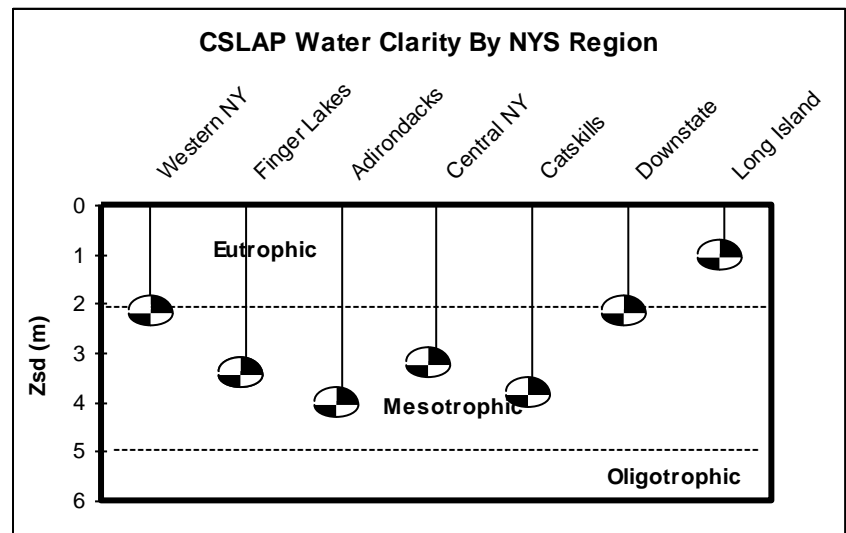


Figure 17b. Water Clarity in CSLAP Lakes by NYS Region

Statewide Variability:

As expected, water clarity is highest in the Adirondacks, Catskills, and Finger Lakes regions, and lowest in Long Island, downstate, and western NY. The differences are more pronounced (at least for the Adirondacks) when “naturally” colored lakes are not considered. However, except for Long Island (for which water clarity is at least partially limited by the shallow water depth), the “typical” lake in each of these regions would be classified as *mesotrophic*.

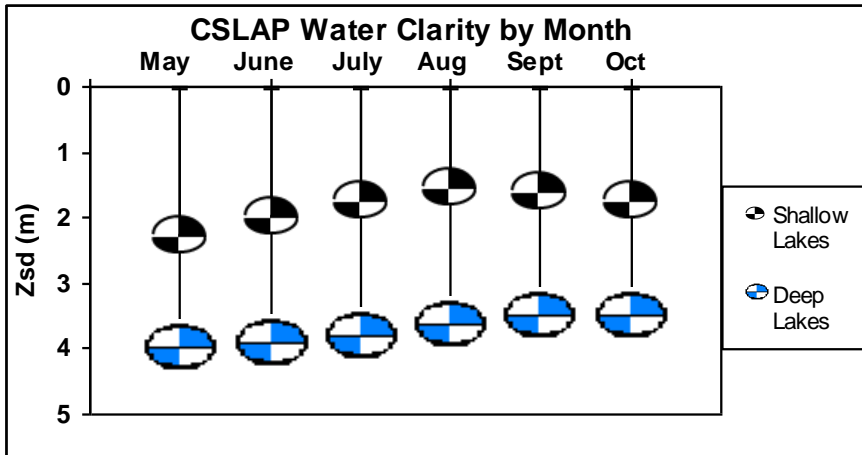


Figure 17c. Water Clarity in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

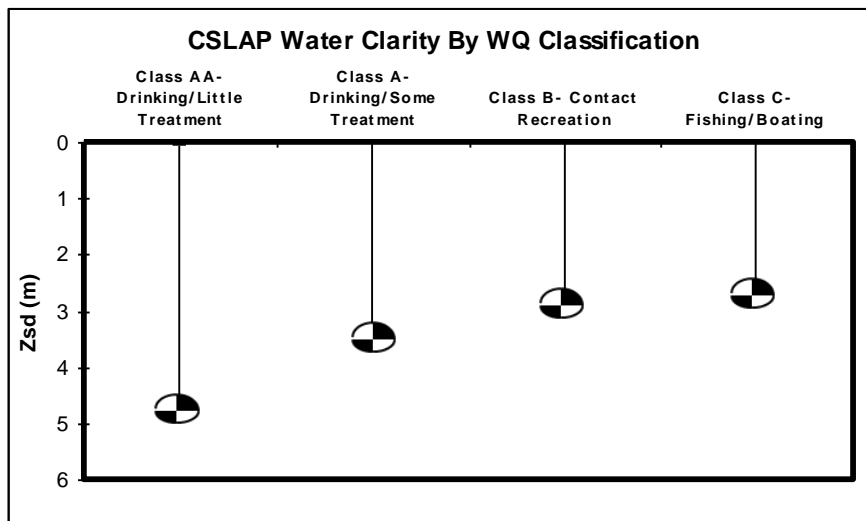


Figure 17d. Water Clarity in CSLAP Lakes by Lake Use

used for potable water (Class AA), and lower clarity found in lakes used primarily for contact and non-contact (fishing and boating) recreation. As with many of the other water-quality indicators, this is due to both geographical and morphometric (depth) differences, although the original designation of these uses may also reflect these measurable and visually apparent water-quality differences.

Seasonal Variability:

Water clarity readings are lower, as expected, in shallow lakes, even when water depth does not physically limit a water clarity measurement. Transparency decreases in both shallow and deep lakes during the course of the sampling season (the drop in clarity in shallower lakes is somewhat more significant), although clarity readings increase from spring to early summer in deeper CSLAP lakes. Water transparency rebounds slightly in shallower lakes in the fall, probably due to a drop in nutrient levels. The lack of “rebound” in deeper lakes may be due to occasional fall algal blooms in response to surface nutrient enrichment after lake turnover (see below).

Lake-Use Variability:

Water transparency decreases as the “sensitivity” of the lake use decreases, with higher clarity found in lakes

**Trophic Indicators:
Phosphorus (TP)**

Annual Variability:

Total phosphorus (TP) has varied annually in most CSLAP lakes. The highest phosphorus readings occurred during 1991, 1996, 1998, 2000, and 2003, the latter four of which corresponded to wet years. However, of the years with the lowest readings, only 1995 (and not 1989, 1997, and 2002) corresponded to dry years, and 2004 was a fairly wet year. The majority of phosphorus readings in CSLAP lakes (40%) correspond to *mesotrophic* conditions (clarity of 2 to 5m), with 30% corresponding to *eutrophic* conditions (< 2m clarity) and 30% corresponding to *oligotrophic* conditions (> 5m clarity); the latter is a much higher percentage than the trophic designation for water clarity.

What Was Expected in 2007?

As noted above, there is not a strong correlation between weather and total phosphorus, and there does not appear to be a consistent long-term pattern in the total phosphorus data.

The data also does not appear to be significantly laboratory-dependent, at least as apparent in Figure 15a. As such, it is difficult to predict whether phosphorus levels might be expected to be higher or lower in most CSLAP lakes in 2007.

What Happened at Eagle Lake in 2007?

Phosphorus readings in Eagle Lake in 2007 were close to the long-term average for the lake, and the typical reading in most CSLAP sampling seasons has been consistently low (despite some significant variability within each sampling season).

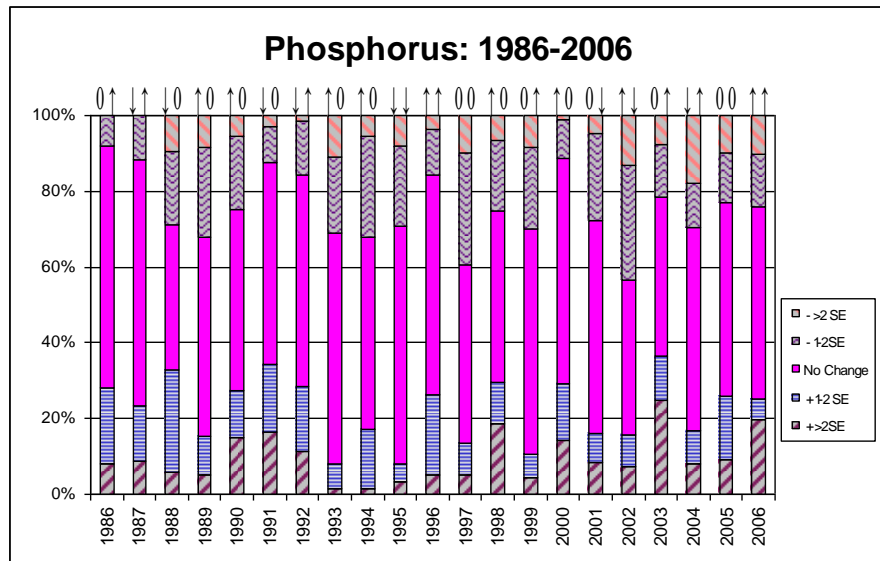


Figure 18a. Annual Change from “Normal” TP in CSLAP Lakes (SE = Standard Error)

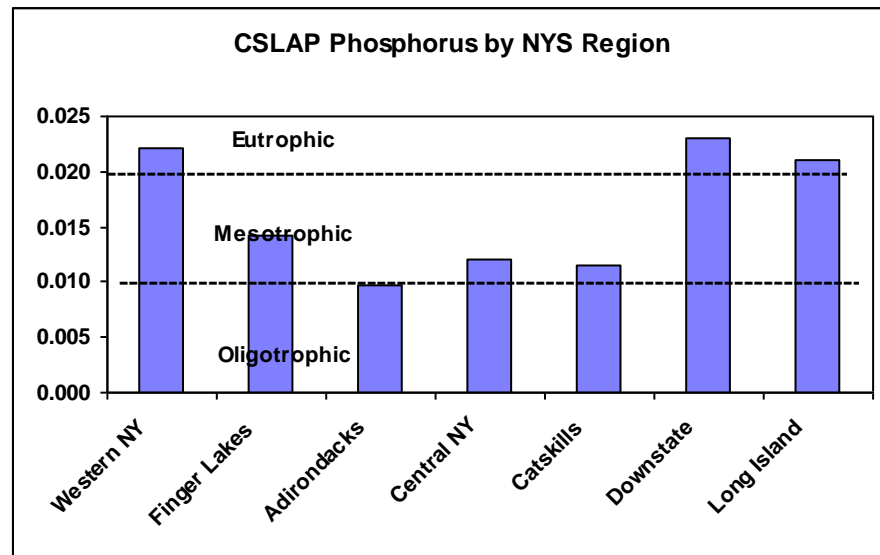


Figure 18b. TP in CSLAP Lakes by NYS Region

Statewide Variability:

As expected, nutrient levels are lowest in the Adirondacks, Catskills, and Central New York (where clarity is highest). Nutrient concentrations were highest in Long Island, downstate, and western NY, where water transparency is lowest. In the latter three regions, the “typical” lake in each of these regions would be classified as *eutrophic*, while only in the Adirondacks could most lakes be described as *oligotrophic*, based on nutrients.

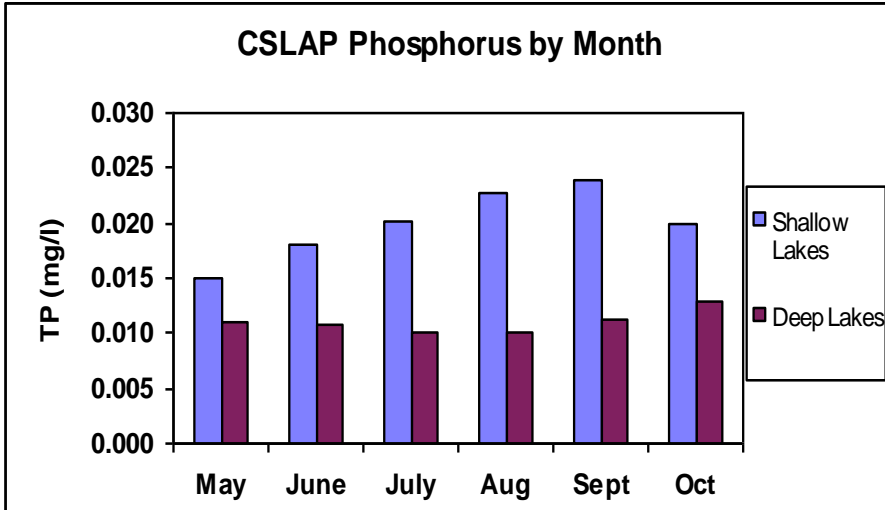


Figure 18c. TP in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

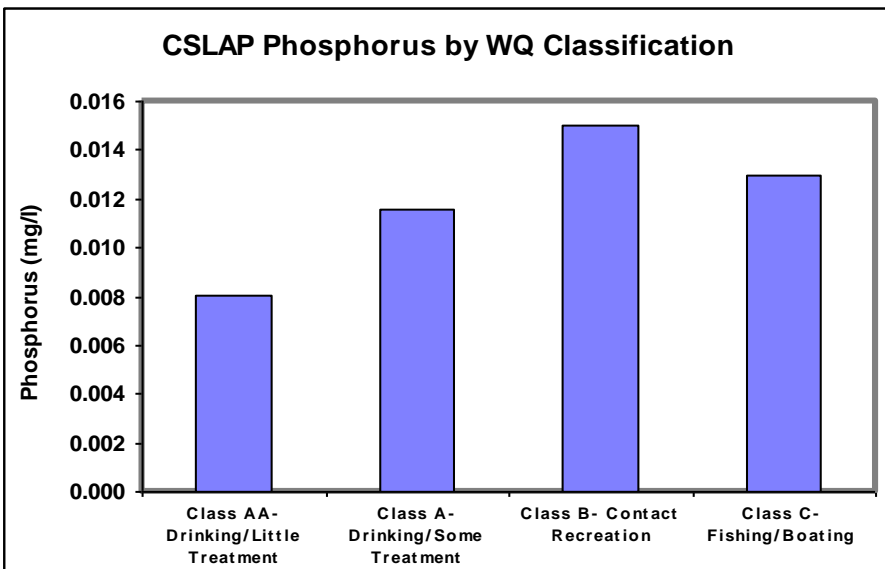


Figure 18d. TP in CSLAP Lakes by Lake Use

recreation), these lakes actually have higher nutrient levels, perhaps reflecting the influence of deepwater nutrient enrichments (these lakes are typically deeper) and the “unofficial” use of Class C waters for bathing and contact recreation.

Seasonal Variability:

Nutrient levels are higher, as expected, in shallow lakes, and phosphorus levels increase in shallow lakes during the course of the sampling season, until dropping in the fall. However, phosphorus levels in deeper lakes are lower and decrease slightly through July, then increase into the fall. The latter phenomenon is due to surface nutrient enrichment after lake turnover (high nutrient water from the lake bottom, due to release of nutrients from poorly oxygenated lake sediments in the summer, migrates to the lake surface when the lake destratifies).

Lake-Use Variability

Phosphorus readings are lower in lakes used for minimally treated potable water intakes (Class AA) and are higher for other lake uses. Although Class B waters are utilized for a “higher” lake use than Class C lakes (contact recreation versus non-contact

Trophic Indicators: Chlorophyll *a* (Chl.*a*)

Annual Variability:

Chlorophyll *a* (chl.*a*) has varied in most CSLAP lakes more significantly than the other trophic indicators, as is typical of biological indicators, which tend to grow “patchy”. With the exception of the very high readings in 1987 (probably due to a lab problem), the highest chlorophyll *a* levels occurred during 1990, 1991, 1994, and 1996, with all but 1991 corresponding to higher spring runoff. However, the lowest readings were in 1986, 2002, and 2005; only the latter corresponded to a particularly dry year. The consistently lower chlorophyll readings in the last five years may also correspond to the shift in laboratories, although both labs use the same analytical methodology and chlorophyll readings were also low in the last few years before changing laboratories. The near majority of chlorophyll readings in CSLAP lakes (53%) correspond to *mesotrophic* conditions

(chlorophyll *a* readings between 2 and 8 µg/l), with 37% corresponding to *eutrophic* conditions (chl.*a* > 8 µg/l) and 10% corresponding to *oligotrophic* conditions (chl.*a* < 2 µg/l); these percentages are more like those for water clarity rather than those for phosphorus.

What Was Expected in 2007?

Chlorophyll *a* levels cannot be well predicted in dry years, as observed in Figure 19a. However, chlorophyll *a* readings are occasionally higher than normal during wet years. Since at least the winter and spring of 2007 was wet in most of the state, algae levels in New York state could be expected to be higher than normal.

What Happened at Eagle Lake in 2007?

Chlorophyll *a* readings have been consistently low and typical of highly unproductive lakes. Readings in 2007 were probably comparable to those measured in previous years.

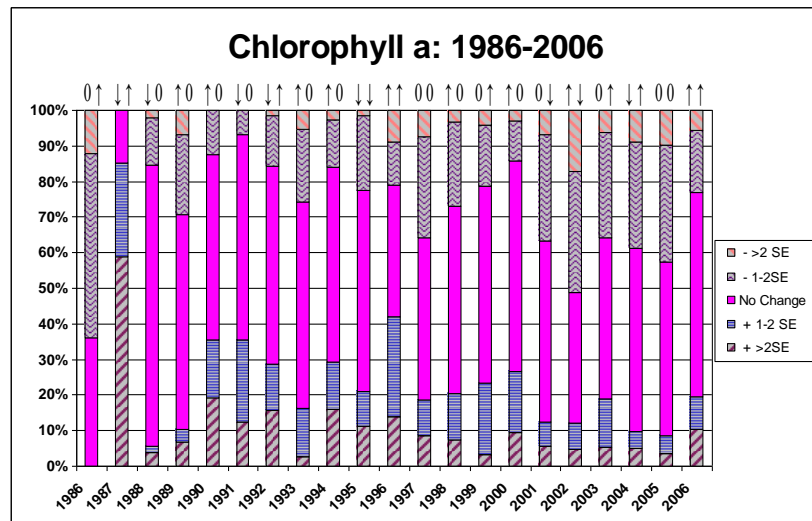


Figure 19a. Annual Change from “Normal” Chlorophyll *a* in CSLAP Lakes (SE = Standard Error)

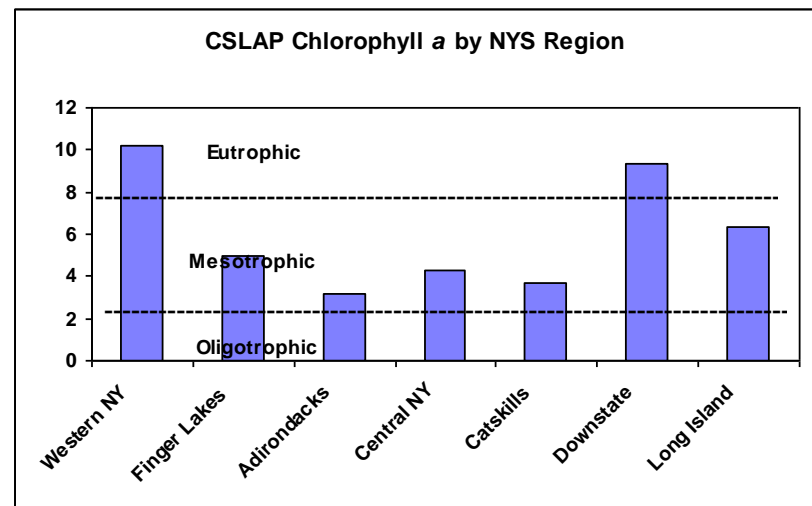


Figure 19b. Chlorophyll *a* in CSLAP Lakes by NYS Region

Statewide Variability:

As with phosphorus, chlorophyll levels are lowest in the Adirondacks, Central New York, and the Catskills (where clarity is highest) and highest in Long Island, downstate, and western NY, where water transparency is lowest. In the latter two regions, the “typical” lake in each of these regions would be classified as *eutrophic*, while lakes in the other regions would be described as *mesotrophic*, based on assessments from chlorophyll *a* readings.

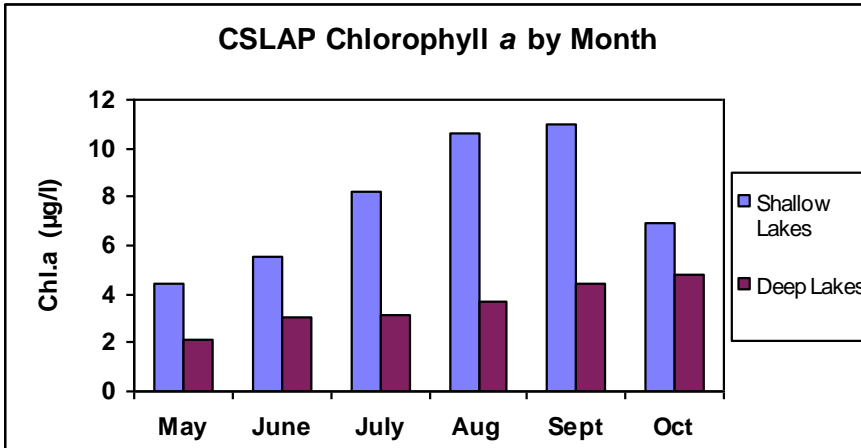


Figure 19c. Chlorophyll *a* in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

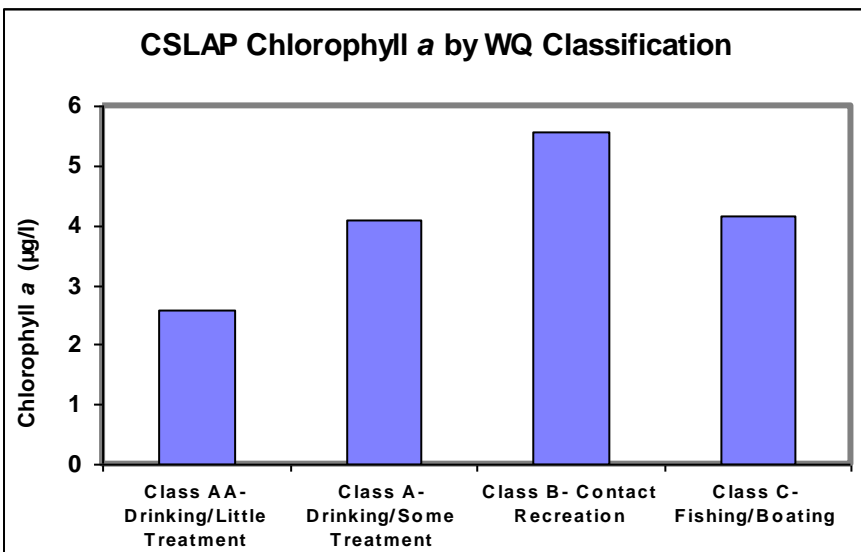


Figure 19d. Chlorophyll *a* in CSLAP Lakes by Lake Use

actually have similar levels, perhaps reflecting the influence of deepwater nutrient enrichments (these lakes are typically deeper) and the “unofficial” use of Class C waters for bathing and contact recreation. This is similar to the use pattern for phosphorus.

Seasonal Variability:

Chlorophyll levels are higher, as expected, in shallow lakes, and increase in both shallow and deep lakes during the course of the sampling season, with chlorophyll readings dropping in shallow lakes in the fall. The steady increase in chlorophyll in both shallow and (to a lesser extent) deep lakes is consistent with the change in phosphorus over the same period, due to steady migration of nutrients released from poorly oxygenated lake sediments during the summer and especially in the fall (as well as drier weather, increased lake use, and other factors).

Lake-Use Variability:

Chlorophyll readings are lower in lakes used for minimally treated potable water intakes (Class AA) and are higher for other lake uses. Although Class B waters are utilized for a “higher” lake use than Class C lakes (contact recreation versus non-contact recreation), these lakes

Water-quality Assessment (QA on the Perception Form)

Annual Variability

Water-quality assessments (the perceived physical condition of the lake or QA on the use-impairment surveys) were least favorable in wet (2000 and 2006) years, but were highly variable in very dry (1995) years, suggesting the lack of correlation between weather and perceived water-quality. These assessments were most favorable in 1992, 1997, and 1999. There is a strong connection between measured and perceived water clarity in most CSLAP lakes, and a comparison of Figures 17a and 20a shows that the most favorable water quality assessments usually occurred in the years with the highest measured water transparency. This occurs despite the lack of a strong connection between water quality assessments and precipitation patterns.

What Was Expected in 2007?

There was not a strong connection between precipitation and perceived water-quality. It is difficult to predict expected conditions in 2007, although water clarity readings were expected to be slightly lower than normal in response to wetter weather in much of the state in 2007.

What Happened at Eagle Lake in 2007?

Water-quality assessments in Eagle Lake have been very stable since 2002, and slightly less favorable than in other lakes with similar water quality conditions .

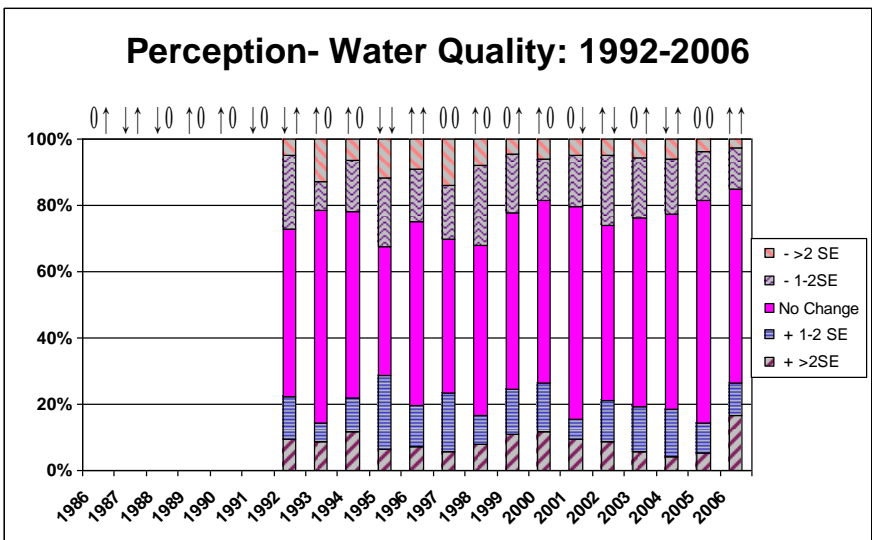


Figure 20a. Annual Change from "Normal" Water-Quality Assessment in CSLAP Lakes (SE = Standard Error)

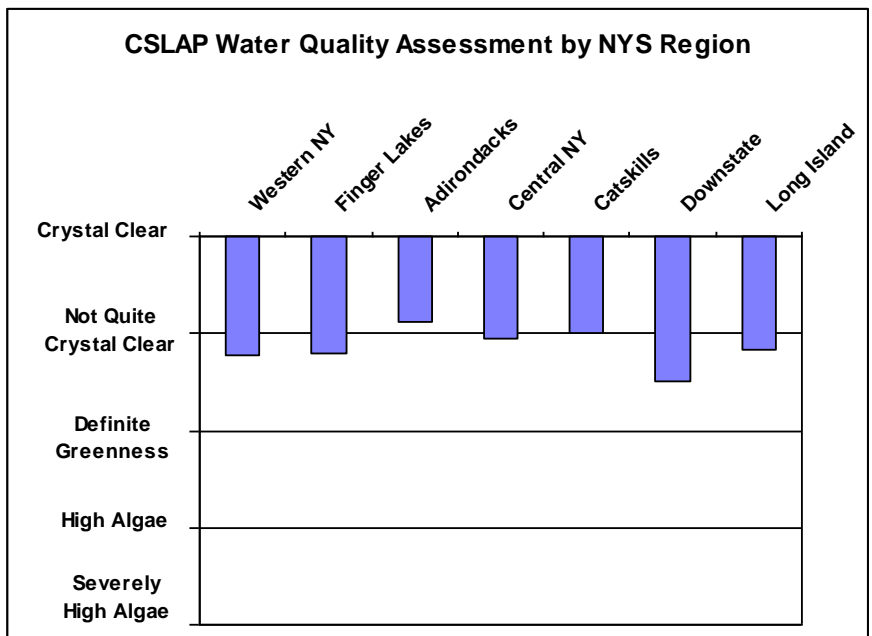


Figure 20b. Water-Quality Assessment in CSLAP Lakes by NYS Region

Statewide Variability:

The most favorable water-quality assessments (at least in support of contact recreation) occurred in the Adirondacks, Catskills, and central New York, as expected, and water-quality assessments were slightly less favorable downstate, in western NY, and on Long Island. This is mostly consistent with the water clarity readings in these regions. However, since the difference between the most favorable (Adirondacks) and least favorable (downstate) assessments is smaller than the measured water transparency differences, this suggests that the relatively low water clarity in the latter regions may often be considered “normal” by lake residents.

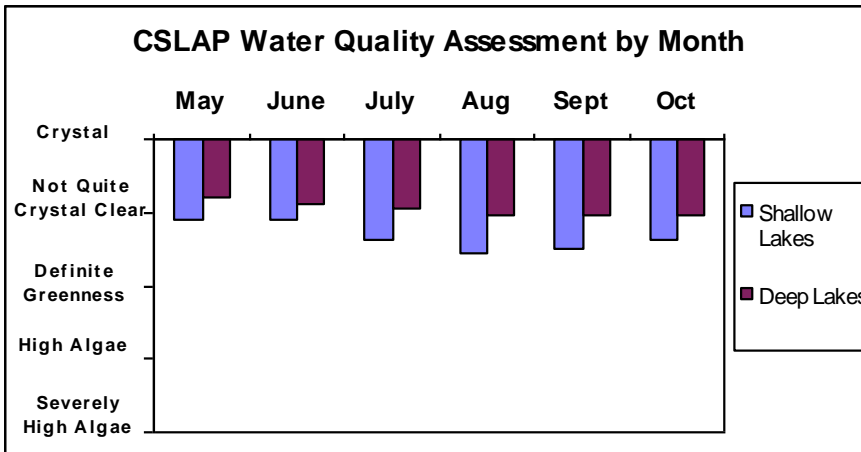


Figure 20c. Water-Quality Assessment in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

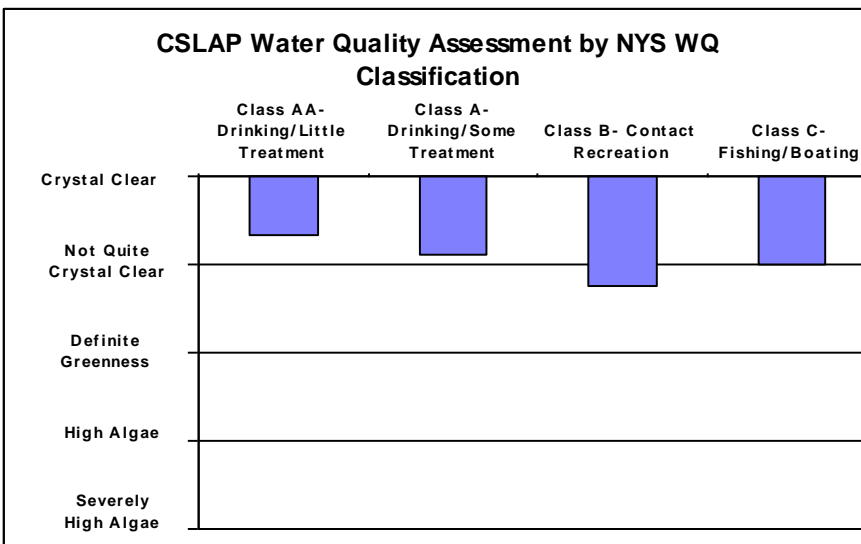


Figure 20d. Water-Quality Assessment in CSLAP Lakes by Lake Use

assessments, perhaps reflecting the influence of deepwater nutrient enrichments (these lakes are typically deeper) and the “unofficial” use of Class C waters for bathing and contact recreation. This is similar to the pattern seen for the trophic indicators.

Seasonal Variability:

Water-quality assessments become less favorable as the summer progresses in both deep and (especially) shallow lakes, coincident with similar patterns for the trophic indicators. However, the seasonal changes in these assessments are not very large. These assessments become slightly more favorable in shallow lakes in the fall, consistent with the improved (measured) water clarity, although overall water-quality assessments are less favorable all year in shallow lakes.

Lake Use Variability:

Water-quality assessments are more favorable in lakes used for potable water intakes (Class AA and Class A) and less favorable for other lake uses. Although Class B waters are utilized for a “higher” lake use than Class C lakes (contact recreation versus non-contact recreation), these lakes actually have similar water-quality

Aquatic Plant (Weed) Assessment (QB)

Annual Variability:

Aquatic-plant assessments (the perceived extent of weed growth in the lake or QB on the use impairment surveys) indicated that weed coverage was greatest in 1992, 1998, 2000, 2002, and 2005 with only 1998 associated with wet weather. Weed growth was less extensive in 1997 and 2003, neither of which exhibited significant changes in precipitation, suggesting the lack of correlation between weather and weed densities. The highest weed growth occurred when the perceived physical condition (clarity) of the lake was also least favorable, such as in 1995 and 2000. These conditions may offer a selective advantage to invasive or exotic weeds (such as *Myriophyllum spicatum*) which can create surface canopies. Despite continuing concerns about increased invasion from exotic weeds, Figure 21a suggests that no long-term trend toward greater aquatic plant coverage is apparent.

What Was Expected in 2007?

There was not a strong connection between precipitation and extent of weed growth, at least as measurable through CSLAP. This makes it difficult to identify expected conditions in 2007. However, aquatic plant densities are often greater when water clarity is lowest (particularly in lakes with exotic weeds), so lower water transparency in 2007 may trigger an increase in weed densities.

And What Happened at Eagle Lake in 2007?

Aquatic plant coverage in all CSLAP sampling seasons have been stable and indicated surface growth of aquatic plants throughout the summer. Assessments in 2007 were comparable to those in previous years, although it is not known if these assessments are representative of overall conditions in the lake.

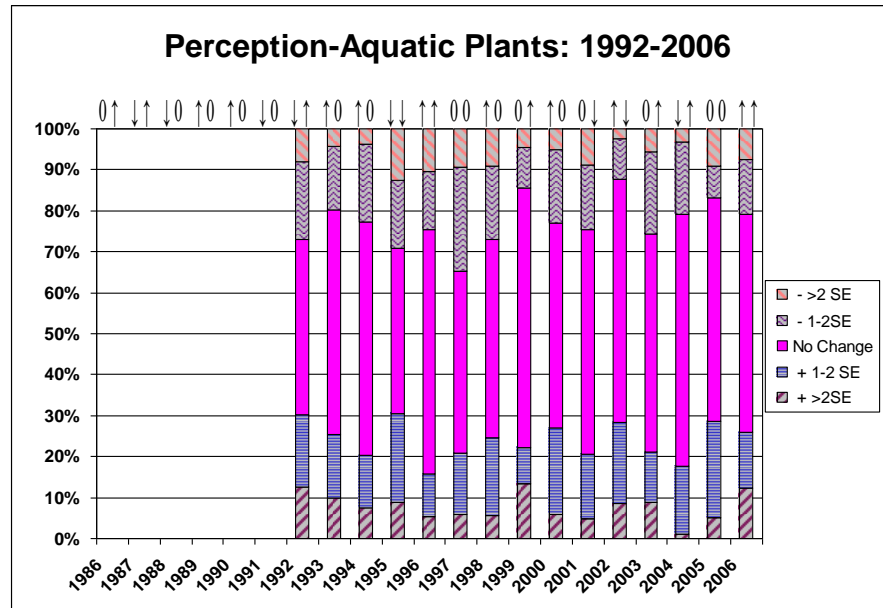


Figure 21a. Annual Change from "Normal" Weed Assessment in CSLAP Lakes (SE = Standard Error)

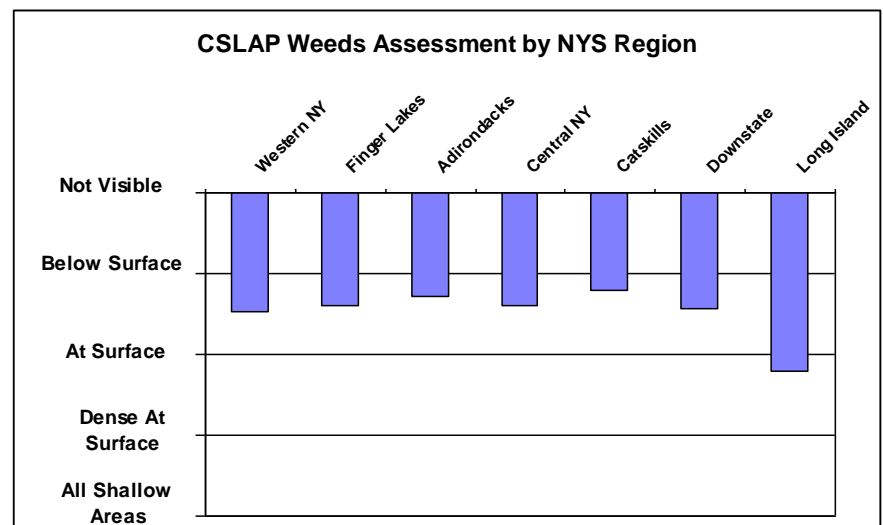


Figure 21b. Weed Assessment in CSLAP Lakes by NYS Region

Statewide Variability:

Aquatic plant growth was most significant in Long Island (and to a lesser extent downstate and in western NY) and least significant in the Catskills and Adirondacks. The former may have a larger concentration of shallow lakes (Long Island) or preponderance of exotic weeds (downstate and western NY), while the latter may correspond to deeper lakes or fewer instances of these invasive weeds, although it is also likely that invasive-weed growth may be increasing in many lakes within these “less impacted” areas.

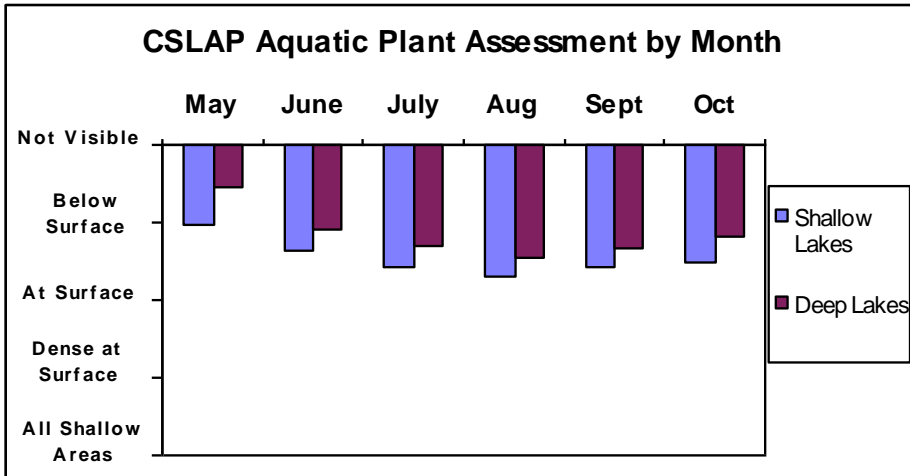


Figure 21c. Weed Assessment in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

Seasonal Variability:

As expected, aquatic-plant densities and coverage increase seasonally (through late summer) in both shallow and deep lakes, with greater aquatic-plant coverage and densities found in shallow lakes. Peak aquatic-plant densities tend to occur in late summer in both deep and shallow lakes. The variability from one lake to another (from very little growth to dense growth at the lake surface) is more pronounced later in the summer. Despite higher clarity in shallow lakes in the fall, aquatic-plant coverage decreases, while the drop in fall plant coverage in deeper lakes is less pronounced.

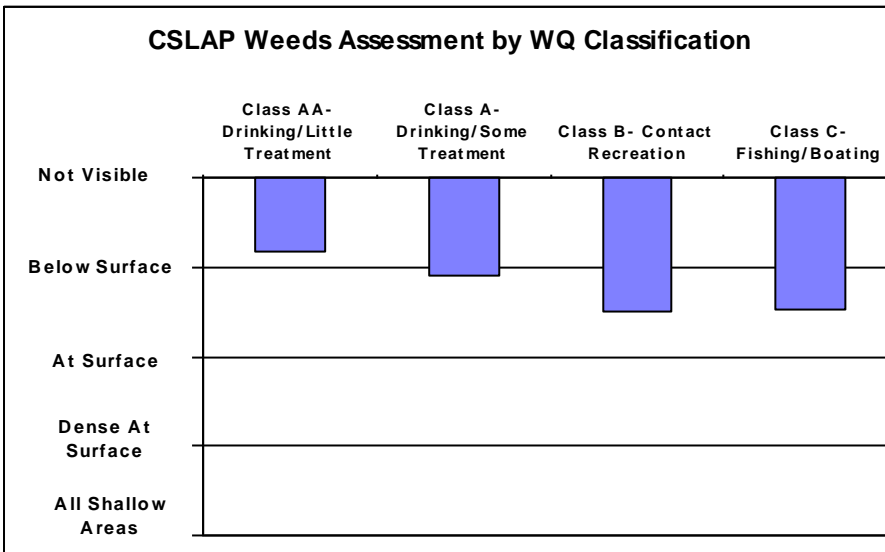


Figure 21d. Weed Assessment in CSLAP Lakes by Lake Use

lake size and depth (Class B lakes tend to be found outside the high elevation areas in the Catskills and Adirondacks, and Class C lakes tend to be shallower than Class AA or Class A lakes).

Lake Use Variability:

Aquatic-plant coverage was more significant in Class B and Class C lakes than in other lakes, but this (again) is probably a greater reflection of geography or

Recreational Assessment (QC)

Annual Variability:

Recreational assessments (the perceived recreational suitability of the lake or QC on the use-impairment surveys) have varied from year to year, with no clear long-term pattern. The most favorable assessments were in 1995, 1997, and 1998. 1997 corresponded to a year with low aquatic-plant (weed) coverage and favorable water quality. This suggests that recreational assessments are influenced by both water-quality conditions and aquatic plant densities. Less favorable assessments occurred in 1992, 2000, and 2006. Extensive weed growth was reported in 1992 and 2000, and poor water quality was more common in 2000 and 2006. The extent of “normal” conditions (the middle bar in Figure 22a) has generally not changed significantly since perception surveys were first conducted in 1992.

What Was Expected in 2007?

There is not a strong connection between precipitation and perceived recreational conditions. While it is reasonable to assume that recreational assessments will be less favorable if either water quality perceptions are unfavorable or aquatic plant coverage increases, changes in water quality or plant coverage is difficult to predict. As noted above, given the 2007 weather patterns and their expected impact on water transparency and weeds, it is more likely that recreational assessments will be less favorable than more favorable.

What Happened at Eagle Lake in 2007?

Recreational assessments have been very stable in all CSLAP sampling seasons, including 2007.

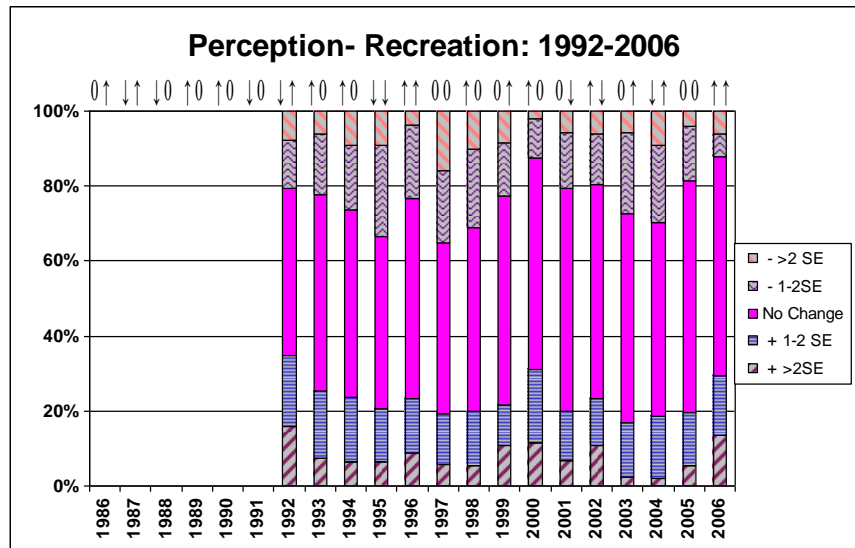


Figure 22a. Annual Change from “Normal” Recreational Assessment in CSLAP Lakes (SE = Standard Error)

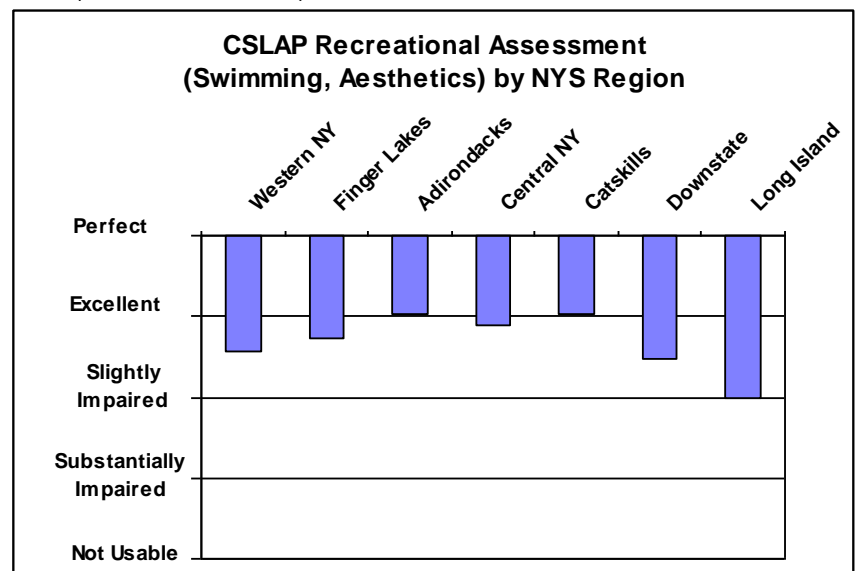


Figure 22b. Recreational Assessment in CSLAP Lakes by NYS Region

Statewide Variability:

Recreational assessments are most favorable in the Adirondacks and Catskills, and less favorable in Long Island and (to a lesser extent) downstate and in western New York. This appears to be in response to less favorable assessments of water-quality and aquatic plant growth, respectively. Except for (the assessments in the small number of CSLAP lakes in) Long Island, overall recreational assessments in all regions are, in general, highly favorable.

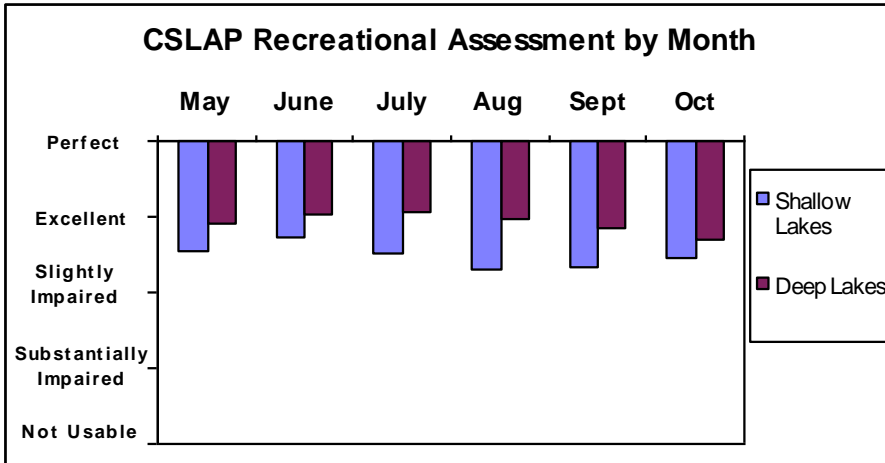


Figure 22c. Recreational Assessment in Shallow (<20ft deep) and Deep CSLAP Lakes by Month

Seasonal Variability:

Recreational assessment in both shallow and deep lakes tends to improve from spring to early summer and then degrade through the summer, improving in shallow lakes in the fall. As expected, this generally corresponds to seasonal increases in aquatic plant coverage in deep lakes and also to seasonally degrading water-quality in shallow lakes. Overall recreational assessments are more favorable in deep lakes every month of the sampling season, although the differences are less pronounced in late spring and early fall (and winter, when every lake looks nice!).

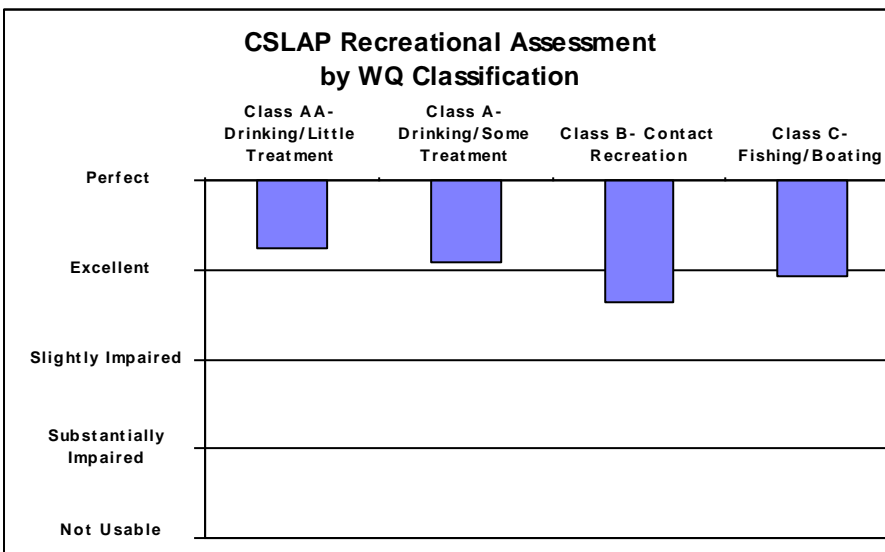


Figure 22d. Recreational Assessment in CSLAP Lakes by Lake Use

Lake Use Variability:

Recreational assessment becomes less favorable as the designated lake use becomes less sensitive (drinking water to contact recreation), although recreational assessments of Class B and C lakes are only slightly less favorable than in

Class AA and A lakes. This may be considered a validation of these classifications (recognizing, again, that many Class C lakes continue to fully support contact recreation and perhaps even potable-water use).

So What Have We Learned Through CSLAP?

After more than twenty years and more than 15,000 samples collected from more than 220 lakes throughout New York State, we have learned a lot about the lakes of New York State as a direct result of the work of nearly 1,500 volunteers through CSLAP. Some of these findings have been summarized in other places in this report, but these and other findings can be distilled here:

- Water quality conditions in most CSLAP lakes have not changed significantly in the last twenty years. While there have been some water quality trends, as discussed below, the majority of the changes observed in these lakes appear to be within the normal range of variability expected in most lakes. This is not to discount the important work done by many NYSFOLA lake associations—improvements in septic management, reductions in lawn fertilization, erosion and stormwater management, and invasive species prevention may have minimized or at least slowed down the steady progression toward lake succession and the continued onslaught of overdevelopment and global climate change. Unfortunately, it is not yet known if these findings can be extrapolated to the entirety of New York State lakes, even though the typical CSLAP lake is similar to the typical New York State lake (in the “developed” portions of the state).
- For those lakes that exhibited significant change, there was no clear pattern of change for most water quality indicators measured through CSLAP. However, there were some exceptions:
 - The most common change occurred with conductivity, which increased significantly in about 20% of CSLAP lakes (versus decreasing conductivity in about 10% of these lakes).
 - Water color increased in 15-20% of these lakes, with the majority of the increase occurring in the last five years. While this corresponded to a warmer and wetter period for many NYS lakes, this also corresponded to the change in laboratories.
 - pH has decreased in twice as many lakes than it has increased, although this decrease occurred in only about 10% of the CSLAP lakes.
 - Water temperature readings have increased in more than 10% of the CSLAP lakes. While a similar change was not apparent with the air temperature data, the latter reflects an instantaneous measurement that might not reflect larger scale changes.
- The frequency of phosphorus readings exceeding 20 parts per billion (or $\mu\text{g/l}$) is very similar to the frequency of water clarity readings below 2 meters. Since the former corresponds to the state guidance value for Class B (swimming) lakes, this suggests that water clarity readings may be a useful surrogate for evaluating potential impacts of excessive algae to swimming and contact recreation
- For many CSLAP lakes, there appears to be a strong correlation between water transparency and precipitation—lower water clarity readings occur in response to heavy rainfall and/or runoff. While phosphorus readings and algae levels also increase as a result of higher precipitation, the correlation is not as strong, probably due to increasing turbidity and lower transmission of light into the water, less sunlight, and the impact of water color on water transparency
- There is a strong correlation between water quality perception and standard eutrophication indicators—water clarity, chlorophyll *a*, and total phosphorus. This has significant implications for developing water quality standards or criteria for these water quality indicators, since poor water quality perception is closely connected to recreational and aesthetic impacts and provides an impetus for managing these resources.

VI. DETAILED EAGLE LAKE WATER-QUALITY SUMMARY

CSLAP is intended to provide a database to help lake associations understand lake conditions and foster sound lake protection and pollution prevention decisions. This individual lake summary for 2006 contains two forms of information. The raw data and graphs present a snapshot or glimpse of water-quality conditions at each lake. They are based on (at most) eight or nine sampling events during the summer. As lakes are sampled through CSLAP for a number of years, the database for each lake will expand, and assessments of lake conditions and water-quality data become more accurate. For this reason, lakes new to CSLAP for only one year will not have information about annual trends.

Raw Data

Two “data sets” are provided below. The data presented in Table 2 include an annual summary of the minimum, maximum, and average for each of the CSLAP sampling parameters, including data from other sources for which sufficient quality-assurance/quality-control documentation is available for assessing the validity of the results. This data may be useful for comparing a particular data point for the current sampling year with historical data or information. Tables 3 through 5 includes more detailed summaries of the 2007 and historical data sets, including some evaluation of water-quality trends, comparison against existing water-quality standards, and whether 2007 represented a typical year.

Graphs

The second form of data analysis for your lake is presented in the form of graphs. These graphs are based on the raw data sets to represent a snapshot of water-quality conditions at your lake. The more sampling that has been done on a particular lake, the more information that can be presented on the graph, and the more information you have to identify annual trends for your lake. For example, a lake that has been doing CSLAP monitoring consistently for five years will have a graph depicting five years’ worth of data, whereas a lake that has been doing CSLAP sampling for only one year will only have one. Therefore, it is important to consider the number of sampling years of information in addition to where the data points fall on a graph when trying to draw conclusions about annual trends. There are certain factors not accounted for in this report that lake managers should consider:

- **Local weather conditions** (high or low temperatures, rainfall, droughts or hurricanes). Due to delays in receiving meteorological data from NOAA stations within NYS, weather data from individual weather stations or the present sampling season are not included in these reports. Some of the variability reported below can be attributed more to weather patterns than to a “real” water trend or change. However, it is presumed that much of the sampling “noise” associated with weather is dampened over multiple years of data collection and thus should not significantly influence the limited trend analyses provided for CSLAP lakes with longer and larger databases.
- **Sampling season and parameter limitations.** Because sampling is generally confined to June-September, this report does not look at CSLAP parameters during the winter and other seasons. Winter conditions can impact the usability and water-quality of a lake. In addition, there are other sampling parameters (fecal coliform, dissolved oxygen, etc.) that may be responsible for chemical and biological processes and changes in physical measurements (such as water clarity) and the perceived conditions in the lake. *The 2007 CSLAP report attempts to standardize some comparisons by limiting the evaluation to the summer recreational season and the most common sampling periods (mid-June through mid-September), in the event that samples are collected at other times of the year (such as May or October) during only some sampling seasons.*

TABLE 2: CSLAP Data Summary for Eagle Lake

Year	Min	Avg	Max	N	Parameter
2000-07	4.50	6.31	8.20	63	CSLAP Zsd
2007	5.10	5.94	6.65	8	CSLAP Zsd
2006	4.95	6.24	7.55	8	CSLAP Zsd
2005	4.50	5.92	7.35	8	CSLAP Zsd
2004	5.60	6.23	6.80	8	CSLAP Zsd
2003	4.75	6.09	8.20	8	CSLAP Zsd
2002	5.50	6.39	7.05	8	CSLAP Zsd
2001	6.95	7.24	7.60	7	CSLAP Zsd
2000	5.00	6.51	7.95	8	CSLAP Zsd
1999	6.40	7.90	9.50	3	LCI Zsd
1932	7.01	7.01	7.01	1	DEC Zsd
Year	Min	Avg	Max	N	Parameter
2000-07	0.002	0.006	0.013	61	CSLAP Tot.P
2007	0.004	0.006	0.006	8	CSLAP Tot.P
2007	0.004	0.005	0.006	8	CSLAP Hyp TP
2006	0.003	0.005	0.009	8	CSLAP Tot.P
2006	0.005	0.006	0.008	8	CSLAP Hyp TP
2005	0.004	0.006	0.010	8	CSLAP Tot.P
2005	0.005	0.007	0.010	7	CSLAP Hyp TP
2004	0.003	0.006	0.013	7	CSLAP Tot.P
2004	0.003	0.005	0.009	8	CSLAP Hyp TP
2003	0.002	0.006	0.008	7	CSLAP Tot.P
2003	0.003	0.007	0.013	8	CSLAP Hyp TP
2002	0.002	0.006	0.009	8	CSLAP Tot.P
2002	0.002	0.007	0.013	8	CSLAP Hyp TP
2001	0.004	0.006	0.012	7	CSLAP Tot.P
2000	0.004	0.007	0.010	8	CSLAP Tot.P
1999	0.006	0.007	0.008	2	LCI Tot.P
1999	0.008	0.010	0.011	2	LCI Hypo Tot.P
Year	Min	Avg	Max	N	Parameter
2000-07	0.00	0.01	0.12	62	CSLAP NO3
2007	0.00	0.02	0.12	8	CSLAP NO3
2006	0.01	0.02	0.03	8	CSLAP NO3
2005	0.01	0.01	0.03	8	CSLAP NO3
2004	0.01	0.01	0.03	8	CSLAP NO3
2004	0.01	0.01	0.02	8	CSLAP HyNO3
2003	0.00	0.01	0.02	8	CSLAP NO3
2003	0.00	0.01	0.02	8	CSLAP HyNO3
2002	0.00	0.00	0.01	8	CSLAP NO3
2002	0.00	0.00	0.01	8	CSLAP HyNO3
2001	0.01	0.01	0.01	7	CSLAP NO3
2000	0.01	0.01	0.01	7	CSLAP NO3
1999	0.01	0.01	0.01	2	LCI NO3

DATA SOURCE KEY

CSLAP	New York Citizens Statewide Lake Assessment Program
LCI	the NYSDEC Lake Classification and Inventory Survey conducted during the 1980s and again beginning in 1996 on select sets of lakes, typically 1 to 4x per year
DEC	other water-quality data collected by the NYSDEC Divisions of Water and Fish and Wildlife, typically 1 to 2x in any give year
ALSC	the NYSDEC (and other partners) Adirondack Lake Survey Corporation study of more than 1500 Adirondack and Catskill lakes during the mid 1980s, typically 1 to 2x
ELS	USEPA's Eastern Lakes Survey, conducted in the fall of 1982, 1x
NES	USEPA's National Eutrophication Survey, conducted in 1972, 2 to 10x
EMAP	USEPA and US Dept. of Interior's Environmental Monitoring and Assessment Program conducted from 1990 to present, 1 to 2x in four year cycles
Additional data source codes are provided in the individual lake reports	

CSLAP DATA KEY:

The following key defines column headings and parameter results for each sampling season:

Min	Minimum reading for the parameter
Avg	Geometric average (mean) reading for the parameter
Max	Maximum reading for the parameter
N	Number of samples collected
Zsd	Secchi disk transparency, meters
Tot.P	Total Phosphorus as P, in mg/l (Hypo = bottom sample)
NO3	Nitrate + Nitrite nitrogen as N, in mg/l
NH4	Ammonia as N, in mg/l
TDN	Total Dissolved Nitrogen as N, in mg/l
TN	Total Nitrogen as N, in mg/l
TP/TN	Phosphorus/Nitrogen ratios, unitless (calculated from TDN)
Ca	Calcium, in mg/l
Tcolor	True color, as platinum color units
pH	(negative logarithm of hydrogen ion concentration), standard pH
Cond25	Specific conductance corrected to 25°C, in µmho/cm
Chl.a	Chlorophyll a, in µg/l
QA	Survey question re: physical condition of lake: (1) crystal clear; (2) not quite crystal clear; (3) definite algae greenness; (4) high algae levels; and (5) severely high algae levels
QB	Survey question re: aquatic plant populations of lake: (1) none visible; (2) visible underwater; (3) visible at lake surface; (4) dense growth at lake surface; (5) dense growth completely covering the nearshore lake surface
QC	Survey question re: recreational suitability of lake: (1) couldn't be nicer; (2) very minor aesthetic problems but excellent for overall use; (3) slightly impaired; (4) substantially impaired, although lake can be used; (5) recreation impossible
QD	Survey question re: factors affecting answer QC: (1) poor water clarity; (2) excessive weeds; (3) too much algae/odor; (4) lake looks bad; (5) poor weather; (6) litter, surface debris, beached/floating material; (7) too many lake users (boats, PWCs, etc); (8) other

TABLE 2: CSLAP Data Summary for Eagle Lake (cont)

Year	Min	Avg	Max	N	Parameter
2002-07	0.00	0.02	0.12	48	CSLAP NH4
2007	0.01	0.04	0.12	8	CSLAP NH4
2006	0.01	0.02	0.04	8	CSLAP NH4
2005	0.01	0.01	0.04	8	CSLAP NH4
2004	0.01	0.01	0.02	8	CSLAP NH4
2004	0.01	0.02	0.09	8	CSLAP HyNH4
2003	0.00	0.01	0.02	8	CSLAP NH4
2003	0.00	0.01	0.02	8	CSLAP HyNH4
2002	0.01	0.03	0.05	8	CSLAP NH4
2002	0.01	0.02	0.04	8	CSLAP HyNH4
Year	Min	Avg	Max	N	Parameter
2002-07	0.10	0.40	1.13	47	CSLAP TDN
2007	0.44	0.60	1.13	8	CSLAP TDN
2006	0.36	0.52	0.61	8	CSLAP TDN
2005	0.10	0.22	0.61	8	CSLAP TDN
2004	0.15	0.28	0.57	7	CSLAP TDN
2004	0.20	0.39	1.05	7	CSLAP HyTDN
2003	0.11	0.28	0.41	8	CSLAP TDN
2003	0.03	0.26	0.38	7	CSLAP HyTDN
2002	0.33	0.48	0.74	8	CSLAP TDN
2002	0.31	0.43	0.58	8	CSLAP HyTDN
Year	Min	Avg	Max	N	Parameter
2002-07	9.98	102.46	439.85	46	CSLAP TN/TP
2007	158.67	245.83	439.85	8	CSLAP TN/TP
2006	52.52	106.73	177.59	8	CSLAP TN/TP
2005	12.30	46.21	140.48	8	CSLAP TN/TP
2004	15.95	67.39	135.61	6	CSLAP TN/TP
2004	21.80	96.21	328.55	7	CSLAP HyTN/TP
2003	9.98	53.52	133.67	8	CSLAP TN/TP
2003	1.87	52.61	124.40	7	CSLAP HyTN/TP
2002	46.00	86.29	172.97	8	CSLAP TN/TP
2002	34.03	83.93	214.16	8	CSLAP HyTN/TP
Year	Min	Avg	Max	N	Parameter
2000-07	2	8	19	60	CSLAP TColor
2007	7	10	13	8	CSLAP TColor
2006	2	8	14	8	CSLAP TColor
2005	6	8	11	7	CSLAP TColor
2004	4	12	19	8	CSLAP TColor
2003	2	8	15	7	CSLAP TColor
2002	2	7	16	8	CSLAP TColor
2001	3	5	8	7	CSLAP TColor
2000	3	5	8	7	CSLAP TColor

TABLE 2: CSLAP Data Summary for Eagle Lake (cont)

Year	Min	Avg	Max	N	Parameter
2000-07	6.31	7.59	8.80	60	CSLAP pH
2007	7.26	7.88	8.46	8	CSLAP pH
2006	7.35	7.74	8.33	7	CSLAP pH
2005	7.30	7.90	8.80	8	CSLAP pH
2004	6.31	7.34	8.04	8	CSLAP pH
2003	6.33	7.31	7.72	7	CSLAP pH
2002	6.71	7.58	7.95	8	CSLAP pH
2001	6.80	7.48	7.94	7	CSLAP pH
2000	6.75	7.46	8.15	7	CSLAP pH
1999	7.30	7.50	7.70	2	LCI pH
1932	7.20	7.20	7.20	1	DEC pH
Year	Min	Avg	Max	N	Parameter
2000-07	102	137	168	61	CSLAP Cond25
2007	102	127	151	8	CSLAP Cond25
2006	115	129	145	8	CSLAP Cond25
2005	129	138	147	8	CSLAP Cond25
2004	115	137	168	8	CSLAP Cond25
2003	146	150	157	7	CSLAP Cond25
2002	140	142	147	8	CSLAP Cond25
2001	117	139	153	7	CSLAP Cond25
2000	129	136	139	7	CSLAP Cond25
1999	140	140	140	2	LCI Cond25
Year	Min	Avg	Max	N	Parameter
2002-07	9.9	11.8	13.0	9	CSLAP Ca
2007	9.9	11.1	12.4	2	CSLAP Ca
2006	11.0	11.6	12.3	2	CSLAP Ca
2005	11.8	12.0	12.3	2	CSLAP Ca
2004	11.8	11.8	11.8	1	CSLAP Ca
2003	12.0	12.5	13.0	2	CSLAP Ca
2002				0	CSLAP Ca
Year	Min	Avg	Max	N	Parameter
2000-07	0.16	1.10	8.20	55	CSLAP Chl.a
2007	0.41	0.89	1.91	8	CSLAP Chl.a
2006	0.35	0.91	1.94	8	CSLAP Chl.a
2005	0.70	0.94	1.11	7	CSLAP Chl.a
2004	0.16	1.08	2.50	7	CSLAP Chl.a
2003	0.44	0.87	1.47	8	CSLAP Chl.a
2002	0.75	1.03	1.23	6	CSLAP Chl.a
2001	0.77	1.01	1.51	4	CSLAP Chl.a
2000	0.42	2.13	8.20	7	CSLAP Chl.a
1999	1.07	1.43	2.29	4	LCI Chl.a

TABLE 2: CSLAP Data Summary for Eagle Lake (cont)

Year	Min	Avg	Max	N	Parameter
2000-07	2	2.0	2	58	QA
2007	2	2.0	2	8	QA
2006	2	2.0	2	8	QA
2005	2	2.0	2	8	QA
2004	2	2.0	2	8	QA
2003	2	2.0	2	7	QA
2002	2	2.0	2	8	QA
2001	2	2.0	2	3	QA
2000	2	2.0	2	8	QA
Year	Min	Avg	Max	N	Parameter
2000-07	1	2.8	3	58	QB
2007	2	2.9	3	8	QB
2006	1	2.8	3	8	QB
2005	3	3.0	3	8	QB
2004	3	3.0	3	8	QB
2003	2	2.7	3	7	QB
2002	2	2.8	3	8	QB
2001	3	3.0	3	3	QB
2000	1	2.6	3	8	QB
Year	Min	Avg	Max	N	Parameter
2000-07	2	3.0	4	58	QC
2007	3	3.1	4	8	QC
2006	2	3.0	4	8	QC
2005	3	3.0	3	8	QC
2004	3	3.0	3	8	QC
2003	3	3.0	3	7	QC
2002	2	2.9	3	8	QC
2001	3	3.0	3	3	QC
2000	3	3.0	3	8	QC

- Statistical analyses.** True assessments of water-quality trends and comparison to other lakes involve rigid statistical analyses. Such analyses are generally beyond the scope of this program, in part due to limitations on the time available to summarize data from more than 100 lakes in the five months from data receipt to the next sampling season. This may be due in part to the inevitable inter-lake inconsistencies in sampling dates from year to year and in part to the limited scope of monitoring. Where appropriate, some statistical summaries, utilizing both parametric and non-parametric statistics, have been provided within the report (primarily in Table 2).
- Mean versus Median.** Much of the water-quality summary data presented in this report is reported as the mean, or the average of all of the readings in the period in question (summer, annual, year to year). However, while mean remains one of the most useful, and often most powerful, ways to estimate the most typical reading for many of the measured water-quality indicators, it is a less useful and perhaps misleading estimate

when the data are not “normally” distributed (most common readings in the middle of the range of all readings, with readings less common toward the end of the range).

In particular, comparisons of one lake to another, such as comparisons within a particular basin, can be greatly affected by the spread of the data across the range of all readings. For example, the average phosphorus level of nine lakes with very low readings (say 10 $\mu\text{g/l}$) and one lake with very high readings (say 110 $\mu\text{g/l}$) could be much higher (in this case, 20 $\mu\text{g/l}$) than in the “typical lake” in this set of lakes (much closer to 10 $\mu\text{g/l}$). In this case, median, or the middle reading in the range, is probably the most accurate representation of “typical”.

This report will include the use of both mean and median to evaluate “central tendency,” or the most typical reading, for the indicator in question. In most cases, “mean” is used most often to estimate central tendency. However, where noted, “median” may also be used.

**TABLE 3- Current and Historical Data Summaries for Eagle Lake
Eutrophication Indicators**

Parameter	Year	Minimum	Average	Maximum
Zsd	2007	5.10	5.94	6.65
(meters)	All Years	4.50	6.31	8.20
Parameter	Year	Minimum	Average	Maximum
Phosphorus	2007	0.004	0.006	0.006
(mg/l)	All Years	0.002	0.006	0.013
Parameter	Year	Minimum	Average	Maximum
Chl.a	2007	0.41	0.89	1.91
(µg/l)	All Years	0.16	1.11	8.20

Parameter	Year	Was 2007 Clarity the Highest or Lowest on Record?	Was 2007 a Typical Year?	Trophic Category	Zsd Changing?	% Samples Violating DOH Beach Std?
Zsd	2007	Within Normal Range	Yes	Oligotrophic	Decreasing?	0
(meters)	All Years			Oligotrophic		0
Parameter	Year	Was 2007 TP the Highest or Lowest on Record?	Was 2007 a Typical Year?	Trophic Category	TP Changing?	% Samples Exceeding TP Guidance Value
Phosphorus	2007	Within Normal Range	Yes	Oligotrophic	No	0
(mg/l)	All Years			Oligotrophic		0
Parameter	Year	Was 2007 Algae the Highest or Lowest on Record?	Was 2007 a Typical Year?	Trophic Category	Chl.a Changing?	
Chl.a	2007	Within Normal Range	Yes	Oligotrophic	No	
(µg/l)	All Years			Oligotrophic		

Minimum allowable water clarity for siting a new NYS swimming beach = 1.2 meters
 NYS Total Phosphorus Guidance Value for Class B and Higher Lakes = 0.020 mg/l

The CSLAP dataset indicates that Eagle Lake is an *oligotrophic*, or highly unproductive lake, based on phosphorus, chlorophyll *a* and Secchi disk readings. This has been a consistent assessments in the lake. Water clarity readings have dropped slightly in recent years, despite stable phosphorus readings and algae levels that have also been lower in recent years. Lower water transparency readings have corresponded to slightly higher color readings. There is not a strong correlation between changes in clarity and algae, and between changes in phosphorus and algae, although it is likely that significant increases in nutrients and algae would trigger a decrease in water clarity. Water clarity readings increase slightly over the course of a typical sampling season, in part due to seasonally decreasing surface phosphorus readings and to deepwater phosphorus levels that are close to those measured at the lake surface. Phosphorus readings consistently are below the state guidance value for lakes used for contact recreation (swimming), and Secchi disk transparency readings regularly exceed the minimum recommended water clarity for swimming beaches (= 1.2 meters). In short, the productivity of Eagle Lake varies only slightly from year to year, perhaps due to influences from weather conditions. Water transparency readings have decreased slightly in the last several years, perhaps due to wetter weather and due to the influence of water color.

**TABLE 4- Current and Historical Data Summaries for Eagle Lake (cont.)
Other Water-Quality Indicators**

Parameter	Year	Minimum	Average	Maximum
Nitrate	2007	0.00	0.02	0.12
(mg/l)	All Years	0.00	0.01	0.12
Parameter	Year	Minimum	Average	Maximum
NH4	2007	0.01	0.04	0.12
(mg/l)	All Years	0.00	0.02	0.12
Parameter	Year	Minimum	Average	Maximum
TDN	2007	0.44	0.60	1.13
(mg/l)	All Years	0.10	0.40	1.13
Parameter	Year	Minimum	Average	Maximum
True Color	2007	7	10	13
(ptu)	All Years	2	9	46
Parameter	Year	Minimum	Average	Maximum
pH	2007	7.26	7.88	8.46
(std units)	All Years	6.31	7.60	8.80
Parameter	Year	Minimum	Average	Maximum
Conductivity	2007	102	127	151
(µmho/cm)	All Years	102	137	168
Parameter	Year	Minimum	Average	Maximum
Calcium	2007	9.9	11.1	12.4
(mg/l)	All Years	9.9	11.8	13.0

These data indicate Eagle Lake is a weakly colored, weakly alkaline (above neutral pH) lake with low nitrate and ammonia levels, and soft water. Water transparency readings may have been affected by higher water color in recent years, although the very high water clarity readings in the lake are the result of very low algae levels. Water color readings have varied slightly from year to year, and these readings have increased slightly over the last several years. Nitrate and ammonia readings are regularly at or below the analytical detection limit, in both surface and bottom waters, and neither appear to represent a threat to water-quality. The small changes in nitrogen have probably been within the normal range of variability for Eagle Lake, although nitrate readings have been slightly higher in recent years. pH readings are indicative of alkaline (above neutral) lakes, and nearly all readings have been within the state water-quality standards (=6.5 to 8.5). These readings were slightly higher than normal in the last few years, but it is likely that these pH readings are adequate to support most aquatic organisms. Most conductivity readings have been typical of lakes with moderately soft water. Calcium levels are near the threshold found to support zebra mussels, although it is not believed that these exotic animals have been found in Eagle Lake.

**TABLE 4- Current and Historical Data Summaries for Eagle Lake (cont.)
Other Water-Quality Indicators (cont)**

Parameter	Year	Was 2007 Nitrate the Highest or Lowest on Record?	Was 2007 a Typical Year?	Nitrate High?	Nitrate Changing?	% Samples Exceeding NO3 Standard	
Nitrate	2007	Highest at Times	Yes	No	Increasing?	0	
(mg/l)	All Years			No		0	
Parameter	Year	Was 2007 Ammonia the Highest or Lowest on Record?	Was 2007 a Typical Year?	Ammonia High?	Ammonia Changing?	% Samples Exceeding NH4 Standard+	
NH4	2007	Highest at Times	Yes	No	No	0	
(mg/l)	All Years			No		0	
Parameter	Year	Was 2007 TDN the Highest or Lowest on Record?	Was 2007 a Typical Year?	TDN High?	TDN Changing?	Ratios of TN/TP Indicate P or N Limitation?	
TDN	2007	Highest at Times	Yes	No	No	P Limitation	
(mg/l)	All Years			No		P Limitation	
Parameter	Year	Was 2007 Color the Highest or Lowest on Record?	Was 2007 a Typical Year?	Colored Lake?	Color Changing?		
True Color	2007	Within Normal Range	Yes	No	Increasing?		
(ptu)	All Years			No			
Parameter	Year	Was 2007 pH the Highest or Lowest on Record?	Was 2007 a Typical Year?	Acceptable Range?	pH Changing?	% Samples > Upper pH Standard+	% Samples < Lower pH Standard+
pH	2007	Within Normal Range	Yes	Yes	No	0	0
(std units)	All Years			Yes		2	5
Parameter	Year	Was 2007 Conductivity Highest or Lowest on Record?	Was 2007 a Typical Year?	Relative Hardness	Conduct. Changing?		
Conductivity	2007	Lowest at Times	Yes	Intermediate	No		
(µmho/cm)	All Years			Intermediate			
Parameter	Year	Was 2007 Calcium Highest or Lowest on Record?	Was 2007 a Typical Year?	Support Zebra Mussels?	Calcium Changing?		
Calcium	2007	Lowest at Times	Yes	Uncertain	No		
(mg/l)	All Years			Uncertain			

NYS Nitrate standard = 10 mg/l

NYS Ammonia standard = 2 mg/l (as NH₃-NH₄)

NYS pH standard- 6.5 < acceptable pH < 8.5

TABLE 5- Current and Historical Data Summaries for Eagle Lake

Lake Perception Indicators (1= most favorable, 5= least favorable)

Parameter	Year	Minimum	Average	Maximum
QA	2007	2	2.0	2
(Clarity)	All Years	2	2.0	2
Parameter	Year	Minimum	Average	Maximum
QB	2007	2	2.9	3
(Plants)	All Years	1	2.8	3
Parameter	Year	Minimum	Average	Maximum
QC	2007	3	3.1	4
(Recreation)	All Years	2	3.0	4

Parameter	Year	Was 2007 Clarity the Highest or Lowest on Record?	Was 2007 a Typical Year?	Clarity Changed?	%Frequency 'Definite Algae Greenness'	%Frequency 'Severe Algae Levels'	%Frequency 'Slightly Impaired' Due to Algae	%Frequency 'Substantially Impaired' Due to Algae
QA	2007	Highest and Lowest	Yes	No	0	0	0	0
(Clarity)	All Years				0	0	6	0
Parameter	Year	Was 2007 Weed Growth the Heaviest on Record?	Was 2007 a Typical Year?	Weeds Changed?	%Frequency Surface Weeds	%Frequency Dense Weeds	%Frequency 'Slightly Impaired' Due to Weeds	%Frequency 'Substantially Impaired' Due to Weeds
QB	2007	Heaviest at Times	Yes	No	88	0	88	13
(Plants)	All Years				86	0	83	3
Parameter	Year	Was 2007 Recreation the Best or Worst on Record?	Was 2007 a Typical Year?	Recreation Changed?	%Frequency Slightly Impaired	%Frequency Substantially Impaired		
QC	2007	Worst at Times	Yes	No	100	13		
(Recreation)	All Years				97	3		

Recreational, water quality, and aquatic plant assessments of Eagle Lake have varied little since CSLAP sampling began in 2000, perhaps as expected given the stability in water quality conditions. Eagle Lake has consistently been described as “not quite crystal clear”, an assessment slightly less favorable than in other lakes with similar water clarity readings. However, this also suggests that recent drops in water clarity may not be apparent to most lake users. Aquatic plants regularly grow to the lake surface, and have frequently been cited as impacting recreational uses of the lake. As a result, the lake has consistently been described as “slightly impaired” for some recreational uses, although these assessments are sometimes impacted by poor weather. These assessments vary little during the sampling season, consistent with seasonally stable water quality and aquatic plant coverage.

Eagle Lake has been described by the CSLAP sampling volunteers as “slightly impaired” during 97% of the CSLAP sampling sessions, and “substantially” impaired 3% of the time. Slightly impaired conditions were associated with excessive weeds during 86% of the CSLAP sampling sessions, and with poor water clarity and excessive algae growth 6% of the time. “Substantially” impaired conditions were associated with excessive weeds 3% of the time, but never with excessive algae and poor water clarity.

How Do the 2007 Data Compare to Historical Data from Eagle Lake?

Seasonal Comparison of Eutrophication, Other Water-quality, and Lake-Perception Indicators—2007 Sampling Season and in the Typical or Previous Sampling Seasons at Eagle Lake

Figures 23 and 24 compare data for the measured eutrophication parameters for Eagle Lake in 2007 and since CSLAP sampling began at Eagle Lake. Figures 25 and 26 compare nitrogen to phosphorus ratios, figures 27 through 34 compare other sampling indicators, and figures 35 and 36 compare volunteer perception responses during the same periods.

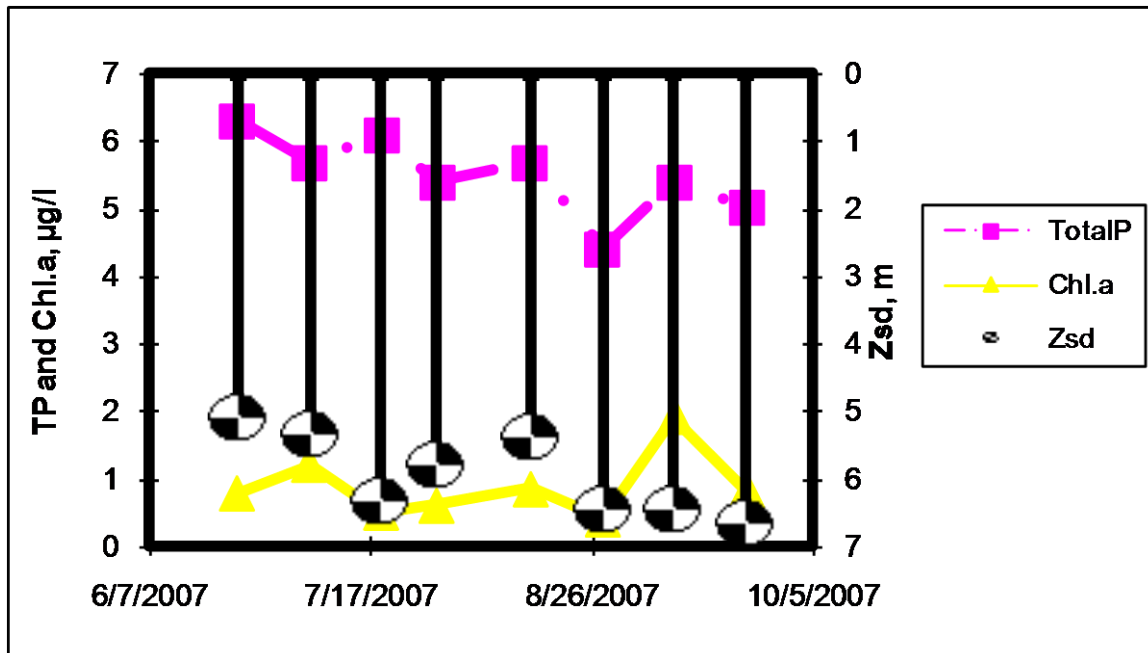


Figure 23. 2007 Eutrophication Data for Eagle Lake

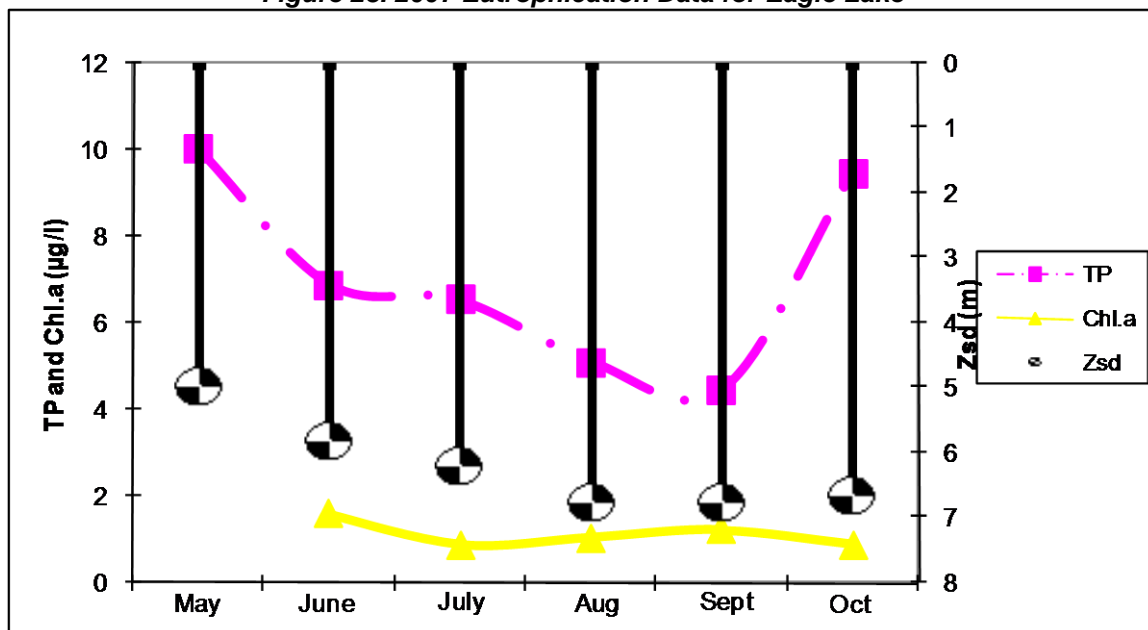


Figure 24- Eutrophication Data in a Typical (Monthly Mean) Year for Eagle Lake

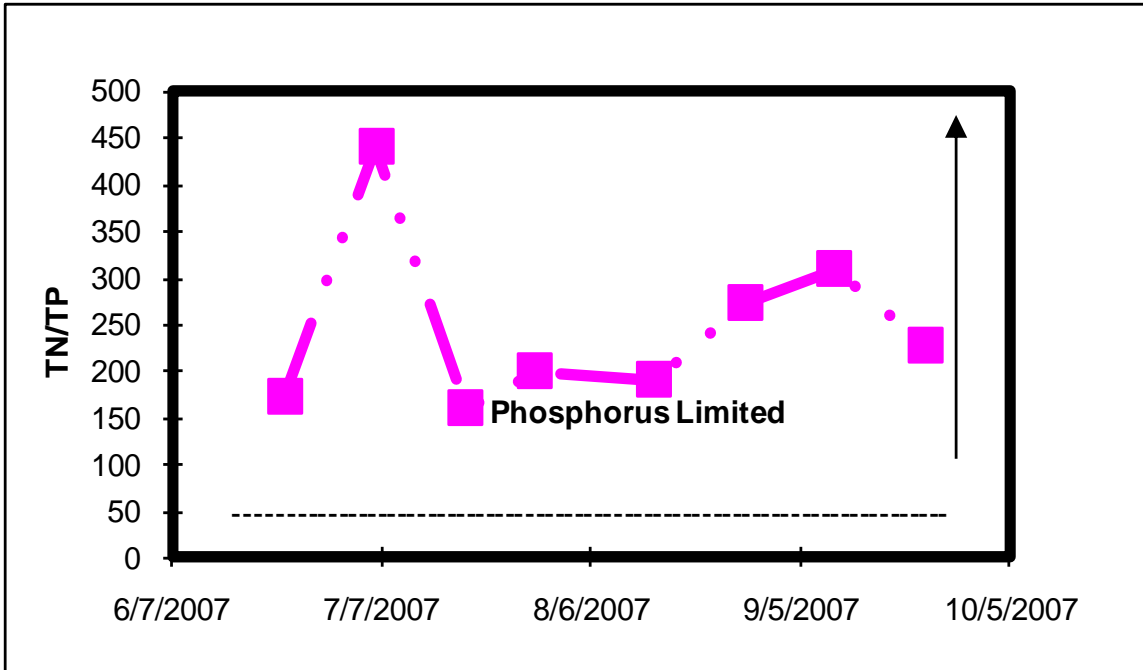


Figure 25. 2007 Nitrogen-to-Phosphorus Ratios for Eagle Lake

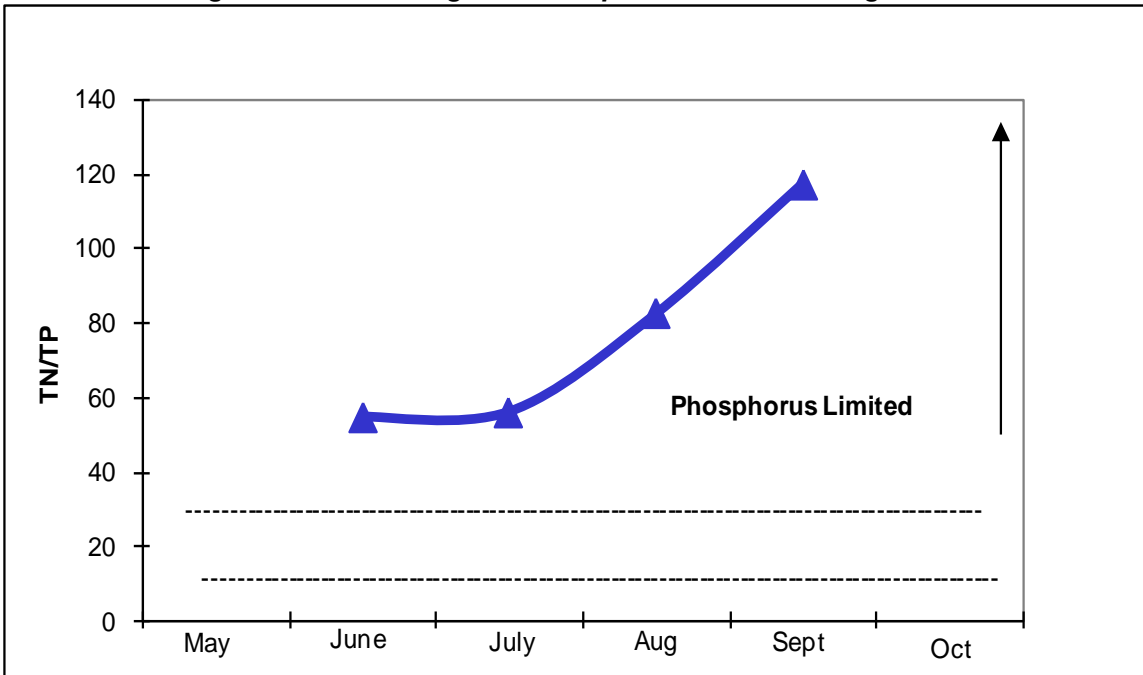


Figure 26- Nitrogen-to-Phosphorus Ratios in a Typical (Monthly Mean) Year for Eagle Lake

Annual Averages, 1996-06

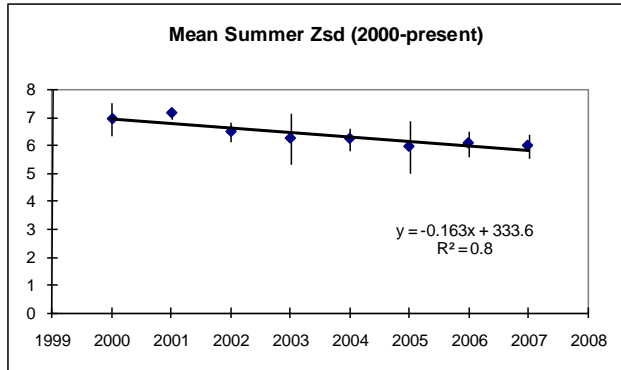


Figure 27. Annual Average Summer Water Clarity for Eagle Lake

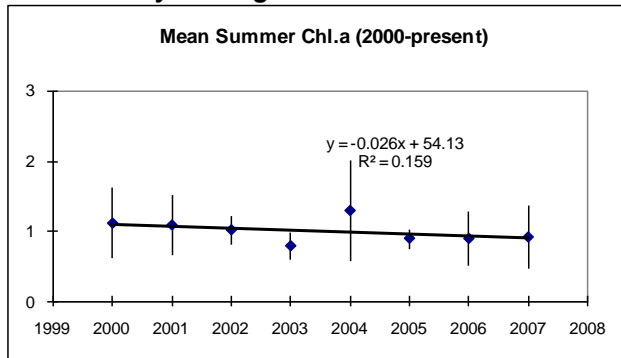


Figure 28. Annual Average Summer Chlorophyll a for Eagle Lake

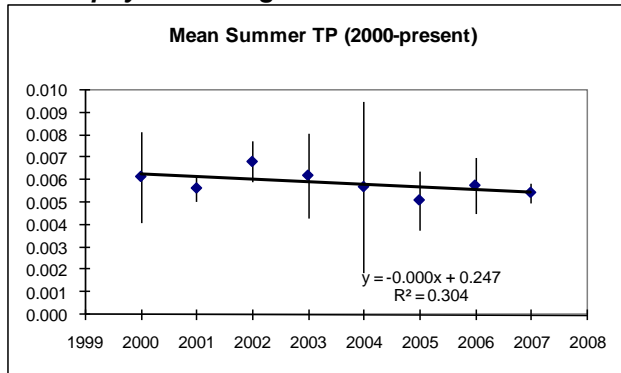


Figure 29. Annual Average Summer Total Phosphorus for Eagle Lake

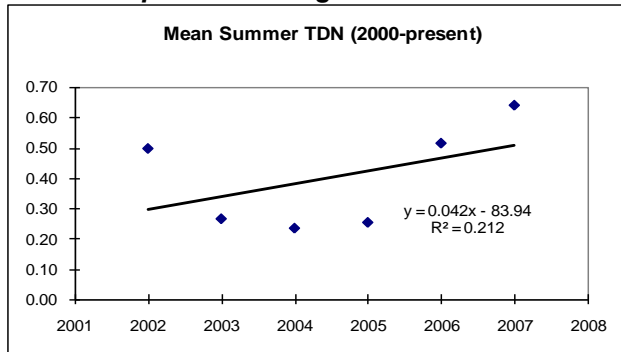


Figure 30. Annual Average Summer Total Nitrogen for Eagle Lake

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Clarity: 2001, 2000
 Lowest Clarity: 2005, 2007
 Long Term Trend?: Decreasing
Discussion: Water transparency readings have decreased over the last eight years, despite stable to decreasing chlorophyll and phosphorus readings.

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Chl.a: 2003, 2006
 Lowest Chl.a: 2004, 2000
 Long Term Trend?: None apparent
Discussion: Chlorophyll readings have varied from only slightly year to year. This variability has not been consistent with changes in precipitation or other water quality indicators

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest TP: 2002, 2003
 Lowest TP: 2004, 1994
 Long Term Trend?: None apparent?
Discussion: Phosphorus readings have decreased since 2000, but the variability from sample to sample appears to be larger than the small decrease from 2000 to 2007

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Total N: 2007, 2006, 2002
 Lowest Total N: 2004, 2005
 Long Term Trend?: No apparent trend
Discussion: Total nitrogen readings have varied from year to year, and no longer-term trends are apparent.

Annual Averages, 1996-06

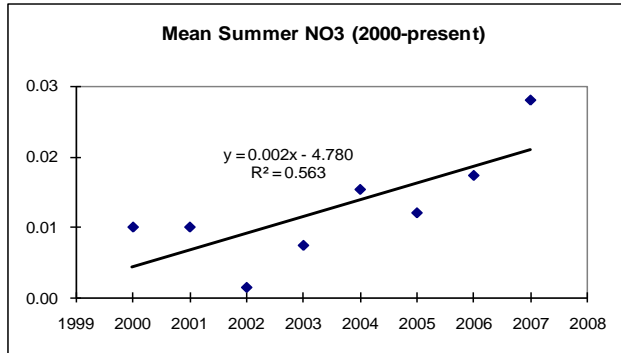


Figure 31. Annual Average Summer Nitrate for Eagle Lake

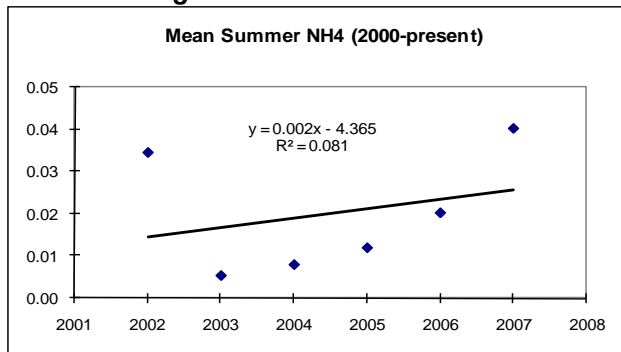


Figure 32. Annual Average Summer Ammonia for Eagle Lake

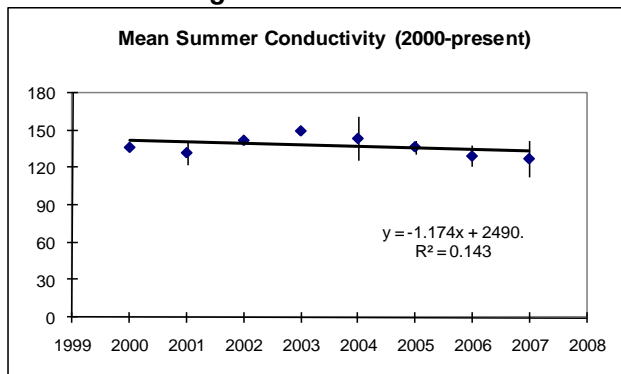


Figure 33. Annual Average Summer Conductivity for Eagle Lake

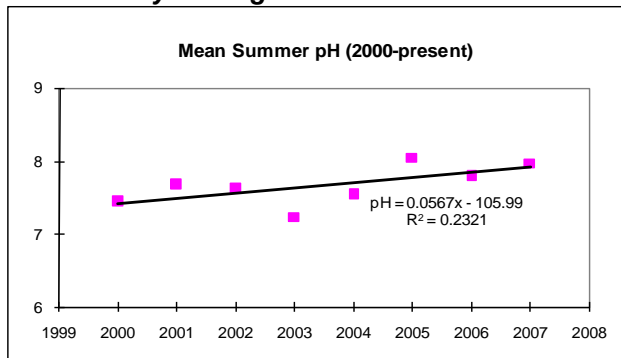


Figure 34. Annual Average Summer pH for Eagle Lake

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Nitrate: 2007, 2006
 Lowest Nitrate: 2002, 2003
 Long Term Trend?: Increasing?
Discussion: Nitrate readings have steadily increased from 2002 through 2007, although all readings continue to be very low

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Ammonia: 2007, 2002
 Lowest Ammonia: 2003, 2004
 Long Term Trend?: No apparent trend
Discussion: Ammonia readings have varied from year to year, with little connection between precipitation and ammonia levels in the lake

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Cond.: 2003, 2004
 Lowest Cond.: 2006, 2007
 Long Term Trend?: No apparent trend
Discussion: Conductivity readings have decreased in the last five years, after increasing from 2000 to 2003. This “reversal” may reflect the influence of wetter weather in recent years, but these trends are not statistically significant

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest pH: 2005, 2007
 Lowest pH: 2003
 Long Term Trend?: None apparent
Discussion: Data in Figure 34 suggest a weak but statistically insignificant trend toward increasing pH, with no apparent influence from weather or other measured water quality indicators

Annual Averages, 1996-06

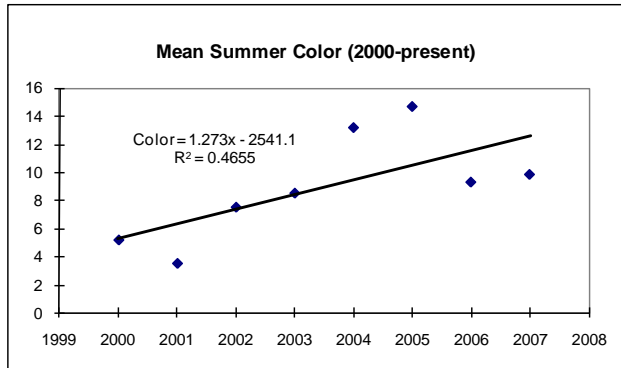


Figure 35. Annual Average Summer Color for Eagle Lake

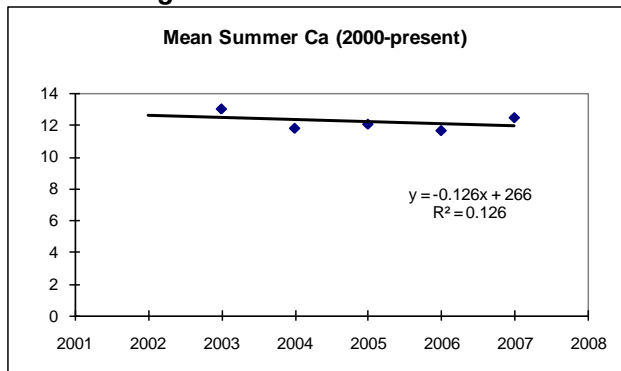


Figure 36. Annual Average Summer Calcium for Eagle Lake

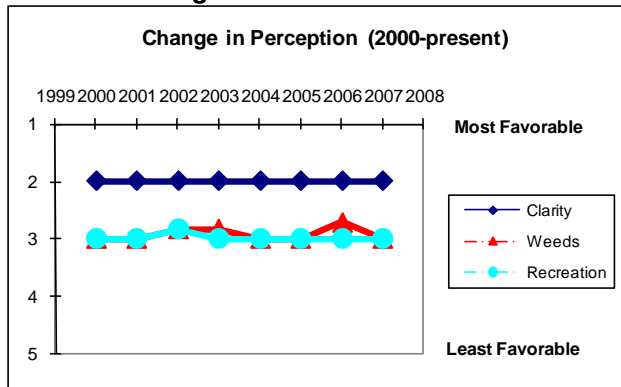


Figure 37. Annual Average Summer Lake Perception for Eagle Lake

(QA = clarity, ranging from (1) crystal clear to (3) definite algae greenness to (5) severely high algae levels;
 QB = weeds, ranging from (1) not visible to (3) growing to the surface to (5) dense growth covers lake;
 QC = recreation, ranging from (1) could not be nicer to (3) slightly impaired to (5) lake not usable)

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Color: 2005, 2004
 Lowest Color: 2001, 2000
 Long Term Trend?: Slight increase
Discussion: Water color readings have been higher in recent years, mirroring a trend found in most CSLAP lakes. However, readings in 2006 and 2007 were lower than in the previous two years, suggesting a less robust trend

Wettest Years: 2000, 2006
 Driest Years: 2002
 Highest Calcium: 2003, 2007
 Lowest Calcium: 2006
 Long Term Trend?: None apparent
Discussion: While there appears to have been a slight decrease in calcium readings since 2003, most readings have generally hovered around 12 ppm.

Wettest Years: 2000, 2006
 Driest Years: 2002
 Most Favorable WQ: same every year
 Least Favorable WQ: same every year
 Highest Weed Cov. all years except 2006, 2003
 Lowest Weed Cov. 2006, 2003
 Most Favorable Rec. 2002
 Least Favorable Rec. all other years
 Long Term Trend?: None apparent
Discussion: Recreational, water quality, and aquatic plant assessments have been very consistent from year to year, with no trends apparent for any of these indicators

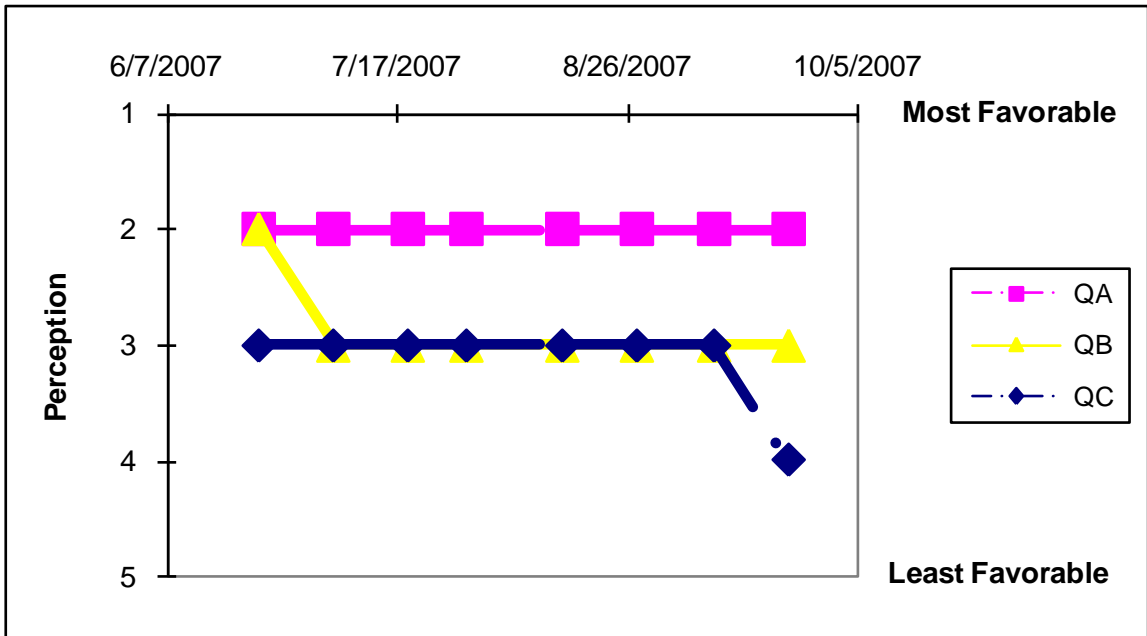


Figure 38. 2007 Lake Perception Data for Eagle Lake

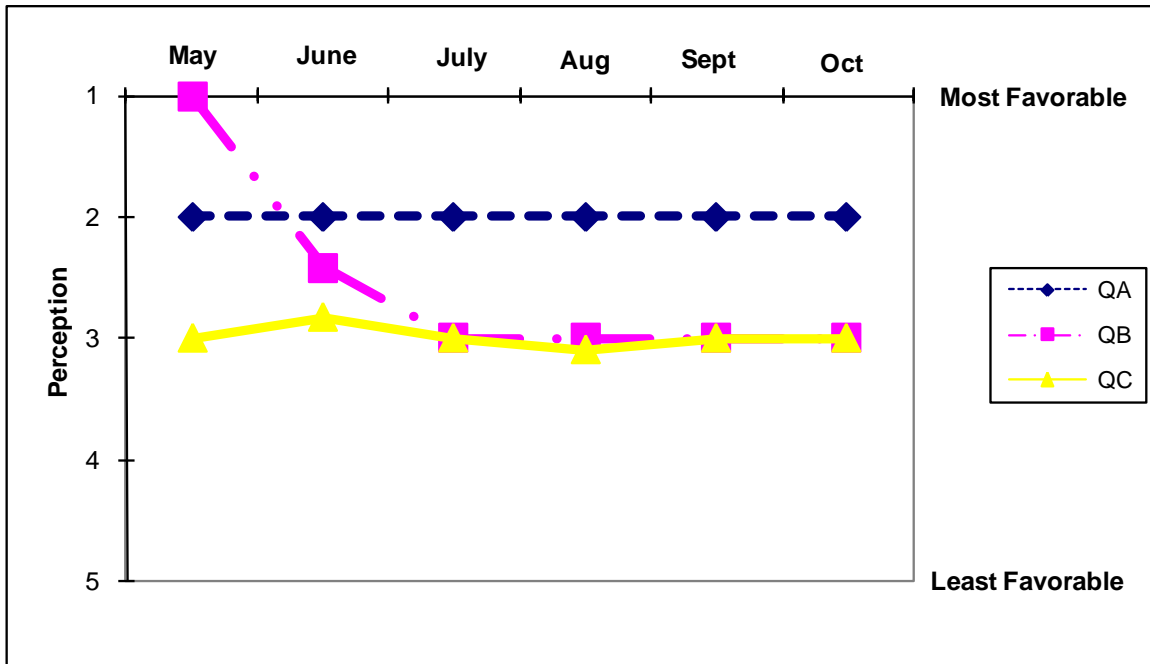


Figure 39- Lake Perception Data in a Typical (Monthly Mean) Year for Eagle Lake

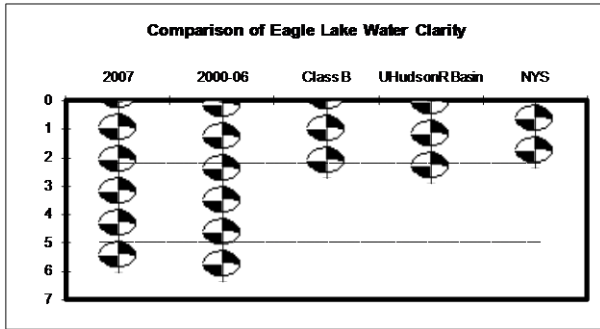


Figure 40. Comparison of 2007 Secchi Disk Transparency to Lakes With the Same Water-Quality Classification, Neighboring Lakes, and Other CSLAP Lakes

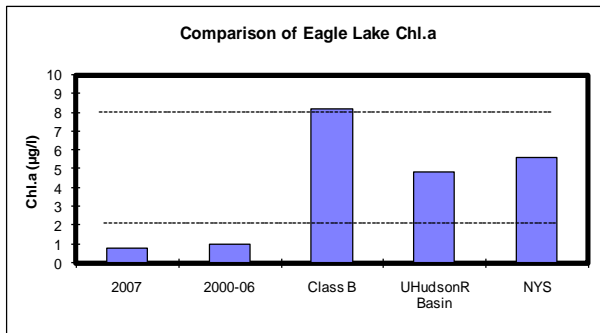


Figure 41. Comparison of 2007 Chlorophyll *a* to Lakes with the Same Water-Quality Classification, Neighboring Lakes, and Other CSLAP Lakes

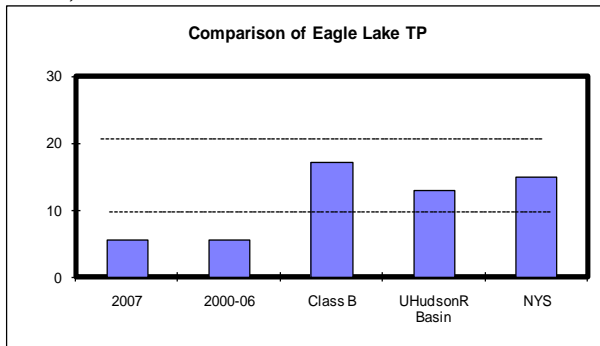


Figure 42. Comparison of 2007 Total Phosphorus to Lakes With the Same Water-Quality Classification, Neighboring Lakes, and Other CSLAP Lakes

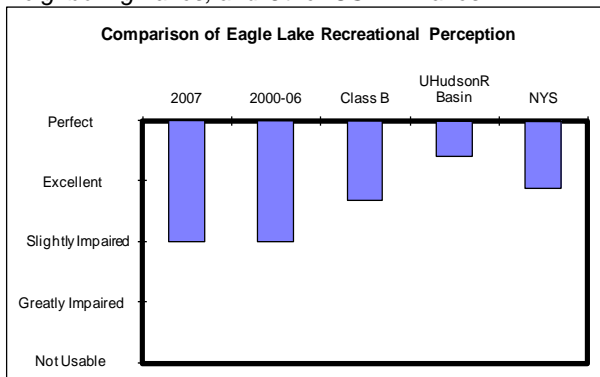


Figure 43. Comparison of 2007 Recreational Perception to Lakes With the Same Water-Quality Classification, Neighboring Lakes, and Other CSLAP Lakes

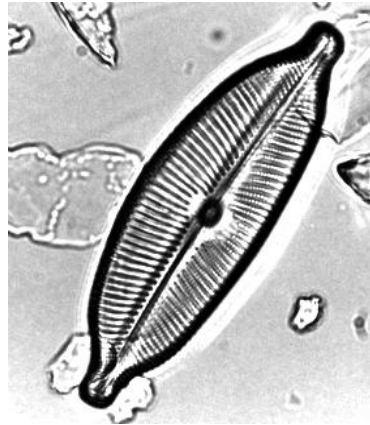
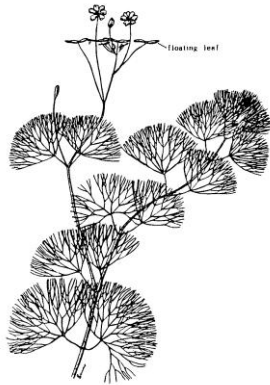
How does Eagle Lake compare to other lakes?

Annual Comparison of Median Readings for Eutrophication Parameters and Recreational Assessment For Eagle Lake in 2007 to Historical Data for Eagle Lake, Neighboring Lakes, Lakes with the Same Lake Classification, and Other CSLAP Lakes

The graphs to the left illustrate comparisons of each eutrophication parameter and recreational perception at Eagle Lake—in 2007, other lakes in the same drainage basin, lakes with the same water-quality classification (each classification is summarized in Appendix B), and all of CSLAP. Readers should note that differences in watershed types, activities, lake history and other factors may result in differing water-quality conditions at your lake relative to other nearby lakes. In addition, the limited database for some regions of the state precludes a comprehensive comparison to neighboring lakes.

Based on these graphs, the following conclusions can be made about Eagle Lake in 2007:

- Using water clarity as an indicator, Eagle Lake is less productive than other Class B lakes, other Upper Hudson River basin lakes, and other NYS lakes.
- Using chlorophyll *a* concentrations as an indicator, Eagle Lake is less productive than other Upper Hudson River basin lakes, other Class B lakes, and other NYS lakes.
- Using total phosphorus concentrations as an indicator, Eagle Lake is less productive than other NYS lakes, other Upper Hudson River basin lakes and other Class B lakes.
- Using QC on the field-observations form as an indicator, Eagle Lake is less suitable for recreation than other Upper Hudson River basin lakes, other Class B lakes, and other NYS lakes.



VII. AQUATIC PLANTS

a. Macrophytes:

Aquatic plants should be recognized for their contributions to lake beauty as well as for providing food and shelter for other life in the lake. Emergent and floating plants such as water lilies floating on the lake surface may provide aesthetic appeal with their colorful flowers; sedges and cattails help to prevent shoreline erosion and may provide food and cover for birds. Submergent plants like pondweeds and leafy waterweed harbor insects, provide nurseries for amphibians and fish, and provide food for birds and other animals. Those who enjoy fishing at the lake appreciate a diverse plant population. Aquatic plants can be found throughout the *littoral zone*, the near-shore areas in which sufficient light reaches the lake bottom to promote photosynthesis. Plant growth in any particular part of the lake is a function of available light, nutrition and space, bottom substrate, wave action, and other factors, and extensive plant growth can occur in both “clean” and “polluted” lakes. A large portion of aquatic vegetation consists of the microscopic algae referred to as phytoplankton; the other portion consists of the larger rooted plants called macrophytes.

As invasive plants colonize and spread into a lake, native plant species can be threatened or even eliminated from aquatic plant communities. The most susceptible of these are those that reside in marginal regions, limited by water depth, sediment type, or inability to compete for space. As a result, many plants identified as *rare, threatened or endangered (RTE) species* are protected under New York State law. *The New York State Natural Heritage Program has identified Northern or alpine pondweed (Potamogeton alpinus), a threatened plant species, in Eagle Lake.*

Of particular concern to many lakefront residents and recreational users are the *non-indigenous macrophytes* that can frequently dominate native aquatic plants and crowd out more beneficial plant species. The invasive plant species may be introduced to a lake by waterfowl, but in most cases they are introduced by fragments or seedlings that remain on watercraft from already-infested lakes. Once introduced, these species have tenacious survival skills, crowding out, dominating and eventually aggressively overtaking the indigenous (native) plant communities in a variety of water-quality conditions. When this occurs, they interfere with recreational activities such as fishing, swimming or water skiing. These species need to be properly identified to be effectively managed.

Non-native Invasive Macrophyte Species

For many years, four common non-native invasive species were considered the most important exotic aquatic plant species in New York lakes and ponds:

- **Eurasian watermilfoil** (*Myriophyllum spicatum*)
- **Curly-leaf pondweed** (*Potamogeton crispus*)
- **Eurasian water chestnut** (*Trapa natans*)
- **Fanwort** (*Cabomba caroliniana*)

If these plants are not present, efforts should be made to continue protecting the lake from the introduction of these species.

In addition, there are a number of other submergent or floating non-native invasive species that are becoming increasingly problematic in New York:

- **Parrotfeather** (*Myriophyllum aquaticum*)
- **Variable watermilfoil** (*Myriophyllum heterophyllum*)
- **Brazilian elodea** (*Egeria densa*)
- **European frogbit** (*Hydrocharis morsus-ranae*)

Whether the role of the lake manager is to better understand the lake ecosystem or better manage the aquatic plant community, knowledge of plant distribution is paramount to the management process. There are many procedures available for assessing and monitoring aquatic vegetation. The CSLAP Sampling Protocol contains procedures for a “semi-quantitative” plant-

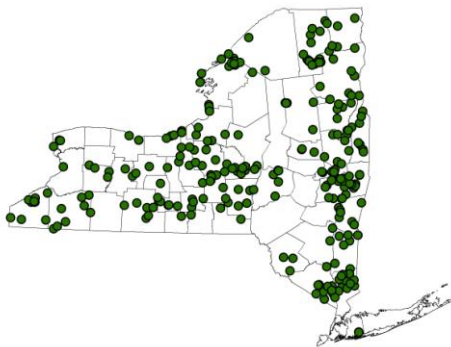


Figure 44a. *Myriophyllum spicatum* distribution in New York State



Figure 44b. *Potamogeton crispus* distribution in New York State

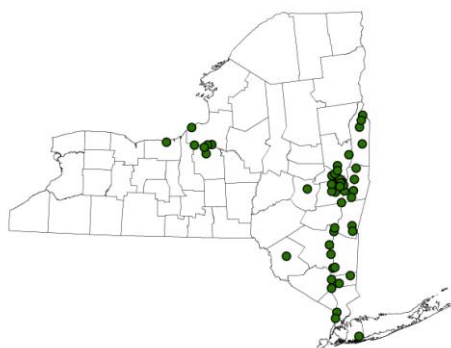


Figure 44c. *Trapa natans* distribution in New York State

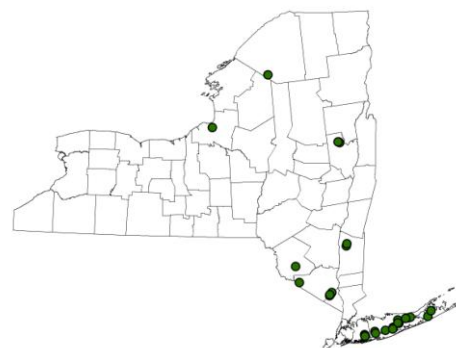


Figure 44d. *Cabomba caroliniana* distribution in New York State

monitoring program. Volunteers collect plant specimens and provide field information and qualitative abundance estimates for an assessment of the macrophyte communities within critical areas of the lake. While these techniques are no substitute for professional plant surveys, they can help provide better information for lake managers. Lake associations planning to devote significant time and expenditures toward a plant-management program are advised to pursue more extensive plant surveying activities.

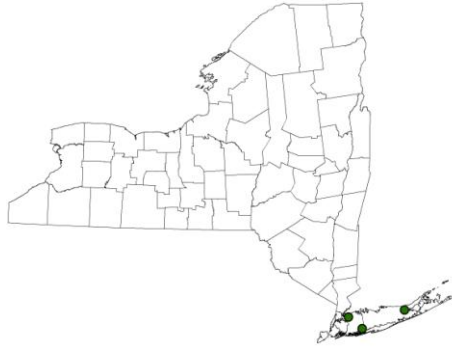


Figure 44e. *Myriophyllum aquaticum* distribution in New York State

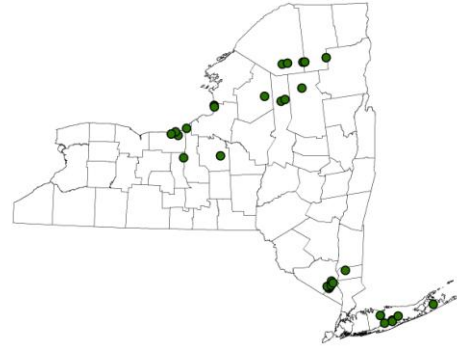


Figure 44f. *Myriophyllum heterophyllum* distribution in New York State

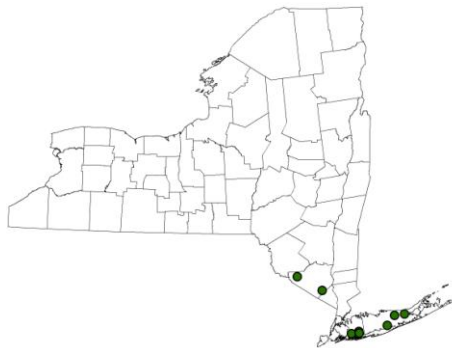


Figure 44g. *Egeria densa* distribution in New York State

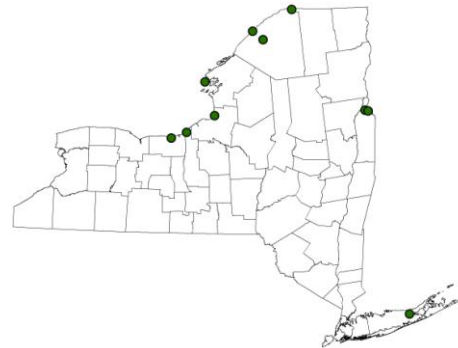


Figure 44h. *Hydrocharis morsus-ranae* distribution in New York State

Formal and informal survey work has been effective in developing statewide distribution maps of each of the major submergent exotic species, and CSLAP data has figured prominently in this process. As of 2007, the statewide distribution maps of confirmed identifications are shown on Figures 44a to 44h.

Aquatic plant surveys have not been conducted through CSLAP at Eagle Lake, although the presence of Eurasian watermilfoil (Myriophyllum spicatum) has been verified by other sources. Aquatic plant populations were evaluated as part of the 1932 Biological Survey:

ABUNDANT	COMMON	FREQUENT	RARE
<i>Potamogeton robbinsii</i> (Robbins pondweed)	<i>Potamogeton dimorphus</i> (spiral pondweed)	<i>Sparganium fluctuans</i> (floating bur reed)	<i>Sparganium eurycarpum</i> (broadfruit bur reed)
<i>Najas flexilis</i> (slender naiad)	<i>Potamogeton epihydrus</i> (ribbonleaf pondweed)	<i>Potamogeton amplifolius</i> (large leaf pondweed)	<i>Potamogeton americanus</i> (longleaf pondweed)
	<i>Elodea canadensis</i> (common waterweed)	<i>Potamogeton compressus</i> (grass-wrack pondweed)	<i>Sagittaria latifolia</i> (broadleaf arrowhead)
	<i>Scirpus subterminalis</i> (water bulrush)	<i>Potamogeton gramineus</i> (variable leaf pondweed)	<i>Eleocharis palustris</i> (common spikerush)
	<i>Eriocaulon septangulare</i> (pipewort)	<i>Potamogeton natans</i> (floating leaf pondweed)	<i>Dulichium arundinaceum</i> (threeway sedge)
	<i>Brasenia schreberi</i> (watershield)	<i>Potamogeton praelongus</i> (white-stemmed pondweed)	<i>Lobelia dortmanna</i> (Dortmann's cardinal flower)
	<i>Isoetes echinospora</i> (spiny-spored quillwort)	<i>Potamogeton pusillus</i> (small pondweed)	<i>Bidens beckii</i> (water marigold)
		<i>Vallisneria americana</i> (eelgrass, tapegrass)	
		<i>Eleocharis acicularis</i> (needle spikerush)	
		<i>Pontederia cordata</i> (pickerelweed)	
		<i>Nymphaea odorata</i> (white water lily)	
		<i>Nymphozanthus advena</i> (yellow water lily)	
		<i>Utricularia vulgaris</i> (common bladderwort)	

It is likely that many of these aquatic plants continue to be found in the lake, although Eurasian watermilfoil, probably the most dominant plant in the lake at this time, was not present in the lake in the 1930s (nor was it found in at least most, if not all, NYS lakes at this time).

b. Algae

Microscopic algae referred to as phytoplankton make up much of aquatic vegetation found in lakes. For this reason, and because phytoplankton are the primary producers of food (through photosynthesis) in lakes, they are the most important component of the complex food web that governs ecological interactions in lakes.

In a lake, phytoplankton communities are usually very diverse and are comprised of hundreds of species having different requirements for nutrients, temperature and light. In many lakes, including those of New York, diatom populations are greatest in the spring, due to a competitive advantage in cooler water and relatively high levels of silica. In most lakes, however, diatom densities rarely reach nuisance portions in the spring. By the summer, green algae take advantage of warmer temperatures and greater amounts of nutrients (particularly nitrogen) in the warm water and often increase in density. These algae often grow in higher densities than do diatoms or most other species, although they are often not the types of algae

most frequently implicated in noxious algae blooms. Later in the summer and in the early fall, blue-green algae, which possess the ability to utilize atmospheric nitrogen to provide this required nutrient, increase in response to higher phosphorus concentrations. This often happens right before turnover or destratification in the fall. These algae are most often associated with taste and odor problems, bloom conditions, and the “spilled paint” slick that prompts the most complaints about algae. Each lake possesses a unique blend of algal communities, often varying in population size from year to year and with differing species proportional in the entire population. The most common types range from the mentioned diatoms, green, and blue-green algae, to golden-brown algae to dinoflagellates and many others, dominating each lake community.

So how can this be evaluated through CSLAP? CSLAP does assess algal biomass through the chlorophyll *a* measurement. While algal differentiation is important, many CSLAP lake associations are primarily interested in “how much?,” not “what kind?,” and this is assessed through the chlorophyll *a* measurement. Phytoplankton communities have not been regularly identified and monitored through CSLAP, in part due to the cost and difficulty in analyzing samples and in part due to the difficulty in using a one-time sample to assess long-term variability in lake conditions. A phytoplankton analysis may reflect a temporary, highly unstable and dynamic water-quality condition.

Prior to 1998, nearly all CSLAP lakes were sampled once for phytoplankton identification, but since then, phytoplankton sampling has not been a regular part of CSLAP. For these sampled lakes, a summary of the most abundant phytoplankton species is included below. Algal species frequently associated with taste and odor problems are specifically noted in this table, although it should be mentioned that these samples, like all other water samples collected through CSLAP, come from near the center of the lake, a location not usually near water intakes or swimming beaches. Since algal communities can also be spatially quite variable, even a preponderance of taste- and odor-causing species in the water samples might not necessarily translate to potable-water-intake or aesthetic impairments, although the threat of such an impairment might be duly noted in the “Considerations” section below.

Phytoplankton surveys have not been conducted through CSLAP at Eagle Lake.

VIII: PRIORITY WATERBODY LISTS AND IMPACTS TO LAKE USE

The Priority Waterbody List (PWL) is presently an inventory of all waters in New York State (lakes, ponds, reservoirs, rivers, streams, and estuaries) known to have designated water uses with some degree of impairment, or those threatened by potential impairment. However, the PWL is slowly evolving into an inventory of all waterbodies for which sufficient information is available to assess the condition and/or usability of the waterbody. PWL waterbodies are identified through a broad network of county and state agencies, with significant public outreach and input, and the list is maintained and compiled by the NYSDEC Division of Water. Monitoring data from a variety of sources, including CSLAP, have been utilized by state agencies to evaluate lakes for inclusion on the PWL, and the process for incorporating lakes data has become more standardized.

Specific numeric criteria have recently been developed to characterize sampled lakes in the available use-based PWL categories (*precluded, impaired, stressed, or threatened*). Evaluations utilize the NYS phosphorus guidance value, water-quality standards, criteria utilized by other states, and the trophic ranges described earlier to supplement the other more antidotal inputs to the listing. The procedures by which waterbodies are evaluated are known as the Consolidated Assessment and Listing Methodology (CALM) process. This process is undertaken on an annual rotating basin, with waterbodies in several drainage basins evaluated each year. Each of the 17 drainage basins in the state is assessed within every 5 years. In general, waterbodies that violate pertinent water-quality standards (such as those listed in Table 6) at a frequency of greater than 25% are identified as *impaired*, at a frequency of 10-25% are identified as *stressed*, and at a frequency of 0-10% are identified as *threatened*, although some evidence of use impairment (including through CSLAP lake-perception surveys) might also be required. Mean (average) phosphorus levels are evaluated against the state guidance value. Evidence of use prohibitions (via beach closures, etc.) is often required to identify a waterbody as *precluded*, while evidence of actual use restrictions or necessary management must accompany an *impaired* listing, at least for lakes evaluated in recent years.

Lakes that have been identified as *precluded* or *impaired* on the PWL are likely candidates for the federal 303(d) list, an “Impaired Waters” designation mandated by the federal Clean Water Act. Lakes on this list must be closely evaluated for the causes and sources of these problems. Remedial measures must be undertaken, under a defined schedule, to solve these water-quality problems. This entire evaluation and remediation process is known as the “TMDL” process, which refers to the Total Maximum Daily Load calculations necessary to determine how much (pollution that causes the water-quality problems) is too much.

Eagle Lake is not presently among the lakes listed on the Upper Hudson River Basin PWL.

TABLE 6- Water-Quality Standards Associated With Class B and Higher Lakes

<u>Parameter</u>	<u>Acceptable Level</u>	<u>To Protect.....</u>
Secchi Disk Transparency	> 1.2 meters*	Swimming
Total Phosphorus	< 0.020 mg/L and Narrative*	Swimming
Chlorophyll a	none	NA
Nitrate Nitrogen	< 10 mg/L and Narrative*	Drinking Water
Ammonia Nitrogen	2 mg/L*	Drinking Water
True Color	Narrative*	Swimming
pH	< 8.5 and > 6.5*	Aquatic Life
Conductivity	None	NA

Narrative Standards and Notes:

Secchi Disk Transparency: The 1.2 meter (4 feet) guidance is applied for safety reasons (to see submerged swimmers or bottom debris) and strictly applies only to citing new swimming beaches, but may be appropriate for all waterbodies used for contact recreation (swimming).

Phosphorus and Nitrogen: “None in amounts that will result in the growths of algae, weeds and slimes that will impair the waters for their best usages” (Class B= swimming)

-The 0.020 mg/l threshold for TP corresponds to a guidance value, not a standard; it strictly applies to Class B and higher waters but may be appropriate for other waterbodies used for contact recreation (swimming).

NYS (and other states) is in the process of identifying numerical nutrient (phosphorus and perhaps Secchi disk transparency, chlorophyll *a*, and nitrogen) standards, but this is unlikely to be finalized within the next several years.

-The 10 mg/L Nitrate standard strictly applies to only Class A or higher waters, but is included here because some Class B lakes are informally used for potable-water intake.

-For the form of ammonia (NH₃+NH₄) analyzed, a 2 mg/l human health standard applies to Class A or higher waters. Lower un-ionized ammonia standards apply to all classes of NYS lakes, this form is not analyzed through CSLAP.

Color: “None in amounts that will adversely affect the color or impair the waters for their best usages” (for Class B waters, this is swimming).

pH: The standard applies to all classes of waterbodies

1. Water-quality Standards Evaluation on Eagle Lake:

pH readings have exceeded the NYS water-quality standards (=6.5 to 8.5) during 2% of the CSLAP sampling sessions at Eagle Lake, and failed to reach these standards 5% of the time. Phosphorus levels at Eagle Lake have never exceeded the phosphorus guidance value for NYS lakes (=0.020 mg/l) during any of the CSLAP sampling sessions, and water transparency readings have exceeded the minimum recommended water clarity for swimming beaches (= 1.2 meters) during 100% of the CSLAP sampling sessions. It is not known whether any of the narrative water-quality standards listed in Table 3 have been violated at Eagle Lake; none of the other numeric standards summarized in Table 3 have been violated.

2. Lake Uses:

Water-quality monitoring programs are devised to evaluate lake conditions as they relate to a variety of lake indicators, from water-quality standards to trophic conditions to invasive species to other measures of the physical, chemical, and biological integrity of these ecological systems. One of these indicators is intended to be lake uses--whether these lakes and ponds can be used for potable water, swimming and bathing, fishing and use of the water by aquatic life, and aesthetics. This is consistent with the broad goals of the 1972 federal Clean Water Act, the governing legislation for federal and state management of lakes and ponds, which states that a fundamental goal of environmental management was to make all waterbodies “fishable and swimmable” by 1983.

The “fishability” of a lake or pond is a function of water-quality (are there pollutants that will kill the fish or render them inedible?); substrate and habitat (is there enough cold water and high oxygen for coldwater fish?; is there enough food for the fish? is there enough cover from predators or structure for fishermen?); space (is there enough flowing water for survival or reproduction?; is there enough room to support all of the various fish species in the lake?), and even access (can anglers get to the areas where the fish can be found?).

Likewise, the “swimmability” of a lake or pond also depends on water-quality (will I get sick due to bacterial contamination from sewage, stormwater or waterfowl?); safety (can swimmers or bottom debris be seen in deeper water?); aesthetics (is the water too green, too weedy, or too cold?; is the bottom too mucky?); user conflicts (can I swim where people use PWCs?); the physical characteristics of the lake and shoreline (how quickly does the lake get too deep? is the shoreline flat enough for a beach?); legal considerations (will the threat of litigation

prevent a lake community from establishing public beaches?), and also access (can swimmers from less hospitable parts of the lake or from the outside swim at a beach?).

Although other designated lake uses are not identified as primary goals of the Clean Water Act, they should be evaluated as part of the lake-assessment process. These include potable water, non-contact recreational uses such as boating, aquatic life support unrelated to fishing, and aesthetics. Similar questions could be posed about the suitability of a particular lake or pond for this use, although many of the concerns addressed in evaluating the fishability or swimmability of a waterbody are pertinent to evaluating drinking-water quality, the ability of a lake to support power boating or sailing, or the adequacy of the lake bottom for salamanders, frogs, and other valued biota.

CSLAP is not really designed to answer many of these questions, at least directly. Some of these issues relate to the physical characteristics of the entire shoreline and bottom of the lake or pond and cannot be easily evaluated in simple water-quality surveys. Other important water-quality indicators, such as bacteria, cannot be sampled at the frequency needed to compare lake conditions to existing water-quality standards or are limited by logistic considerations. Other indicators, such as sediment toxins, are too expensive to be included in standard water-quality monitoring programs. It is anticipated that future generations of CSLAP will look to better address some of these questions through expanded monitoring and partnerships with other monitoring agencies, academic institutions, lake residents, and other parties invested in the lake-assessment and management process. It is also anticipated that data from other sources will be more completely included in the lake- and pond-assessment process in the future. Until that time, however, it should again be stated that these assessments are both preliminary and incomplete, based on data presently collectable through the monitoring programs summarized in this report.

Eagle Lake is a Class B lake, which means it is designated for support of contact recreation (swimming and bathing), non-contact recreation (such as boating), aesthetics, and aquatic life (including fishing). As such, Eagle Lake should be evaluated for its best intended uses—support of swimming, aquatic life, non-contact recreation, and aesthetics.

a. Potable Water

Eagle Lake is not classified for potable water use, and it is assumed that the lake does not presently sustain this use.

b. Swimming/Contact Recreation

Eagle Lake is presently used for swimming, bathing, or other forms of contact recreation, although the frequency of and opportunities for swimming are not evaluated through CSLAP. As noted above, it is classified for bathing and swimming.

A number of water-quality indicators are measured in CSLAP that relate to the suitability of lake for swimming and contact recreation. Water clarity measurements can be used to evaluate the lake against the NYS Department of Health guidelines for siting new swimming beaches (= 4 feet). Public-perception data collected through CSLAP assess swimming conditions, and regional

or statewide criteria connecting water transparency readings (or nutrient and algae levels) to recreational-use impacts will likely be developed in the near future. However, there remains a relatively strong correlation between contact recreational conditions and phosphorus readings, with recreational-use impacts generally corresponding to the state guidance value for phosphorus (= 20 parts per billion total phosphorus). Algae levels are measured as chlorophyll *a*, while rooted aquatic-plant populations are broadly quantified through CSLAP, and are linked to potential impacts on swimming and aesthetics. These water-quality-based and perception-based evaluations of swimming conditions are outlined below.

1. Water-quality Evaluation of Swimming/Contact Recreation

These data showed that none of the Eagle Lake samples possessed total phosphorus readings exceeding 20 parts per billion (=µg/l), which corresponds to the state phosphorus guidance value. Water transparency readings were never less than 2 meters during any of the CSLAP sampling sessions. This roughly corresponds to the distinction between *eutrophic* and *mesotrophic* lakes and a water clarity reading that would roughly be equivalent to the state phosphorus guidance value. Perhaps more importantly, this may correspond to the saddle point between high-quality and reduced-quality swimming, based on lake perception data (see below).

Although there is no state water-quality standard for chlorophyll *a*, readings exceeding 8 µg/l generally correspond to water clarity readings lower than 2 meters and total phosphorus readings in excess of 20 µg/l- each of these indicator thresholds marks the distinction between *mesotrophic* and *eutrophic* lake. 2% of the Eagle Lake samples corresponded to chlorophyll *a* readings > 8 µg/l.

Bacteria data have not been collected through CSLAP on Eagle Lake or (if collected by the lake association or local community) have not been forwarded to the NYSDEC for evaluation.

2. Lake Perception Evaluation

Lake perception data from CSLAP provide insights into recreational (swimming) conditions, perceptions of water clarity, and the density and coverage of aquatic plants. Recreational assessments indicating “beautiful, could not be nicer” and “..excellent for swimming, boating, and overall enjoyment” conditions suggest no limits to recreational use. The frequency of “slightly” to “substantially” impaired conditions may be closely related to the need to implement lake-management actions. These surveys also assess the extent to which these impacts are influenced by excessive weed growth, nuisance algae or poor water clarity.

The evaluation of these survey results, and the extrapolation of these results to a lake-wide assessment, is restricted by the small sample size and the potential for responses that are not representative of the responses from the typical lake resident, whether due to the impact of local conditions or different goals for different lake users. However, these assessments may serve as an instructive starting point for evaluating impacts on lake uses.

The CSLAP volunteers described Eagle Lake as “slightly” impaired during 97% of the CSLAP sampling sessions, and “substantially” impaired 3% of the time. Slightly impaired conditions were associated with excessive algae or poor water clarity during 6% of the sampling sessions, and with excessive weeds 83% of the time. Substantially impaired conditions were never associated with excessive algae, and were associated with excessive weeds during 3% of the CSLAP sampling sessions.

3. Overall Evaluation- Swimming and Contact Recreation

The CSLAP dataset at Eagle Lake, including water chemistry data, physical measurements, and volunteer samplers’ perception data, suggests that swimming and contact recreation may be stressed by excessive weeds.

c. Aquatic Life/Non-Contact Recreation

Eagle Lake supports fishing and other forms of non-contact recreation. Other forms of non-contact recreation, such as boating, may be a function of access points, whether the lake shoreline is inhabited, and water depth, but it is also presumed that Eagle Lake may be used for boating.

While water-quality plays a role in evaluating non-contact recreation, particularly cold-water fisheries, the information needed to properly evaluate fishing quality, angler success, and boating enjoyment and viability are not collected in most routine monitoring programs. It is anticipated that future generations of the CSLAP report will include more comprehensive evaluations of non-contact recreational conditions in lakes and ponds, as databases containing this information become more readily available, but until that time, only ancillary measures can be evaluated.

The primary indicators from these monitoring programs used to evaluate fisheries, aquatic life, and non-contact recreation (boating, etc.) include lake perception surveys, aquatic plant densities (and the presence of invasive exotic plants), and water-quality indicators related to fish habitat and survival, such as pH and ammonia. While other water-quality indicators, such as other forms of nitrogen, can also be used to evaluate water-quality impacts to aquatic life, these indicators are generally found at low enough levels to minimize their utility in evaluating lake conditions. Dissolved oxygen can be very useful in evaluating habitat, but temperature and oxygen profiles are not collected through CSLAP. These datasets can provide at least some insights into the ability of lakes and ponds to support these uses.

1. Fisheries and Aquatic Life Evaluation

pH data are collected through CSLAP. Fish consumption advisories are issued by the NYS Department of Health, and fishing regulations are instituted by the NYSDEC. Lake recreational perception data related to non-contact recreation (fishing and boating) and aesthetics are also collected through CSLAP, and these can be used to evaluate fisheries and aquatic life impacts to Eagle Lake.

These data indicate that pH readings in Eagle Lake samples exceeded the state water-quality standards (= 6.5 to 8.5) during 2% of the CSLAP sampling sessions, and failed to reach these standards 5% of the time. While laboratory pH is not as accurate as field pH for evaluating lake acidity, these data suggest that fisheries or aquatic life impacts do not occur as a result of depressed or elevated pH.

It is not known if fishing regulations result in any impact to the use of Eagle Lake for fishing. The lake possesses some coolwater or coldwater fish species susceptible to low oxygen levels in coldwater habitats (deepwater conditions during the summer, and throughout the water column during other times of the year), and it is unlikely that significant oxygen deficits occur in the lake, based on LCI dissolved oxygen readings and CSLAP deepwater nutrient (phosphorus and ammonia) readings, although slight hypolimnetic oxygen deficits have been recorded since at least the 1930s.

2. Boating (Recreation) and Aesthetics Evaluation

Impacts to non-contact recreation, such as boating and aesthetics, can only be peripherally evaluated through CSLAP. Sampling volunteers can report that the lake “looks bad,” as a direct measure of impacts to lake aesthetics, while “poor water clarity,” “excessive algae growth,” and “excessive weed growth” may be indirect measures of these impacts.

The CSLAP volunteers never reported that Eagle Lake “looks bad” during any of the CSLAP sampling sessions. Surface weed growth was reported during 86% of the CSLAP sampling sessions, but dense weed growth was never reported.

3. Overall Evaluation- Aquatic Life and Non-Contact Recreation

The CSLAP dataset on Eagle Lake, including water chemistry data, physical measurements, and volunteer samplers’ perception data, suggest that aquatic life should be fully support, although non-contact recreation may be stressed by excessive weeds.

IX: CONSIDERATIONS FOR LAKE MANAGEMENT

CSLAP is intended for a variety of uses, such as collecting needed information for comprehensive lake management, although it is not capable of collecting all the needed information. To this end, this section includes a broad summary of the major lake problems and “considerations” for lake management. These include only those lake problems that may have been defined by CSLAP sampling, such as physical condition (algae and water clarity), aquatic plant coverage (type and extent of weed populations), and recreational suitability of the lake, as related to contact recreation. These broad categories may not encompass the most pressing issue at a particular time at any given CSLAP lake, for example, local concerns about filamentous algae or concerns about other parameters not analyzed in the CSLAP sampling. While there is some opportunity for CLSAP-trained volunteers to report and assess some site-specific conditions or concerns on the CSLAP Field Observations Form, such as algae blooms or shoreline vegetation, this section is limited to the confines of this program. The categories

represent the most common, broadest issues within the lake management as reported through CSLAP.

Each summarized management strategy is more extensively outlined in *Diet for a Small Lake*, and this joint NYSDEC-NYSFLA publication should be consulted for more details and for a broader context of in-lake- or watershed- management techniques. These “considerations” should not be construed as “recommendations,” because there is insufficient information available through CSLAP to assess whether or how a lake should be managed. Issues associated with local environmental sensitivity, permits, and broad community-management objectives also cannot be addressed here. Rather, the following section should be considered as “tips” or a compilation of suggestions for a lake association to manage problems defined by CSLAP water-quality data or articulated by perception data. When appropriate, lake-specific management information, and other lake-specific or local “data” (such as the presence of a controllable outlet structure) is reported in **bold** in this “considerations” section.

The primary focus of CSLAP monitoring is to evaluate lake condition and impacts associated with lake eutrophication. Because lake eutrophication is often manifested in excessive plant growth, whether algae or aquatic macrophytes (weeds), it is likely that lake-management activities, whether promulgated to reduce algae or weed growth or to maintain water clarity and the existing makeup and density of aquatic plants in the lake, will need to address watershed inputs of nutrients and sediment to the lake, because both can contribute to either algal blooms or excessive weed growth. A core group of nutrient and sediment control activities will likely serve as the foundation for most comprehensive lake-management plans and activities and can be summarized below.

a. GENERAL CONSIDERATIONS FOR ALL CSLAP LAKES

Nutrient controls can take several forms, depending on the original source of the nutrients:

- Septic systems can be regularly pumped or upgraded to reduce the stress on the leach fields which can be replaced with new soil or moving the discharge from the septic tank to a new field). Pumpout programs are usually quite inexpensive, particularly when lakefront residents negotiate a bulk rate discount with local pumping companies. Upgrading systems can be expensive, but may be necessary to handle the increased loading from camp expansion or conversion to year-round residency. Replacing leach fields alone can be expensive and limited by local soil or slope conditions, but may be the only way to reduce actual nutrient loading from septic systems to the lake. It should be noted that upgrading or replacing the leach field may do little to change any bacterial loading to the lake, since bacteria are controlled primarily within the septic tank, not the leach field.
- Stormwater runoff control plans include street cleaning, artificial marshes, sedimentation basins, runoff conveyance systems, and other strategies aimed at minimizing or intercepting pollutant discharge from impervious surfaces. The NYSDEC has developed a guide called Reducing the Impacts of Stormwater Runoff to provide more detailed information about developing a stormwater management plan. This is a strategy that cannot generally be tackled by an individual homeowner, but rather requires the effort and cooperation of lake residents and municipal officials.

- There are numerous agriculture management practices such as fertilizer controls, soil erosion practices, and control of animal wastes, which either reduce nutrient export or retain particles lost from agricultural fields. These practices are frequently employed in cooperation with county Soil and Water Conservation District offices, and are described in greater detail in the NYSDEC's Controlling Agricultural Nonpoint Source Water Pollution in New York State. Like stormwater controls, these require the cooperation of many watershed partners, including farmers.
- Streambank erosion can be caused by increased flow due to poorly managed urban areas, agricultural fields, construction sites, and deforested areas, or it may simply come from repetitive flow over disturbed streambanks. Control strategies may involve streambank stabilization, detention basins, revegetation, and water diversion.

Land use restrictions development and zoning tools such as floodplain management, master planning to allow for development clusters in more tolerant areas in the watershed and protection of more sensitive areas, deed or contracts which limit access to the lake, and cutting restrictions can be used to reduce pollutant loading to lakes. This approach varies greatly from one community to the next and frequently involves balancing lake-use protection with land-use restrictions. State law gives great latitude to local government in developing land-use plans.

Lawn fertilizers frequently contain phosphorus, even though nitrogen is more likely to be the limiting nutrient for grasses and other terrestrial plants. By using lawn fertilizers with little or no phosphorus, eliminating lawn fertilizers or using lake water as a “fertilizer” on shoreline properties, fewer nutrients may enter the lake. Retaining the original flora as much as possible, or planting a buffer strip (trees, bushes, shrubs) along the shoreline, can reduce the nutrient load leaving a residential lawn.

Waterfowl introduce nutrients, plant fragments, and bacteria to the lake water through their feces. Feeding the waterfowl encourages congregation which in turn concentrates and increases this nutrient source and will increase the likelihood that plant fragments, particularly from Eurasian watermilfoil and other plants that easily fragment and reproduce through small fragments, can be introduced to a previously uncolonized lake..

Although not really a “watershed control strategy”, establishing **no-wake zones** can reduce shoreline erosion and local turbidity. Wave action, which can disturb flocculent bottom sediments and unconsolidated shoreline terrain is ultimately reduced, minimizing the spread of fertile soils to susceptible portions of the lake.

Do not discard or introduce plants from one water source to another or deliberately introduce a "new" species from a catalogue or vendor. For example, do not empty bilge or bait bucket water from another lake upon arrival at another lake, for this may contain traces of exotic plants or animals. Do not empty aquaria wastewater or plants in the lake.

Boat propellers are a major mode of transport to uncolonized lakes. Propellers, hitches, and trailers frequently get entangled by weeds and weed fragments. Boats not cleaned of fragments after leaving a colonized lake may introduce plant fragments to another location. New introductions of plants are often found near public access sites.

b. SPECIFIC CONSIDERATIONS FOR EAGLE LAKE

Management Focus: **Water Clarity/Algae/Physical Condition/Recreational Condition**

Issue	Through	By?
Maintain water clarity	Maintaining or reducing algae levels	Maintaining or reducing nutrient inputs to the lake

Discussion:

User perception and water quality data indicate a favorable physical condition and water clarity of the lake. This places the focus of water clarity management on maintaining present conditions, an enviable position for many other lake associations. Although some increase in nutrient loading is inevitable, the lake association should devote efforts to minimize the input of nutrients to the lake, or change activities that otherwise influence water clarity.

Management Focus: **The Impact of Weeds on Recreational Condition**

Problem	Probable Cause	Probable Source
Moderate to Excessive weed growth	Shallow water depth, excessive nutrients and sediment	Excessive pollutant loading from watershed runoff (stormwater, construction sites, agriculture, etc.), septic, bottom disturbance,...

Discussion:

Perception data indicate that aquatic weed growth is perceived to inhibit recreational use of this lake, at least in some parts of the lake or during certain times of the year. Nuisance weed growth in lakes is influenced by a variety of factors- water clarity, sediment characteristics, wave action, competition between individual plant species, sediment nutrient levels, etc. In most cases, excessive weed growth is associated with the presence of exotic, (non-native) submergent plant species such as Eurasian watermilfoil (*Myriophyllum spicatum*), although some lakes are inhibited by dense growth of native species. Some of these factors cannot be controlled by lake association activities, while others can only be addressed peripherally. For example, sediment characteristics can be influenced by the solids loading to the lake. With the exception of some hand harvesting activities, aquatic plant management should only be undertaken when lake uses (recreational, municipal, economic, etc.) are significantly and regularly threatened or impaired. Management strategies can be costly and controversial, and a variety of factors should be weighed. Aquatic plant management most efficiently involves a mix of immediate, in-lake controls, and long-term measures to address the causes and sources of this excessive weed growth.

THE EAGLE LAKE ASSOCIATION HAS BEEN HEAVILY INVOLVED IN LOCAL AND STATEWIDE AQUATIC PLANT MANAGEMENT PLANNING, AND HAS ATTEMPTED TO DIRECT LOCAL EFFORTS TO CONTROL INVASIVE EURASIAN WATERMILFOIL GROWTH WITH THE USE OF AQUATIC HERBICIDES (SPECIFICALLY, FLURIDONE). HAND HARVESTING HAS BEEN USED ON SMALL PLOTS. HOWEVER, THE ENTIRE RANGE OF AQUATIC PLANT MANAGEMENT TECHNIQUES ARE PRESENTED HERE FOR THE SAKE OF COMPLETENESS.

IN-LAKE CONTROL TECHNIQUES

The following strategies primarily address the cause, but not the ultimate source, of problems related to nuisance aquatic plant growth. As such, their effectiveness is necessarily short-term, but perhaps more immediately realized, than strategies that control the source of the problem. Until the sources of the problem are addressed, however, it is likely that these strategies will need to be continuously employed. Some of these are listed in the **Watershed Controls**, since many of the same pollutants contribute to excessive algae growth as well as nuisance weed growth. Except where noted, most of these in-lake techniques do not require permits in most parts of the state, but, as always, the NYDEC Region 5 Offices and the Adirondack Park Agency should be consulted before undertaking these strategies. These techniques are presented within the context of potential management for the conditions (types of nuisance plants, extent of problem) reported through CSLAP. In-lake control methods include: *physical/mechanical plant management techniques, chemical plant management techniques, and biological plant management techniques*

Physical/mechanical control techniques utilize several modes of operation to remove or reduce the growth of nuisance plants. The most commonly employed procedures are the following:

- *Mechanical harvesters* physically remove rooted aquatic plants by using a mechanical machine to cut and transport plants to the shore for proper storage. Mechanical harvesters are probably the most common “formal” plant management strategy in New York State. While it is essentially akin to “mowing the (lake) lawn”, it usually provides access to the lake surface and may remove some lake nutrients if the cut plants are disposed out of the watershed. However, if some shallow areas of the lake are not infested with weeds, they will likely become infested after mechanical harvesting, since fragments frequently wander from cut areas to barren sediment and colonize new plant communities. Harvesters are very expensive, but can be rented or leased. *Rotovators* are rotovating mechanical harvesters, dislodging and removing plants and roots. *Mechanical cutters* cut, but don’t remove, vegetation or fragments. Box springs, sickles, cutting bars, boat props, and anchors often serve as mechanical cutters.
- *Hand harvesting* is the fancy term for lake weeding- pulling out weeds and the root structure by hand. It is very labor intensive, but very plant selective (pull the “weeds”, leave the “plants”); and can be effective if the entire plant is pulled and if the growth area is small enough to be fully cleared of the plant. *Diver dredging* is like hand harvesting with a vacuum cleaner- in this strategy, scuba divers hand-pull plants and place them into a suction hose for removal into a basket in a floating barge. It is also labor intensive and can be quite expensive, but it can be used in water deeper than about 5ft (the rough limit for hand harvesting). It works best where plant beds are dense, but is not very efficient when plant beds or stems are scattered. *This technique has been used at Eagle Lake.*
- *Water level manipulation* is the same thing as *drawdown*, in which the lake surface is lowered, usually over the winter, to expose vegetation and sediments to freezing and drying conditions. Over time this affects the growing characteristics of the plants, and in many cases selectively eliminates susceptible plants. This is obviously limited to lakes that have a mechanism (dam structure, controlled culvert, etc.) for manipulating water level. It is usually very inexpensive, but doesn’t work on all plants and there is a risk of insufficient lake refill

the following spring (causing docks to be orphaned from the waterfront). *It is not believed by the report authors that Eagle Lake can be sufficiently drawn down to utilize this technique.*

- **Bottom barriers** are screens or mats that are placed directly on the lake bottom to prevent the growth of weeds by eliminating sunlight needed for plant survival. The mats are held in place by anchors or stakes, and must be periodically cleaned or removed to detach any surface sediment that may serve as a medium for new growth. The mats, if installed properly, are almost always effective, with relatively few environmental side-effects, but are expensive and do not select for plant control under the mats. It is best used when plant communities are dense but small in area, and is not very efficient for lake-wide control.
- **Sediment removal**, also referred to as dredging, controls aquatic plants by physically removing vegetation and by increasing the depth of the lake so that plant growth is limited by light availability. Dredging projects are usually very successful at increasing depth and controlling vegetation, but they are very expensive, may result in significant side effects (turbidity, algal blooms, potential suspension of toxic materials), and may require significant area for disposal. This procedure usually triggers an extensive permitting process, *particularly in the Adirondack Park.*

Chemical control techniques involve the use of aquatic herbicides to kill undesired aquatic vegetation and prevent future nuisance weed growth. These herbicides come in granular or liquid formulations, and can be applied in spot- or whole-lake treatments. Some herbicides provide plant control by disrupting part of the plants life cycle or ability to produce food, while others have more toxicological effects. Aquatic herbicides are usually effective at controlling plants, but other factors in considering this option include the long term control (longevity), efficiency, and plant selectivity. Effectiveness may also depend on dosage rate, extent of non-target (usually native) plant growth, flushing rate, and other factors. The use of herbicides is often a highly controversial matter frequently influenced by personal philosophies about introducing chemicals to lakes. Some of the more recently registered herbicides appear to be more selective and have fewer side effects than some of the previously utilized chemicals. Chemical control of nuisance plants can be quite expensive, and, with only few exceptions, require permits and licensed applicators. *As discussed above, herbicides appear to be the control strategy of choice, at least among the active lake association members, at Eagle Lake.*

Biological control techniques presently involve the stocking of sterile grass carp, which are herbivorous fish that feed exclusively on macrophytes (and macroalgae). Grass carp, when stocked at the appropriate rate, have been effective at controlling nuisance weeds in many southern states, although their track record in NYS is relatively short, particularly in lakes with shallow or adjacent wetlands or in larger (>100 acre) lakes. These carp may not prefer the nuisance plant species desired for control (in particular Eurasian watermilfoil), and they are quite efficient at converting macrophyte biomass into nutrients that become available for algae growth. This is, however, one of the less expensive means of plant control. *The permitting process for grass carp in the Adirondacks is extensive.*

Naturally occurring biological controls may include native species of *aquatic weevils and moths* which burrow into and ultimately destroy some weeds. These organisms feed on Eurasian watermilfoil, and control nuisance plants in some Finger Lakes and throughout the

Northeast. However, they also inhabit other lakes with varied or undocumented effectiveness for the long term. Because these organisms live in the canopy of weed beds and feed primarily on the top of the plants, harvesting may have a severe negative impact on the population. Research continues about their natural occurrence, and their effectiveness both as a natural or deliberately- introduced control mechanism for Eurasian watermilfoil. *The impact of herbivorous insects on Eurasian watermilfoil in Eagle Lake continues to be evaluated.*

c. SPECIFIC MONITORING CONSIDERATIONS FOR EAGLE LAKE

Discussion:

Eagle Lake has sampled through CSLAP since 2000. More extensive data will help to continue evaluating “normal” conditions on the lake, and to identify water quality or use problems at the lake. However, some additional parameters may be appropriate for evaluation at the lake:

1. *Bacteria*- Eagle Lake is classified for use for contact recreation (swimming), and it is likely that at least some swimming occurs. The use of the lake for swimming and bathing can best be evaluated with bacteriological data. A comparison of sampling results to the state water quality standards requires at least five samples per month. These data cannot be collected through CSLAP.
2. *Algal toxins*- Algal toxins, usually associated with blue-green algae, may affect swimmers and others who ingest small amounts of water (as well as any lake residents who utilize Eagle Lake as a potable water supply). These may be analyzed in standard water samples as part of CSLAP in coming years.
3. *Aquatic plants*- Aquatic plant surveys have not been conducted through CSLAP at Eagle Lake. CSLAP samplers can collect and submit for identification any plant samples thought to be exotic or otherwise invasive, as well as any rare or unusual plants. Sampling protocols are also available to conduct systematic monitoring of aquatic plants for the purpose of evaluating aquatic plant management actions utilized at the lake.
4. *Temperature and oxygen profiles*- the suitability of the lake for supporting sensitive fish, the susceptibility of the lake to nutrient release from bottom sediments and fall algal blooms, and the environment for aquatic plant growth can be evaluated through temperature and oxygen profiles. These can be created through the use of electronic meters or through chemical titrations conducted on site, but, at present, neither of these are collected through CSLAP at Eagle Lake.

Appendix A. Raw Data for Eagle Lake

LNum	PName	Date	Zbot	Zsd	Zsamp	Tot.P	NO3	NH4	TDN	TN/TP	TColor	pH	Cond25	Ca	Chl.a
169	Eagle L	5/30/2000	11.8	5.00	1.5	0.010	0.01				8	7.87	133		1.79
169	Eagle L	6/12/2000	12.4	5.55	1.5	0.006	0.01				7	7.57	129		8.20
169	Eagle L	6/26/2000	11.5	7.95	1.5	0.010	0.01				8	7.94	137		0.56
169	Eagle L	7/10/2000	11.6	5.95	1.5	0.004	0.01				3	7.49	139		1.08
169	Eagle L	7/24/2000	11.5	7.00	1.5	0.004	0.01				7	7.47	136		1.15
169	Eagle L	8/7/2000	11.5	7.40	1.5	0.005	0.01				4	6.82	137		2.08
169	Eagle L	8/22/2000	11.7	7.00	1.5	0.008	0.01				3	8.15	134		0.42
169	Eagle L	9/4/2000	11.5	6.25	1.5	0.006	0.01				6	6.75	138		1.41
169	Eagle L	7/8/2001	11.5	7.00	1.5	0.006	0.01				3	7.91	136		1.51
169	Eagle L	7/22/2001	11.5	7.25	1.5	0.005	0.01				4	7.94	138		0.77
169	Eagle L	8/5/2001	11.5	7.35	1.5	0.005	0.01				3	7.19	117		0.99
169	Eagle L	9/3/2001	11.5	6.95	1.5	0.006	0.01				4	7.64	139		
169	Eagle L	9/30/2001	11.5	7.60	1.5	0.004	0.01				6	6.80	142		
169	Eagle L	10/10/2001	11.5	7.35	1.5	0.012	0.01				5	7.57	145		
169	Eagle L	10/23/2001	11.5	7.20	1.5	0.007	0.01				8	7.31	153		0.78
169	Eagle L	06/10/02	11.5	5.50	1.5	0.009	0.01	0.02	0.56	64.65	9	7.04	142		1.04
169	Eagle L	06/23/02	11.5	6.50	1.5	0.009	0.00	0.03	0.40	46.00	2	7.70	140		
169	Eagle L	07/07/02	11.5	5.90	1.5	0.007	0.00	0.02	0.38	52.77	11	7.93	141		1.23
169	Eagle L	07/19/02	11.5	6.10	1.5	0.007	0.00	0.05	0.54	76.15	16	6.71	147		0.75
169	Eagle L	07/21/02	11.5	6.95	1.5	0.006	0.00	0.03	0.37	66.32	8	7.63	143		
169	Eagle L	08/04/02	11.5	6.30	1.5	0.006	0.00	0.04	0.74	128.35	6	7.95	144		0.99
169	Eagle L	09/01/02	11.5	7.05	1.5	0.007	0.00	0.04	0.55	83.09	2	7.74	141		1.11
169	Eagle L	09/21/02	11.5	6.85	1.5	0.002	0.00	0.01	0.33	172.97	5	7.94	142		1.04
169	Eagle L	6/9/2003	11.5	5.10	1.5	0.005	0.00	0.01	0.36	70.59	7	7.72	147	12.0	1.47
169	Eagle L	6/24/2003	11.5	5.60	1.5	0.008	0.01	0.00	0.32	39.70	5	7.70	148		0.71
169	Eagle L	6/29/2003	11.5	5.10	1.5	0.005	0.01	0.00	0.35	69.03	11	6.74	153		0.92
169	Eagle L	7/13/2003	11.5	6.50	1.5		0.01	0.02	0.23	9.98	9	6.33	157		1.22
169	Eagle L	7/27/2003	11.5	4.75	1.5	0.008	0.00	0.01	0.11	13.12				13.0	0.56
169	Eagle L	8/10/2003	11.5	8.20	1.5	0.005	0.02	0.00	0.41	75.85	9	7.62	149		0.81
169	Eagle L	8/24/2003	11.5	6.35	1.5	0.008	0.00	0.00	0.13	16.23	15	7.60	146		0.44
169	Eagle L	9/8/2003	11.5	7.10		0.002	0.00	0.00	0.32	133.67	2	7.43	147		0.87
169	Eagle L	6/12/2004	11.5	6.80	1.5	0.007	0.01	0.01			9	6.31	120		0.16
169	Eagle L	6/26/2004	11.5	5.60	1.5	0.005	0.03	0.01	0.15	29.44	19	6.69	161		
169	Eagle L	7/10/2004	11.5	6.80	1.5		0.03	0.01	0.32		4	6.94	140		0.20
169	Eagle L	7/24/2004	11.5	6.45	1.5	0.013	0.01	0.01	0.21	15.95	16	8.04	117		1.20
169	Eagle L	8/9/2004	11.5	6.35	1.5	0.004	0.01	0.02	0.19	52.62	15	7.78	157	11.8	1.30
169	Eagle L	8/24/2004	11.5	6.50	1.5	0.003	0.02	0.01	0.29	102.47	17	7.89	168		1.30
169	Eagle L	9/6/2004	11.5	5.60	1.5	0.004	0.01	0.01	0.26	68.24	8	7.88	121		2.50
169	Eagle L	9/18/2004	11.5	5.75	1.5	0.004	0.01	0.01	0.57	135.61	9	7.22	115		0.93
169	Eagle L	6/28/2005	11.5	4.70	1.5	0.008	0.01	0.01	0.15	17.33	11	8.80	147	12.3	0.82
169	Eagle L	7/11/2005	11.5	4.50	1.5	0.005	0.02	0.04	0.24	50.44	9	7.50	134		
169	Eagle L	7/25/2005	11.5	6.20	1.5	0.004	0.03	0.01	0.16	37.53	7	7.59	133		1.11
169	Eagle L	8/7/2005	11.5	7.10	1.5	0.004	0.01	0.01	0.26	64.05	9	8.28	144		0.70
169	Eagle L	8/22/2005	11.5	5.75	1.5	0.005	0.01	0.01	0.11	23.69	46	7.78	138	11.8	0.87
169	Eagle L	9/4/2005	11.5	7.35	1.5	0.004	0.01	0.01	0.61	140.48	6	8.35	129		0.99
169	Eagle L	9/19/2005	11.5	6.25	1.5	0.004	0.01	0.01	0.10	23.88	10	7.63	147		1.08
169	Eagle L	10/1/2005	11.5	5.50	1.5	0.010	0.01	0.01	0.12	12.30	6	7.30	131		0.97
169	Eagle L	6/6/2006	11.5	5.45	1.5	0.006	0.03	0.01	0.44	170.30	13		115	11.0	1.17
169	Eagle L	6/18/2006	11.5	6.25	1.5	0.005	0.02	0.02	0.36	168.10	12	7.88	117		0.87
169	Eagle L	7/2/2006	11.5	4.95	1.5	0.007	0.01	0.01	0.59	188.65	10	8.33	145		0.54
169	Eagle L	7/16/2006	11.5	6.65	1.5	0.006	0.01	0.02	0.61	238.14	4	7.56	128		0.64
169	Eagle L	7/30/2006	11.5	6.25	1.5	0.009	0.02	0.02	0.46	115.55	10	8.02	143	12.3	0.35
169	Eagle L	8/13/2006	11.5	6.35	1.5	0.003	0.01	0.02	0.55	366.93	14	7.58	131		0.76
169	Eagle L	8/27/2006	11.5	6.50	1.5	0.006	0.03	0.04	0.60	240.05	2	7.35	131		1.94
169	Eagle L	9/11/2006	11.5	7.55	1.5	0.003	0.03	0.02	0.59	390.71	2	7.44	123		1.00
169	Eagle L	6/23/2007	11.5	5.10	1.5	0.006	0.02	0.04	0.49	171.7	9	7.3	111	9.9	0.79
169	Eagle L	7/6/2007	11.5	5.35	1.5	0.006	0.01	0.02	1.13	439.9	10	7.3	151		1.21
169	Eagle L	7/19/2007	11.5	6.35	1.5	0.006	0.01	0.01	0.44	158.7	12	8.2	112		0.51
169	Eagle L	7/29/2007	11.5	5.80	1.5	0.005	0.02	0.01	0.49	199.8	13	8.1	130		0.61
169	Eagle L	8/15/2007	11.5	5.40	1.5	0.006	0.01	0.06	0.48	187.9	7	7.6	102	12.4	0.85
169	Eagle L	8/28/2007	11.5	6.45	1.5	0.004	0.12	0.03	0.54	272.8	7	8.2	137		0.41

LNum	PName	Date	Zbot	Zsd	Zsamp	Tot.P	NO3	NH4	TDN	TN/TP	TColor	pH	Cond25	Ca	Chl.a
169	Eagle L	9/10/2007	11.5	6.45	1.5	0.005	0.00	0.12	0.76	309.9	10	8.5	135		1.91
169	Eagle L	9/23/2007	11.5	6.65	1.5	0.005	0.01	0.02	0.51	226.1	8	7.9	142		0.83
169	Eagle L	06/10/02	11.5			0.009	0.01	0.03	0.58	67.04					
169	Eagle L	06/23/02	11.5			0.009	0.00	0.01	0.31	34.03					
169	Eagle L	07/07/02	11.5			0.007	0.00	0.01	0.45	68.53					
169	Eagle L	07/19/02	11.5			0.004	0.00	0.03	0.51	121.37					
169	Eagle L	07/21/02	11.5			0.007	0.00	0.01	0.34	48.04					
169	Eagle L	08/04/02	11.5			0.006	0.00	0.04	0.51	83.32					
169	Eagle L	09/01/02	11.5			0.013	0.00	0.02	0.44	34.96					
169	Eagle L	09/21/02	11.5			0.002	0.00	0.01	0.33	214.16					
169	Eagle L	6/9/2003				0.011	0.00	0.01	0.18	15.80					
169	Eagle L	6/24/2003				0.006	0.01	0.02	0.29	50.61					
169	Eagle L	6/29/2003				0.005	0.01	0.01	0.30	63.42					
169	Eagle L	7/13/2003				0.008	0.00	0.02	0.34	40.48					
169	Eagle L	7/27/2003				0.013	0.00	0.00	0.03	1.87					
169	Eagle L	8/10/2003				0.006	0.02	0.02							
169	Eagle L	8/24/2003				0.004	0.00	0.00	0.29	71.68					
169	Eagle L	9/8/2003			1.5	0.003	0.00	0.00	0.38	124.40					
169	Eagle L	6/12/2004	11.5			0.006	0.01	0.01							
169	Eagle L	6/26/2004	11.5			0.009	0.01	0.01	0.20	21.80					
169	Eagle L	7/10/2004	11.5			0.006	0.02	0.01	0.38	67.34					
169	Eagle L	7/24/2004	11.5			0.004	0.01	0.01	0.27	61.95					
169	Eagle L	8/9/2004	11.5			0.007	0.01	0.01	0.21	31.95					
169	Eagle L	8/24/2004	11.5			0.003	0.02	0.09	1.05	328.55					
169	Eagle L	9/6/2004	11.5			0.004	0.01	0.01	0.25	65.60					
169	Eagle L	9/18/2004	11.5			0.004	0.01	0.01	0.40	96.28					
169	Eagle L	6/28/2005	11.5			0.007									
169	Eagle L	7/11/2005	11.5			0.005									
169	Eagle L	7/25/2005	11.5			0.010									
169	Eagle L	8/7/2005	11.5												
169	Eagle L	8/22/2005	11.5			0.006									
169	Eagle L	9/4/2005	11.5			0.005									
169	Eagle L	9/19/2005	11.5			0.006									
169	Eagle L	10/1/2005	11.5			0.008									
169	Eagle L	6/6/2006	11.5			0.007									
169	Eagle L	6/18/2006	11.5			0.008									
169	Eagle L	7/2/2006	11.5			0.006									
169	Eagle L	7/16/2006	11.5			0.007									
169	Eagle L	7/30/2006	11.5			0.006									
169	Eagle L	8/13/2006	11.5			0.005									
169	Eagle L	8/27/2006	11.5			0.005									
169	Eagle L	9/11/2006	11.5			0.006									
169	Eagle L	6/23/2007	11.5			0.005									
169	Eagle L	7/6/2007	11.5			0.005									
169	Eagle L	7/19/2007	11.5			0.004									
169	Eagle L	7/29/2007	11.5			0.006									
169	Eagle L	8/15/2007	11.5			0.005									
169	Eagle L	8/28/2007	11.5			0.004									
169	Eagle L	9/10/2007	11.5			0.005									
169	Eagle L	9/23/2007	11.5			0.005									

LNum	PName	Date	Zbot	Zsd	Zsamp	QaQc	TAir	TH2O	QA	QB	QC	QD
169	Eagle L	5/30/2000	11.8	5.00	1.5	1	20	15	2	1	3	
169	Eagle L	6/12/2000	12.4	5.55	1.5	1	17	16	2	2	3	25
169	Eagle L	6/26/2000	11.5	7.95	1.5	1	30	24	2	3	3	2
169	Eagle L	7/10/2000	11.6	5.95	1.5	1	22	22	2	3	3	56
169	Eagle L	7/24/2000	11.5	7.00	1.5	1	25	22	2	3	3	2
169	Eagle L	8/7/2000	11.5	7.40	1.5	1	21	23	2	3	3	125
169	Eagle L	8/22/2000	11.7	7.00	1.5	1	26	24	2	3	3	2
169	Eagle L	9/4/2000	11.5	6.25	1.5	1	11	21	2	3	3	25
169	Eagle L	7/8/2001	11.5	7.00	1.5	1	24	21	2	3	3	5
169	Eagle L	7/22/2001	11.5	7.25	1.5	1	24	24				
169	Eagle L	8/5/2001	11.5	7.35	1.5	1	29	25				
169	Eagle L	9/3/2001	11.5	6.95	1.5	1	27	24	2	3	3	2
169	Eagle L	9/30/2001	11.5	7.60	1.5	1	20	18				
169	Eagle L	10/10/2001	11.5	7.35	1.5	1	11	15				
169	Eagle L	10/23/2001	11.5	7.20	1.5	1	13	14	2	3	3	5
169	Eagle L	06/10/02	11.5	5.50	1.5	1	22	16	2	2	3	2
169	Eagle L	06/23/02	11.5	6.50	1.5	1	24	20	2	2	2	5
169	Eagle L	07/07/02	11.5	5.90	1.5	1	22	24	2	3	3	2
169	Eagle L	07/19/02	11.5	6.10	1.5	1	21	25	2	3	3	2
169	Eagle L	07/21/02	11.5	6.95	1.5	1	29	25	2	3	3	2
169	Eagle L	08/04/02	11.5	6.30	1.5	1	26	24	2	3	3	2
169	Eagle L	09/01/02	11.5	7.05	1.5	1	20	22	2	3	3	258
169	Eagle L	09/21/02	11.5	6.85	1.5	1	25	21	2	3	3	2
169	Eagle L	6/9/2003	11.5	5.10	1.5	1	20	16	2	2	3	25
169	Eagle L	6/24/2003	11.5	5.60	1.5	1	30	24	2	2	3	2
169	Eagle L	6/29/2003	11.5	5.10	1.5	1	25	24	2	3	3	25
169	Eagle L	7/13/2003	11.5	6.50	1.5	1	22	24	2	3	3	25
169	Eagle L	7/27/2003	11.5	4.75	1.5	1	24	24	2	3	3	25
169	Eagle L	8/10/2003	11.5	8.20	1.5	1	22	25				
169	Eagle L	8/24/2003	11.5	6.35	1.5	1	18	24	2	3	3	2
169	Eagle L	9/8/2003	11.5	7.10		1	19	21	2	3	3	28
169	Eagle L	6/12/2004	11.5	6.80	1.5	1	20	20	2	3	3	256
169	Eagle L	6/26/2004	11.5	5.60	1.5	1	21	21	2	3	3	25
169	Eagle L	7/10/2004	11.5	6.80	1.5	1	23	22	2	3	3	2
169	Eagle L	7/24/2004	11.5	6.45	1.5	1	22	23	2	3	3	25
169	Eagle L	8/9/2004	11.5	6.35	1.5	1	20	22	2	3	3	2
169	Eagle L	8/24/2004	11.5	6.50	1.5	1	21	22	2	3	3	2
169	Eagle L	9/6/2004	11.5	5.60	1.5	1	21	22	2	3	3	125
169	Eagle L	9/18/2004	11.5	5.75	1.5	1	19	20	2	3	3	125
169	Eagle L	6/28/2005	11.5	4.70	1.5	1	27	26	2	3	3	125
169	Eagle L	7/11/2005	11.5	4.50	1.5	1	29	26	2	3	3	2
169	Eagle L	7/25/2005	11.5	6.20	1.5	1	22	26	2	3	3	25
169	Eagle L	8/7/2005	11.5	7.10	1.5	1	27	28	2	3	3	2
169	Eagle L	8/22/2005	11.5	5.75	1.5	1	20	25	2	3	3	25
169	Eagle L	9/4/2005	11.5	7.35	1.5	1	21	21	2	3	3	2
169	Eagle L	9/19/2005	11.5	6.25	1.5	1	21	21	2	3	3	2
169	Eagle L	10/1/2005	11.5	5.50	1.5	1	19	18	2	3	3	28
169	Eagle L	6/6/2006	11.5	5.45	1.5	1	23	20	2	1	2	25
169	Eagle L	6/18/2006	11.5	6.25	1.5	1	24	21	2	3	3	25
169	Eagle L	7/2/2006	11.5	4.95	1.5	1	23	22	2	3	3	25
169	Eagle L	7/16/2006	11.5	6.65	1.5	1	31	26	2	3	3	2
169	Eagle L	7/30/2006	11.5	6.25	1.5	1	28	26	2	3	3	2
169	Eagle L	8/13/2006	11.5	6.35	1.5	1	22	22	2	3	3	2
169	Eagle L	8/27/2006	11.5	6.50	1.5	1	16	20	2	3	4	24568
169	Eagle L	9/11/2006	11.5	7.55	1.5	1	16	19	2	3	3	2
169	Eagle L	6/23/2007	11.5	5.10	1.5	1	22	21	2	2	3	25
169	Eagle L	7/6/2007	11.5	5.35	1.5	1	24	22	2	3	3	25
169	Eagle L	7/19/2007	11.5	6.35	1.5	1	24	22	2	3	3	2

LNum	PName	Date	Zbot	Zsd	Zsamp	QaQc	TAir	TH2O	QA	QB	QC	QD
169	Eagle L	7/29/2007	11.5	5.80	1.5	1	32	27	2	3	3	2
169	Eagle L	8/15/2007	11.5	5.40	1.5	1	26	25	2	3	3	2
169	Eagle L	8/28/2007	11.5	6.45	1.5	1	22	22	2	3	3	2
169	Eagle L	9/10/2007	11.5	6.45	1.5	1	22	22	2	3	3	2
169	Eagle L	9/23/2007	11.5	6.65	1.5	1	20	19	2	3	4	2
169	Eagle L	06/10/02	11.5			2						
169	Eagle L	06/23/02	11.5			2						
169	Eagle L	07/07/02	11.5			2						
169	Eagle L	07/19/02	11.5			2						
169	Eagle L	07/21/02	11.5			2						
169	Eagle L	08/04/02	11.5			2						
169	Eagle L	09/01/02	11.5			2						
169	Eagle L	09/21/02	11.5			2						
169	Eagle L	6/9/2003				2						
169	Eagle L	6/24/2003				2						
169	Eagle L	6/29/2003				2						
169	Eagle L	7/13/2003				2						
169	Eagle L	7/27/2003				2						
169	Eagle L	8/10/2003				2						
169	Eagle L	8/24/2003				2						
169	Eagle L	9/8/2003			1.5	2						
169	Eagle L	6/12/2004	11.5			2						
169	Eagle L	6/26/2004	11.5			2						
169	Eagle L	7/10/2004	11.5			2						
169	Eagle L	7/24/2004	11.5			2						
169	Eagle L	8/9/2004	11.5			2						
169	Eagle L	8/24/2004	11.5			2						
169	Eagle L	9/6/2004	11.5			2						
169	Eagle L	9/18/2004	11.5			2						
169	Eagle L	6/28/2005	11.5			2						
169	Eagle L	7/11/2005	11.5			2						
169	Eagle L	7/25/2005	11.5			2						
169	Eagle L	8/7/2005	11.5			2						
169	Eagle L	8/22/2005	11.5			2						
169	Eagle L	9/4/2005	11.5			2						
169	Eagle L	9/19/2005	11.5			2						
169	Eagle L	10/1/2005	11.5			2						
169	Eagle L	6/6/2006	11.5			2						
169	Eagle L	6/18/2006	11.5			2						
169	Eagle L	7/2/2006	11.5			2						
169	Eagle L	7/16/2006	11.5			2						
169	Eagle L	7/30/2006	11.5			2						
169	Eagle L	8/13/2006	11.5			2						
169	Eagle L	8/27/2006	11.5			2						
169	Eagle L	9/11/2006	11.5			2						
169	Eagle L	6/23/2007	11.5			2						
169	Eagle L	7/6/2007	11.5			2						
169	Eagle L	7/19/2007	11.5			2						
169	Eagle L	7/29/2007	11.5			2						
169	Eagle L	8/15/2007	11.5			2						
169	Eagle L	8/28/2007	11.5			2						
169	Eagle L	9/10/2007	11.5			2						
169	Eagle L	9/23/2007	11.5			2						

Appendix B. New York State Water-Quality Classifications

- Class N: Enjoyment of water in its natural condition and where compatible, as a source of water for drinking or culinary purposes, bathing, fishing and fish propagation, recreation and any other usages except for the discharge of sewage, industrial wastes or other wastes or any sewage or waste effluent not having filtration resulting from at least 200 feet of lateral travel through unconsolidated earth. These waters should contain no deleterious substances, hydrocarbons or substances that would contribute to eutrophication, nor shall they receive surface runoff containing any such substance.
- Class AA_{special}: Source of water supply for drinking, culinary or food-processing purposes; primary and secondary contact recreation, and fishing. These waters shall be suitable for fish propagation and survival and shall contain no floating solids, settleable solids, oils, sludge deposits, toxic wastes, deleterious substances, colored or other wastes or heated liquids attributable to sewage, industrial wastes or other wastes. There shall be no discharge or disposal of sewage, industrial wastes or other wastes into these waters. These waters shall contain no phosphorus and nitrogen in amounts that will result in growths of algae, weeds and slimes that will impair the waters for their best usages.
- Class A_{special}: Source of water supply for drinking, culinary or food-processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These international boundary waters, if subjected to approved treatment equal to coagulation, sedimentation, filtration and disinfection, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes.
- Class AA: Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation, and fishing. These waters shall be suitable for fish propagation and survival. These waters, if subjected to approved disinfection treatment, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking-water standards and will be considered safe and satisfactory for drinking-water purposes.
- Class A: Source of water supply for drinking, culinary or food-processing purposes; primary and secondary contact recreation, and fishing. These waters shall be suitable for fish propagation and survival. These waters, if subjected to approved treatment equal to coagulation, sedimentation, filtration and disinfection, with additional treatment if necessary to remove naturally

present impurities, will meet New York State Department of Health drinking-water standards and will be considered safe and satisfactory for drinking-water purposes

- Class B Suitable for primary and secondary contact recreation and fishing. These waters shall be suitable for fish propagation and survival.
- Class C: Suitable for fishing and fish propagation and survival. The water quality shall be suitable for primary and secondary contact recreation, although other factors may limit the use for these purposes.
- Class D: Suitable for fishing. Due to such natural conditions as intermittency of flow, water conditions not conducive to propagation of game fishery, or stream bed conditions, the waters will not support fish propagation. These waters shall be suitable for fish survival. The water-quality shall be suitable for primary and secondary contact recreation, although other factors may limit the use for these purposes.
- Class (T): Designated for trout survival, defined by the Environmental Conservation Law Article 11 (NYS, 1984b) as brook trout, brown trout, red throat trout, rainbow trout, and splake.

**APPENDIX C:
SUMMARY OF STATISTICAL METHODS USED TO EVALUATE TRENDS**

1. Non-Parametric Analyses

Kendall tau ranking orders paired observations by one of the variables (arranging water clarity readings by date). Starting with the left-hand (earliest date) pair, the number of times that the variable not ordered (clarity readings) is exceeded by the same variable in subsequent pairs is computed as P, and the number of times in which the unordered variable is not exceeded is computed as Q. This computation is completed for each ordered pair, with N= total number of pairs (samples), and the sum of the differences $S = \Sigma(P-Q)$. The Kendall tau rank correlation coefficient t is computed as:

$$t = 2S/(N*(N-1))$$

Values for t range from -1 (complete negative correlation) to $+1$ (complete positive correlation). As above, strong correlations (or simply “significance”) may be associated with values for t greater than 0.5 (or less than -0.5), and moderate correlations may be associated with values for t between 0.3 and 0.5 (or between -0.3 and -0.5), but the “significance” of this correlation must be further computed. Standard charts for computing the probabilities for testing the significance of S are provided in most statistics text books, and for values of N greater than 10, a standard normal deviate D can be computed by calculating the quotient:

$$D = S\sqrt{18} / \sqrt{[(N(N-1)(2N+5)]}$$

and attributing the following significance:

$$D > 3.29 = 0.05\% \text{ significance}$$

$$2.58 < D < 3.29 = 0.5\% \text{ significance}$$

$$1.96 < D < 2.58 = 2.5\% \text{ significance}$$

$$D < 1.96 = > 2.5\% \text{ significance}$$

For the purpose of this exercise, 2.5% significance or less is necessary to assign validity (or, using the vernacular above, “significance”) to the trend determined by the Kendall tau correlation. It should be noted again that this evaluation does not determine the magnitude of the trend but only whether a trend is likely to occur.

Parametric trends can be defined by standard best-fit linear regression lines, with the significance of these data customarily defined by the magnitude of the best-fit regression coefficient β or R^2 . This can be conducted using raw or individual data points, or seasonal summaries (using some indicator of central tendency, such as mean or median). Because the former can be adversely influenced by seasonal variability and/or imprecision in the length and breadth of the sampling season during any given year, seasonal summaries may provide more realistic measures for long-term trend analyses. However, because the summaries may not adequately reflect variability within any given sampling season, it may be appropriate to compare deviations from seasonal means or medians with the “modeled” change in the mean/median resulting from the regression analyses.

When similar parametric and non-parametric tools are utilized to evaluate long-term trends in NYS lakes, a few assumptions must be adopted:

- Using the non-parametric tools, trend “significance” (defined as no more than approx. 3% “likelihood” that a trend is calculated when none exists) can only be achieved with at least four years of averaged water-quality data. When looking at all summer data points (as opposed to data averaging), a minimum of 40 data points is required to achieve some confidence in data significance. This corresponds to at least five years of CSLAP data. The “lesson” in these assumptions is that data trends assigned to data sets collected over fewer than five years assume only marginal significance.

As noted above, summer data only are utilized (as in the previous analyses) to minimize seasonal effects and different sampling schedules around the fringes (primarily May and September) of the sampling season. This reduces the number of data points used to compile averages or whole data sets but is considered necessary to best evaluate the CSLAP datasets.

2. Parametric Analyses

Parametric analyses are conducted by comparing annual changes in summer mean values for each of the analyzed sampling parameters. Summer is defined as the period from June 15 thru September 15, and roughly corresponds to the window between the end of spring runoff (after ice out) and start of thermal stratification, and the onset of thermal destratification. This period also corresponds to the peak summer recreational season and (for most lakes) the most critical period for water-quality impacts. It also bounds the most frequent range of sampling dates for the majority of both the primarily seasonal volunteers and full-time residents of CSLAP lakes.

Trends in the parametric analyses are determined by the least squares method, in which “significance” requires both a high correlation coefficient ($R^2 > 0.5$) and intra-seasonal variance to be lower than the predicted change (trend) during the period of sampling (roughly corresponding to Δy). Changes in water-quality indicators are also evaluated by the two-sided t-test, in which the change (z statistic) in the mean summer value for each of the indicators by decade of sampling (1980s, 1990s, 2000s) is compared to the t statistic distribution within the 95% confidence interval, with the null hypothesis corresponding to no significant change.

APPENDIX D: BACKGROUND INFO FOR EAGLE LAKE

CSLAP Number	169
Lake Name	Eagle L
First CSLAP Year	2000
Sampled in 2006?	yes
Latitude	435218
Longitude	733702
Elevation (m)	288
Area (ha)	170.9
Volume Code	5
Volume Code Name	Upper Hudson River
Pond Number	438
Qualifier	none
Water-quality Classification	B
County	Essex
Town	Ticonderoga
Watershed Area (ha)	996.5
Retention Time (years)	Not yet determined
Mean Depth (m)	Not yet determined
Runoff (m/yr)	0.510947047
Watershed Number	11
Watershed Name	Upper Hudson River
NOAA Section	3
Closest NOAA Station	North Creek
Closest USGS Gaging Station-Number	4276842
Closest USGS Gaging Station-Name	Putnam Point East of Crown Point Center
CSLAP Lakes in Watershed	Adirondack L, Babcock L, Ballston L, Brant L, Cossayuna L, Eagle L, Efner L, Friends L, Garnet L, Goodnow F, Hedges L, Hunt L, Kellum L, L Forest, L Lauderdale, L Luzerne, Loon L-W, Mayfield L, Moreau L, Paradox L, Piseco L, Sacandaga L, Saratoga L, Schroon L, Summit L-W, Taconic P, Windover L